# DEVELOPMENT AND USE OF A THREE-DIMENSIONAL WATER QUALITY MODEL 

TO PREDICT DISSOLVED OXYGEN CONCENTRATIONS IN THE LOWER CAPE FEAR RIVER ESTUARY, NORTH CAROLINA

James D. Bowen, Solomon Negusse, J. Matthew Goodman, Benoit Duclaud, Mathieu Robin, and Jesslyn Williams

Department of Civil and Environmental Engineering
William States Lee College of Engineering
University of North Carolina at Charlotte
Charlotte, NC 28223

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## EXECUTIVE SUMMARY

An application of the three-dimensional water quality model EFDC (Environmental Fluid Dynamics Code) was developed for the Lower Cape Fear River Estuary, North Carolina. The model was used to investigate the effects of various organic matter and ammonia load reduction scenarios on the dissolved oxygen concentrations within the estuary. The model region included the tidally affected portions of the Cape Fear, Black, and Northeast Cape Fear Rivers near Wilmington, North Carolina, and extended southward to the mouth of the Cape Fear River near Southport, North Carolina.

The model's three-dimensional numerical grid had 1050 horizontal cells and eight vertical layers. The average horizontal grid size was approximately 600 m ; individual cell sizes ranged from 100 to 1300 m . Cell water depths varied for river and estuary channel cells from 2.5 to 12.5 m . EFDC uses a "sigma" vertical grid so that the water depth within a horizontal cell was subdivided into eight layers of uniform thicknesses. The model grid included 241 "marsh" cells that were used to model the off-channel storage of wetlands adjoining the estuary. Flow boundary conditions were utilized at the upstream riverine boundaries; a radiation-separation boundary condition was used at the downstream open boundary. Model input data sets and observed data sets used for calibration and confirmation, and model scenario testing were developed using observed data gathered from various agencies. The period December 1, 2003 to December 31, 2004 was found to have sufficient data to create the necessary model data set for calibration. An additional model data set using data from January 1 to December 31, 2005 was used for model confirmation.

The model calibration period (January 1 - December 31, 2004) had streamflows that were generally below average for the first half of the year, and generally above average for the latter half of the year. After a storm in early May 2004, streamflows remained below average until August. The remainder of 2004 and 2005 until the middle of April had streamflows that were near historical average values. Flows through the summer of 2005 were generally below average. Dissolved oxygen (DO) concentrations at two stations that were in the impaired region and had continuous monitors installed in 2004 (Navassa and the Northeast Cape Fear at Wilmington) showed summertime (April 1 - October 31) dissolved oxygen concentrations that varied between 3 and $7.5 \mathrm{mg} / \mathrm{L}$. At both of these sites the median DO concentration was below $4.5 \mathrm{mg} / \mathrm{L}$ and the $25^{\text {th }}$ percentile DO concentration was below $4.1 \mathrm{mg} / \mathrm{L}$. DO concentrations were generally lower at the Northeast Cape Fear site, and at this site were generally lower near the bottom of the water column.

Model calibration was performed separately on the hydrodynamic and water quality submodels. Hydrodynamic model calibration relied heavily on a set of continuous in-situ water level and water quality monitors that were deployed through the estuary. Hydrodynamic model calibration consisted primarily of varying the location, width, and bottom roughnesses of shallow model cells located adjacent to the river and estuary channels. The distribution of these "marsh" cells utilized information on the location and lateral extent of saltwater marshes and wetland forests adjoining the estuary. The calibrated hydrodynamic model matched well the attenuation of the tidal amplitude signal from the estuary mouth to the upstream model boundaries. The
model was also able to simulate to a reasonable extent the time histories of salinities throughout the estuary.

The twenty-one state variable water quality model available in EFDC included multiple dissolved and particulate organic carbon constituents, as well as organic and inorganic nutrients, dissolved oxygen, and three phytoplankton constituents. To adequately characterize the various organic matter decomposition rates of the riverine and wastewater inputs, both labile and refractory dissolved organic matter constituents were used. The water quality model considered inputs from the three riverine sources at the model boundaries, twenty wastewater point source inputs within the estuary, and fourteen additional point sources that simulated other freshwater inputs to the estuary from tidal creeks and wetlands. Over the three-year time period (20022005) for which the freshwater and point source loadings were developed, approximately $10 \%$ of the organic matter loading and $50 \%$ of the ammonia loading to the estuary came from the twenty wastewater point sources that discharged directly to the estuary.

The dissolved oxygen model utilized a user specified sediment oxygen demand (SOD) that could vary spatially and temporally. Both temporal and spatial variation in SOD was utilized for the model. The temporally variable SOD that was used in this study was based upon an analysis of the available monitoring data. Temporal variation was modeled according the observed variation in measured SOD with changes in water temperature during the SOD measurement. SOD values were then adjusted seasonally according to observed changes in water temperature in the water body. During calibration it was determined that higher SOD values should be used for the Northeast Cape Fear. The ratio between these higher SOD values and the SOD values for remaining portion of the model region was determined during calibration. Discharge and estuary monitoring data were used to quantify the loadings of organic matter, nutrients, and dissolved oxygen to the estuary. As part of calibration SOD values were varied to values above and below the average values from the monitoring program. The calibrated value was found to be nearly identical to the corresponding value from the monitoring program.

The calibrated model matched well time histories of observed dissolved oxygen concentration. Model predictions were compared to over 5200 measured dissolved oxygen concentrations collected at eighteen sites throughout the estuary. The mean model error was less than $0.01 \mathrm{mg} / \mathrm{L}$, and the root mean square error was $0.92 \mathrm{mg} / \mathrm{L}$, which corresponded to $13.8 \%$ of the mean value. The correlation $r^{2}$ was $84.4 \%$. This is an excellent degree of model fit for an estuarine dissolved oxygen model. Time history comparisons were also made for other important water quality constituents, such as nitrate + nitrite, orthophosphate, total nitrogen, and chlorophyll. In each case the model acceptably simulated the spatial and temporal patterns in the observed data set. A water quality model confirmation test run using input and calibration data from 2005 showed a similar degree of fit between observed and predicted dissolved oxygen concentrations. Based upon this work, the calibrated model is considered to be suitable for conducting scenario tests on the effect of changes in organic matter and ammonia loadings on the dissolved oxygen concentrations in the estuary.

A number of scenario tests were conducted to investigate the system's sensitivity to loading changes and other possible changes in the water body. For each scenario, a model run
was used to determine the predicted dissolved oxygen concentrations during the summer period of 2004 (April 1 - October 31) within the state designated impaired region of the estuary. Twice daily dissolved oxygen snapshots were recorded from all eight vertical layers at eighteen sites within the impaired region. More than twenty separate model runs were made and compared to the base case to determine how the dissolved oxygen concentrations might vary for a particular scenario. These eight scenarios were examined:

1. Eliminating wastewater point source loadings
2. Reducing river, creek, and wetland loadings
3. Changing wastewater loadings for various values of sediment oxygen demand
4. Reducing river, creek, and wetland loadings and sediment oxygen demand
5. Eliminating ammonia inputs from wastewater point sources
6. Increasing wastewater inputs to maximum permitted values
7. Deepening of the navigation channel
8. Changing Brunswick County wastewater loadings

Four model runs were made to investigate the effect of completely turning off the discharges from the wastewater treatment plants. These scenarios were run, not because the pollutant removals used were considered achievable or advisable, but instead these scenarios were run to investigate the system's sensitivity to WWTP loads. River, creek, and wetland inputs were maintained at the base condition levels. It was assumed that organic matter and ammonia concentrations in the wastewater inputs would be decreased to $0.0 \mathrm{mg} / \mathrm{L}$. When all WWTP loads were eliminated, the $10^{\text {th }}$ percentile DO concentration increased by approximately $0.3 \mathrm{mg} / \mathrm{L}$, from 4.3 to $4.5 \mathrm{mg} / \mathrm{L}$. The median value DO concentration increased by about 0.10 $\mathrm{mg} / \mathrm{L}$, from 5.6 to $5.7 \mathrm{mg} / \mathrm{L}$. Selectively turning off individual WWTP loads, as expected, had a smaller effect. Turning off the International Paper WWTP had an effect that was roughly $2 / 3$ of that when all the plants were turned off. Turning off both Wilmington domestic wastewater treatment plants had an effect that was roughly $1 / 3$ of that when all discharges were turned off. A fourth scenario considered eliminating the ammonia loading for all wastewater treatment plants. This scenario increased the dissolved oxygen concentration by about $0.1 \mathrm{mg} / \mathrm{L}$ at the $10^{\text {th }}$ percentile level.

The impact of changing the loadings that entered the model region from rivers, creeks, and wetlands were also investigated. Loading reductions of $30 \%, 50 \%$, or $70 \%$ were assumed for the three river inputs (Cape Fear, Black, and Northeast Cape Fear), and from the fourteen creeks and wetland inputs in the estuary. To reduce the loadings, flows were maintained at the observed levels, but concentrations were reduced by $30 \%, 50 \%$, or $70 \%$. Loadings for the twenty wastewater point sources were maintained at the levels in the base case scenario. These load reductions were found to have a significant impact on dissolved oxygen concentrations. At the $10^{\text {th }}$ percentile level, DO concentrations for the $30 \%, 50 \%$, and $70 \%$ load reduction increased by $0.20,0.3$, and $0.40 \mathrm{mg} / \mathrm{L}$ respectively, from $4.3 \mathrm{mg} / \mathrm{l}$ to either $4.5,4.6$ or $4.7 \mathrm{mg} / \mathrm{L}$. Unlike the other scenarios investigated, this level of increase in DO concentration was maintained at the higher percentiles. In fact, the median DO concentration was increased to even a greater extent, increasing from 5.6 to $5.85 \mathrm{mg} / \mathrm{L}$ for the $30 \%$ load reduction, and from 5.6 to $6.2 \mathrm{mg} / \mathrm{L}$ for the $70 \%$ load reduction scenario.

Additional water quality model runs were conducted to investigate the sensitivity of the results to the choice of model kinetic parameters. Because of its relative importance in determining dissolved oxygen concentrations, the sensitivity analysis focused on the impact of various choices for the sediment oxygen demand. During calibration it was found that this parameter had a significant impact on dissolved oxygen concentrations in the estuary. In addition to the calibrated SOD value $\left(0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right.$ at $20^{\circ} \mathrm{C}$ ), runs were made at a lower value ( 0.3 $\mathrm{g} / \mathrm{m}^{2} / \mathrm{d}$ at $20^{\circ} \mathrm{C}$ ) and a higher value $\left(0.5 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right.$ at $\left.20^{\circ} \mathrm{C}\right)$. Model fit to the observed data set was compared for these three cases. All three cases had similar correlation $r^{2}$ values ( $83.4 \%, 83.4 \%$, and $84.5 \%$ ), but on average the low SOD case overpredicted DO concentrations by $0.11 \mathrm{mg} / \mathrm{L}$, while the high SOD case underpredicted DO concentrations by a similar amount ( $0.09 \mathrm{mg} / \mathrm{L}$ ). The impact of turning off the organic matter and ammonia loadings from all wastewater treatment plants was investigated for each of these three SOD cases. Each scenario produced a very similar change in DO concentration when all point sources were turned off. Even though DO concentrations were changed with the different SOD cases, in each scenario, turning off the sources increased DO concentrations by approximately $0.25 \mathrm{mg} / \mathrm{L}$ at the $10^{\text {th }}$ percentile level. It is expected that changes in other model parameters would have a similar result. Although absolute concentrations are affected by changes in parameter values, relative changes in concentration are less affected by changing parameter values.

An additional set of model runs were conducted to investigate the effect of reducing both river loading of organic matter and sediment oxygen demand. Three pairs of model runs were made that reduced river loading or river loading and SOD by $30 \%, 50 \%$, and $70 \%$, using the methods described earlier. As expected, a larger effect was seen when both river loading and SOD were reduced. A $30 \%$ reduction raised the $10^{\text {th }}$ percentile DO concentration from 4.3 to 4.8 $\mathrm{mg} / \mathrm{L}$; a $50 \%$ reduction raised the $10^{\text {th }}$ percentile DO concentration to $5.2 \mathrm{mg} / \mathrm{L}$; a $70 \%$ reduction raised the $10^{\text {th }}$ percentile DO concentration to $5.4 \mathrm{mg} / \mathrm{L}$. In this final case where both river load and SOD were reduced by $70 \%$ from the base case, only about $1 \%$ of the model predicted DO concentrations in the impaired region were less than the water quality standard value of 5.0 $\mathrm{mg} / \mathrm{L}$.

The impact of increasing all wastewater treatment plant loadings to the maximum allowed by permit was also investigated. For cases where both the maximum concentration and flow were specified, both these values were used in the scenario. In permits with only a load or concentration limit, the existing flows were maintained, but concentrations were adjusted to produce the permitted limit. Using these maximum permitted loadings decreased predicted dissolved concentrations in the impaired region by approximately $0.1 \mathrm{mg} / \mathrm{L}$. This level of decrease was fairly uniform across the range of predicted concentrations.

The impact of deepening the navigation channel on predicted dissolved oxygen concentrations in the estuary was also investigated. Model grid files were adjusted to match the channel deepening plan developed by the Wilmington district of the US Army Corps of Engineers that will deepen portions of the Cape Fear and Northeast Cape Fear Rivers. It was found that this deepening would decrease dissolved oxygen concentrations by approximately 0.1 to $0.2 \mathrm{mg} / \mathrm{L}$. The percentage of dissolved oxygen concentrations in the summertime in the impaired region that would be below $5.0 \mathrm{mg} / \mathrm{L}$ would increase slightly with the channel deepening from $32 \%$ to $34 \%$.

Three model scenarios considered the impact of changing the wastewater flow of a single discharger (Brunswick County wastewater treatment plant). Based on the discharge monitoring reports for the model period, the base condition had a time-averaged flow of 0.38 MGD. The three scenarios tested had time-averaged flow of 1.65, 4.65 and 15 MGD. These scenarios therefore represented the effect of changing the wastewater treatment flow by factors of 4.3, 12.1, and 39.1. Despite the relatively large changes in the assumed wastewater flow, there were only very small differences in the predicted dissolved oxygen concentrations, for summertime conditions in the impaired region. Across the entire range of dissolved oxygen concentrations, the differences between the base case and each of the three scenarios were always less than 0.05 $\mathrm{mg} / \mathrm{L}$.

An analysis was performed to attribute the observed oxygen depletion within the estuary to riverine loadings of degradable waste, wastewater discharges, and sediment oxygen demand. It was found that for three sites in the impaired region during the summer, less than $10 \%$ of the dissolved oxygen deficit was attributable to wastewater discharges. Riverine loadings and sediment oxygen demand each contributed similar amounts to the remaining $90 \%$ of the deficit.

The report concludes with some recommendations as to additional development of the model that seems justified. Both this model study and a previous effort demonstrated the importance of benthic fluxes of oxygen. As this model used a prescribed sediment oxygen demand, albeit one that varies temporally and spatially, further refinement of the sediment model seems warranted. In addition, some additional work seems justified to separately consider the effects of wetland and riverine loadings to dissolved oxygen conditions in the estuary.

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## 1. INTRODUCTION

### 1.1 Background

The Cape Fear River watershed is the largest in North Carolina with a drainage area of 9,149 square miles. The watershed is nearly 200 miles long and is completely contained within the state (NC DENR 2004). The watershed is divided into upper, middle, and lower regions. The upper region contains the Triad metropolitan area of Greensboro, High Point, and WinstonSalem. The Lower Cape Fear River Estuary occupies the southernmost 60 kilometers of the watershed, from upstream of Wilmington to the Atlantic Ocean (Figure 1). The estuary includes not only the tidally affected portion of the Cape Fear River, but also the lowermost portions of the Black and Northeast Cape Fear Rivers. The Wilmington metropolitan area is located within the estuarine portion of the Lower Cape Fear watershed.

The river and its watershed are vital resources to the state providing not only as a water supply for drinking and industrial use, but also by providing habitat for countless aquatic plants and animals. As such, reaches within the Cape Fear River watershed have been deemed by the North Carolina government to have designated uses such as water supply, primary and secondary recreation, aquatic life propagation, and agriculture depending on the desired use of the water and its quality (see classifications listing at http://h2o.enr.state.nc.us/csu/). State (NC Department of Environment and Natural Resources) and Federal (Environmental Protection Agency) agencies are charged with maintaining, protecting, and restoring the quality of surface waters. Section 303(d) of the Federal Clean Water Act (CWA) requires all states to develop a list of waters not meeting water quality standards or which are not meeting their designated use. Listed waters must be prioritized, and a management strategy or total maximum daily load (TMDL) analysis must subsequently be developed for all listed waters.

A portion of the Lower Cape Fear River Estuary occupying 5,616.7 acres is currently on the state's 303(d) list (NC Division of Water Quality 2008). This listing, having the following description:

From upstream mouth of Toomers Cr. to a line across the river between Lilliput Creek and Snow's cut
is given in part because of violations of the state's water quality standard for dissolved oxygen ( $5.0 \mathrm{mg} / \mathrm{L}$ for surface waters, and $4.0 \mathrm{mg} / \mathrm{L}$ for swamp waters) (NCDEHNR 1994). Because of the history of dissolved oxygen water quality violations, and the 303(d) listing of this portion of the estuary, the NC is currently conducting a dissolved oxygen TMDL analysis of the estuary. As part of this TMDL analysis, a three-dimensional water quality model of the estuarine portion of the Cape Fear River watershed has been developed, tested, and utilized to run management scenarios. This report summarizes the development, testing, and use of this model.

### 1.2 Study Objectives

The overall objective of the study was to develop a water quality model of the Lower Cape Fear River estuary that would be suitable for use during the dissolved oxygen TMDL analysis performed by the NC Division of Water Quality (DWQ). During the early investigative phase of the project, it was determined that a three-dimensional numerical model would be necessary. Once the model was selected, developed and tested, the project objective was to estimate the nature and magnitude of changes in the water quality of the estuary that would result from changes in inputs of oxygen demanding wastes to the estuary. This objective was achieved by formulating, running, and analyzing several sets of model scenarios that simulated water quality conditions in the estuary for various loading levels of oxygen demanding wastes. A secondary model objective was to assess the reliability of model predictions. This objective was achieved through model confirmation and model sensitivity testing.

### 1.3 Organization of the Study Report

In order to achieve the model study objectives, the project was divided into the following phases:

- model selection,
- collection of model forcing and calibration data,
- creation of the model computational grid,
- creation of model input files,
- model calibration,
- model confirmation (verification and sensitivity analysis), and
- scenario testing.

The remaining chapters describe each of the phases listed above. Chapter 2 introduces the threedimensional model and describes the process by which the computational grid was created. This section also reviews the data sources used to create model input and calibration data files. Chapter 3 provides additional details on the specifications of model inputs that were used to create the various model input files. This is followed in Chapter 4 by a description of hydrodynamic and water quality model calibration. Model evaluation, which included a model confirmation test and sensitivity testing, is described in Chapter 5. The sensitivity of the system to changes in inputs of organic matter or other changes in the system is described in chapter 6. In this chapter are described five separate scenario tests that were performed to investigate how water quality in the estuary might change with changes in organic matter loadings. Discussion of the results and some conclusions that come from the study are provided in Chapter 7.
References are provided in the final chapter.


Figure 1. The Cape Fear River Basin, North Carolina

## 2. MODEL AND SYSTEM DESCRIPTION

### 2.1 Model Description

After a review of the previous monitoring and modeling work done in the Lower Cape Fear River Estuary (e.g. Cahoon et al., 1999; Lin et al. 2006, Mallin et al., 2003, Tetra Tech 2001), it was decided that a three-dimensional water quality model was needed. Because of its flexibility in water quality simulation, its frequent use for hydrodynamic simulation, and its support by EPA, the Environmental Fluid Dynamics Code (EFDC) was selected as the model to be used for this application. EFDC is a hydrodynamic and water quality model that can be used to model one, two or three-dimensional aquatic systems. EFDC uses stretched or sigma vertical coordinates and Cartesian or curvilinear-orthogonal horizontal coordinates to represent the physical characteristics of a water body. It solves the three-dimensional, vertically hydrostatic, free surface, turbulent averaged equations of motion for a variable-density fluid. Dynamically coupled transport equations for turbulent kinetic energy, turbulent length scale, salinity and temperature are also solved. The EFDC model allows for drying and wetting in shallow areas by a mass conservation scheme (Hamrick, 1992).

A water quality model with 21 state variables (Table 1) has been developed and implemented within EFDC to form a three-dimensional coupled hydrodynamic and water quality model (water quality formulation can be found in Park et al., 1995). The model simulates temporal and spatial distributions of several water quality parameters such as dissolved oxygen, suspended algae, several components of carbon, nitrogen, phosphorous, silica as well as fecal coliform bacteria. The model predicts time and space variable concentrations by coupling the fate of these constituents (Figure 2) and performing a mass balance for each constituent at each water quality model time step (Tetra Tech, Inc., 2002a, 2002b).

The complete EFDC model contains three submodels (1: Hydrodynamics and Temperature, 2: Water Quality, 3: Sediment Diagenesis). The complete model, which was originally developed by John Hamrick (Hamrick, 1992) is currently maintained by Tetra Tech, Inc.. The hydrodynamics and temperature submodel is currently supported by the Environmental Protection Agency (EPA). In this application, the hydrodynamics, temperature, and water quality submodels were utilized. The sediment diagenesis model was not used, instead fluxes between the water column and sediments were prescribed.

Table 1. EFDC Water Quality State Variables. Abbreviations refer to constituents as shown in Figure 2.

| Water Quality State Variable | Abbreviation | Unit |
| :---: | :---: | :---: |
| Cyanobacteria (blue-green <br> algae) | Bc | $\mathrm{g} / \mathrm{m}^{3}$ |
| Diatoms (algae) | Bd | $\mathrm{g} / \mathrm{m}^{3}$ |
| Green algae (others) | Bg | $\mathrm{g} / \mathrm{m}^{3}$ |
| Refractory particulate organic <br> carbon | RPOC | $\mathrm{g} / \mathrm{m}^{3}$ |
| Labile particulate organic carbon | LPOC | $\mathrm{g} / \mathrm{m}^{3}$ |
| Dissolved organic carbon | DOC | $\mathrm{g} / \mathrm{m}^{3}$ |
| Refractory particulate organic <br> phosphorus | RPOP | $\mathrm{g} / \mathrm{m}^{3}$ |
| Labile particulate organic <br> Phosphorous | LPOP | $\mathrm{g} / \mathrm{m}^{3}$ |
| Dissolved organic Phosphorous | DOP | $\mathrm{g} / \mathrm{m}^{3}$ |
| Total phosphate | TPO | $\mathrm{g} / \mathrm{m}^{3}$ |
| Refractory particulate organic <br> Nitrogen | RPON | $\mathrm{g} / \mathrm{m}^{3}$ |
| Labile particulate organic <br> Nitrogen | LPON | $\mathrm{g} / \mathrm{m}^{3}$ |
| Dissolved organic Nitrogen | DON | $\mathrm{g} / \mathrm{m}^{3}$ |
| Ammonium | NH | $\mathrm{g} / \mathrm{m}^{3}$ |
| Nitrate Nitrogen | NO | $\mathrm{g} / \mathrm{m}^{-}$ |
| Particulate biogenic silica | SU | $\mathrm{g} / \mathrm{m}^{3}$ |
| Available silica | SA | $\mathrm{g} / \mathrm{m}^{3}$ |
| Chemical Oxygen Demand | COD | $\mathrm{g} / \mathrm{m}^{3}$ |
| Dissolved Oxygen | DO | $\mathrm{g} / \mathrm{m}^{3}$ |
| Total active metal | TAM | $\mathrm{g} / \mathrm{m}^{3}$ |
| Fecal Coliform Bacteria | FCB | $\mathrm{Mpn} / \mathrm{m}^{3}$ |



Figure 2. Schematic Description of the State Variables within the EFDC Water Quality Model.

### 2.2 Description of the Lower Cape Fear River Basin

The Cape Fear River flows for 200 miles through the North Carolina piedmont, crosses the coastal plain, and empties into the Atlantic Ocean near Southport, North Carolina. The river begins near Greensboro and Winston-Salem as two rivers, the Deep River and the Haw River. These two rivers converge near Moncure to form the Cape Fear River. The Black River joins the Cape Fear 15 miles above Wilmington, and the Northeast Cape Fear River enters the system at Wilmington (Figure 1).

The Lower Cape Fear River Estuary occupies the southernmost 60 kilometers of the watershed, between Wilmington and the Atlantic Ocean. The estuary includes not only the tidally affected portion of the Cape Fear River, but also the lowermost portions of the Black and Northeast Cape Fear Rivers. This area of the river basin is extremely important for saltwater animals because of its function as a nursery for juvenile fish, crabs, and shrimp. The Cape Fear River system is North Carolina's largest river system whose basin covers 9,000 square miles and encompasses streams in 29 of the state's 100 counties. It is the most industrialized of all of North Carolina's rivers (Cahoon et al., 1999; Mallin et al., 2003). The river is also an important natural resource that supports many uses including industry, transportation, recreation, drinking water and aesthetic enjoyment.

The hydrology of the Cape Fear Estuary is similar to that of most of the middle to large estuaries located along the US Atlantic Coast. During the summer and the beginning of the fall, the estuary has relatively low flows, whereas it has high flows from the end of the fall to the beginning of the spring. From a hydrodynamic perspective, the Cape Fear has approximately a two-meter tide range and strong tidal currents ( $>0.5 \mathrm{~m} / \mathrm{s}$ ) in the navigational channel of the open estuary and in the narrow tidal river channels of the three tributaries (Ensign et al. 2004).

A region of the estuary near the junction of the Cape Fear, Black, and Northeast Cape Fear Rivers is prone to low dissolved oxygen concentrations during the warm summer and early fall seasons (Mallin et al. 2003) as a result of natural and anthropogenic (both point and nonpoint source) organic matter loadings. Dissolved oxygen concentrations are frequently below saturation concentrations, and typically exhibit an unusual vertical distribution in which surface concentrations are consistently lower than bottom concentrations (Lin et al. 2006). Previous modeling studies of the estuary have replicated this behavior, and have shown reaeration to be an important component of the dissolved oxygen budget for this region (Tetra Tech 2001).

A conceptual model of dissolved oxygen in the Lower Cape Fear River Estuary considers that the mass balance of dissolved oxygen in the water-column is simultaneously affected by several sources (Figure 3). Biologically or chemically degradable organic matter, quantified according to a biochemical oxygen demand load (BOD in kg/day), is introduced into the estuary from a variety of sources including the three major rivers (Cape Fear River, Northeast Cape Fear River, Black River), wastewater treatment plant discharges, and local inputs of non-point source loadings from the fringing wetlands and tidal creeks within the estuary. These same sources may also contribute or deplete the estuaries water column of dissolved oxygen as they either add or remove dissolved oxygen from the water. Likewise the bottom sediments of the estuary, which accepts the organic matter that settles from the water column, will also produce a water column dissolved oxygen sink as the organic matter in the bottom sediments is aerobically decomposed by benthic microbes and animals. Sources of dissolved oxygen to the estuary water column includes phytoplankton that produces oxygen as they photosynthesize new plant material, and the overlying atmosphere, which will replenish the water column with dissolved oxygen whenever surface water dissolved oxygen concentration are less than $100 \%$ saturated. Together these dissolved oxygen sources and sinks and BOD sources produces a dissolved oxygen concentration that varies both in time and space within the estuary. In general the cumulative effects result in dissolved oxygen concentrations that are lowest in the warm summer months when organic matter decomposition occurs at a maximum rate while dissolved oxygen saturation concentrations, which decrease with increasing temperature and salinity, are at their seasonal minimum values.


Figure 3. A Conceptual Model of the Dissolved Oxygen Mass Balance in the Lower Cape Fear River Estuary

### 2.3 Previous Modeling of the Lower Cape Fear River Estuary

A three-dimensional water quality model of the Lower Cape Fear River Estuary has previously been developed for the City of Wilmington and New Hanover County (Tetra Tech 2001). This work had these stated objectives:

- Simulation of the mixing and transport of the City/County existing and proposed future Northside and Southside facility effluents.
- Simulation of the impact of existing and proposed future Northside and Southside facility pollutant loads for oxygen-demanding substances.
- Evaluation of multiple sources and cumulative loads of oxygen-demanding substances to the lower Cape Fear River estuary.
- Analysis of the various processes affecting dissolved oxygen and their relative contribution to ambient dissolved oxygen deficit levels.

Included in the model analysis was a prediction of the near-field and far-field dilution of the two Wilmington wastewater treatment plants and a water quality analysis for various scenarios that considered changes to wastewater treatment plant loadings, changes to sediment oxygen demand values within the estuary, changes to river loadings of organic matter and ammonia, and changes arising from elimination of the fringing wetland areas of the estuary. The hydrodynamic model was calibrated using continuous monitoring of water surface elevation and salinity during the summer of 1993 and long-term measurements of water surface elevation at the NOAA tide gauge at Wilmington. A radiation separation boundary condition was used for the open boundary at the mouth of the Cape Fear River Estuary. The water quality model calibration utilized in-stream data from the Lower Cape Fear Program and wastewater treatment plant loading data from the NC Division of Water Quality. The water quality calibration was based upon model predicted and observed data comparisons for the entire 1998 calendar year. The scenario analysis also considered the 1998 calendar year.

The scenario analyses focused on changes to dissolved oxygen concentration that would be seen at five monitoring stations within the Cape Fear and Northeast Cape Fear Rivers (Navassa, NCF6, M61, M54, M42). The scenario analysis considered time-averaged changes in dissolved oxygen concentrations at these five stations. It was concluded based upon the results of the scenario analyses that sediment oxygen demand had the largest effect on dissolved oxygen concentrations, followed by either the point source loads or the wetland effect, depending on the station examined. Overall the tributary loadings of organic matter had the smallest effect on dissolved oxygen concentrations at the five stations examined, although there were significant differences in the relative magnitudes of the effects at the five stations. The study concluded with recommendations for additional studies and water quality modeling of the estuary aimed at better quantifying the relative magnitude of the factors causing dissolved oxygen depletion.

### 2.4 Model Grid and Bathymetry

The model's three-dimensional numerical grid had 1050 horizontal cells and eight vertical layers. The average horizontal grid size was approximately 600 m ; individual cell side dimensions ranged from 100 to 1300 m (Figure 3). Cell water depths varied for river and estuary channel cells from 2.5 to 12.5 m . EFDC uses a "sigma" vertical grid so that the total water depth within a horizontal cell was subdivided into eight layers of uniform thicknesses.

Of the 1050 horizontal cells in the Lower Cape Fear River, the model grid included 241 "marsh" cells that were used to model the off-channel storage of wetlands adjoining the estuary. The model also has 809 channel cells that provide the majority of transport into and out of the estuary. Horizontal sizes and location of the channel cells were set to match as much as possible the estuary bathymetric information obtained from NOAA (available at
http://NOSDataExplorer.noaa.gov). The size, location, water depth, and bottom roughness of the marsh cells was determined during calibration of the hydrodynamic model. The method for specifying marsh cell characteristics will be described in more detail in the calibration section of this report (see Chapter 4).

The overall distribution of marsh cells within the estuary was initially set according to wetland delineations conducted by the North Carolina Division of Coastal Management NC Division of Coastal Management, 1999). Marsh cells were assumed to be located in regions of the estuary that had two of the wetland types identified in the NC DCM delineation: riverine swamps and saltwater marshes. Of the wetland types identified by the NC DCM, these two represented the majority of wetlands adjoining the Cape Fear River in the estuary. According to the Division of Coastal Management, within the Lower Cape Fear River estuary, saltwater marshes can be found south of Wilmington (Figure 4), while the riverine marshes are generally located farther upstream (Figure 4). The marsh cells in the calibrated EFDC model are also located both upstream and downstream of Wilmington (Figure 5).


Figure 4. Channel Cells Within the Model Region


Figure 5. Riverine Swamps and Saltwater Marshes in the Lower Cape Fear River Estuary


Figure 6. Model Grid Showing Location and Size of Marsh Cells

### 2.4 Monitoring Data Used for Model Setup and Calibration

Monitoring data is needed to define the model system characteristics and system inputs of mass, momentum, and energy. Table 2 summarizes the data types and sources used for setting up and running the Lower Cape Fear River estuary water quality model.

Table 2. EFDC Input Files and Data Sources

| EFDC Input <br> Filename | Description of Data <br> Contained in File | Data Sources |
| :--- | :--- | :--- |
| QSER.INP | Flow time series data at flow <br> specified model boundaries <br> and point source locations | US Geological Survey, NC Division of <br> Water Quality |
| ASER.INP | Meteorological time series <br> data | National Weather Service |
| PSER.INP | Water surface elevation time <br> series data at elevation <br> specified model boundaries | National Oceanographic and <br> Atmospheric Administration (NOAA) |
| TSER.INP | Temperature time series data <br> at all model boundaries and <br> point source inputs | Lower Cape Fear River Program, US <br> Geological Survey, NC Division of <br> Water Quality |
| WSER.INP | Wind time series data | National Weather Service |
| DXDY.INP | Horizontal cell lengths, widths, <br> depths, bottom roughness | National Oceanographic and <br> Atmospheric Administration (NOAA) |
| LXLY.INP | Horizontal cell size location, <br> orientation relative to E-W, N- <br> S direction | National Oceanographic and <br> Atmospheric Administration (NOAA) |
| WQ3DWC.INP | Sediment oxygen demand <br> specification | NC Division of Water Quality |
| SALT.INP | Initial condition for salinity for <br> every model cell and layer | US Geological Survey, NC Division of <br> Water Quality |
| TEMP.INP | Initial condition for <br> temperature for every model <br> cell and layer | Lower Cape Fear River Program, NC <br> Division of Water Quality |
| WQPSL.INP | Time series mass load for each <br> water quality constituent at <br> each flow boundary or point <br> source input | Lower Cape Fear River Program, NC <br> Division of Water Quality |
| CWQSRXX.INP <br> (XX indicates <br> $o n s t i t u e n t ~$ | Time series concentration <br> boundary condition at <br> elevation specified boundaries <br> number) | US Division of Water Quality, Lower <br> Cape Fear River Program |

Much of the water quality information for specifying initial and boundary conditions and calibration data came from the Lower Cape Fear River Program (e.g. Mallin et al. 2003). This water quality monitoring program, which is administered by the Center for Marine Science at UNC Wilmington, collects monthly data at a large number of sites throughout the estuary (Figure 6). The following monitoring data from this program were used for model specification and calibration:

> Physical Parameters: temperature, salinity
> Water Quality Parameters: nitrate-nitrite, ammonium, total nitrogen, orthophosphate, total phosphorus, total kjedahl nitrogen, dissolved oxygen
> Biological Parameters: chlorophyll-a, biological oxygen demand

Additional water quality and physical data were collected using continuous in-situ monitors. This monitoring program was administered by the US Geological Survey, with financial support from the NC Division of Water Quality. The following data, which were collected every 15 minutes at each site, were used for model specification:

Physical parameters: temperature, salinity, water depth or water elevation, streamflow Water Quality Parameters: dissolved oxygen

The continuous monitors were operated for roughly a one-year period beginning in November 2003 at six sites (Figure 7) throughout the estuary. Two of the sites (NE Cape Fear at Wilmington, Marker 12) had a near-surface and near-bottom monitor. The other stations had the monitors installed in the upper portion of the water column. The actual start and stop date of the monitoring differed somewhat for each of the sites. Two additional sites operated by the USGS that were upstream of the model region on the Black and NE Cape Fear Rivers were also used for establishing boundary conditions for flow (Figure 7).

Several other special studies administered by the NC Division of Water Quality collected important additional physical, water quality, and biological data. A short-term special study collected data from 12 sites within the estuary (Figure 7) during July and August 2004 (NC Division of Water Quality 2004). The study data were used along with the Lower Cape Fear River Program data to specify boundary conditions and calibration data for the model. The following data were collected and used in this study:

Field parameters: dissolved oxygen, temperature, salinity
Chemical parameters: 5-day biochemical oxygen demand, 30-day biochemical oxygen , total organic carbon, dissolved organic carbon, chlorophyll, ammonia nitrogen, total kjeldahl nitrogen (tkn), nitrite + nitrate nitrogen, total phosphorus, ortho-phosphorus


Figure 7. Lower Cape Fear River Monitoring Program Monitoring Stations


Figure 8. USGS Continuous Monitoring Stations and the Division of Water Quality Special Study Monitoring Stations

A second special study conducted by the NC Division of Water Quality surveyed river cross-sections at 21 stations throughout the estuary (Figure 8). These data were used to specify the width and depth of each channel cell. A third sampling program performed by the NC Division of Water Quality used chambers to measure in-situ sediment oxygen demand (SOD). Measurements were taken at five separate locations (Figure 10) on five separate days during the Summer and Fall of 2003.

Several other agencies also provided data necessary to specify model forcings or calibration data. Meteorological data (wind speed and direction, cloud cover, air and dewpoint temperature, short-wave solar radiation) were provided by the National Weather Service (NWS). The data were collected by the NWS at the Wilmington Airport. Water surface elevations at several locations (Southport, Sunset Beach, Wilmington) was provided by the National Oceanographic and Atmospheric Administration (NOAA).


Figure 9. Cross-Sections Surveyed by the Division of Water Quality


Figure 10. Division of Water Quality SOD Monitoring Stations

## 3. SPECIFICATION OF MODEL INPUT FILES

Using the complete set of monitoring data described in the previous section, a set of model input data files was created for a time period beginning in November 2003 and ending January 1, 2006. Flow, temperature, and water quality inputs were specified for all of the freshwater that was expected to enter the estuary. These sources included upstream inputs from the rivers ( 3 sources: Cape Fear, Black, and Northeast Cape Fear Rivers), wastewater loadings entering directly to the estuary ( 20 sources), and watershed loadings that enter directly into the estuary ( 14 sources). The overall approach to the specification of input loadings was to consider all loading categories (rivers, tidal creeks, wetlands, wastewater treatment plants), and to quantify these loadings according to the best information available on the particular loading to the estuary. For instance, loadings from tidal creeks downstream of Wilmington and outside of the impaired region were nonetheless estimated based upon available streamflow and/or drainage area data. In addition to this "upstream" source information, water surface elevation and water quality conditions were specified at the model's open boundary near Southport, NC.
Specification of all of these model input data sets is described in Sections 3.1, 3.2, and 3.3. In addition to water inputs from point sources, flow boundaries, and elevation boundaries, the estuary's surface waters also exchange mass, momentum, and energy with the sediments below and the atmosphere above. Quantification of this transport requires another set of model input files. Specification of these model inputs is described in Sections 3.4 and 3.5.

### 3.1 Riverine and Wetland Inputs

### 3.1.1 Flow Specification

Upstream flow boundary conditions for the Cape Fear, Black, and Northeast Cape Fear Rivers were based upon continuous gauging stations operated by the United States Geological Survey at the Cape Fear River Lock \& Dam \#1 (USGS Station 02105769), the Black River near Tomahawk (USGS Station 02106500) and the Northeast Cape Fear River near Chinquapin (USGS Station 02108000). The Cape Fear station was located at the model boundary while the other two stations were located approximately 40 km upstream of the model boundary. To better quantify the river flows entering the modeled region from the Black River and the NE Cape Fear River the NC Department of Environment and Natural Resources (NC DENR) funded the creation of two additional shorter-term flow monitoring sites. The US Geological Survey installed the sites (Black River near Currie, USGS Station 02107544, and NE Cape Fear at Burgaw, USGS Station 02108566) collected the data, and distributed the data via their web site. These data, and additional measurements of drainage area also performed by the USGS were used to establish the needed flow time histories, both at the model boundaries, and for the creeks and wetlands that drain directly to the estuary.

With the extra data collected by the USGS, data from the two sites on the Black River and the two sites on the NE Cape Fear River were pooled to create a continuous time history at the appropriate model boundary. To model the full three-year time period modeled (November

2003 - January 2006) it was necessary to use data from both the upstream and downstream sites for each river to construct the complete flow time history at the model boundary. By examining time periods when flow data were available at both sites (e.g. Black at Tomahawk and Currie), a time lag was established for the more upstream of the two stations. This lag represented the typical travel time of flood waves from the upper to the lower station. The flows from the upstream station after lagging were then scaled according to the drainage area ratios of the two sites. This lagged and scaled upstream site data site was then combined with the downstream site to create the full time history at each of the three upstream model boundaries.

The three flow time histories data were also used to quantify the flow draining directly to the estuary from creeks and riparian wetlands. These flow estimates utilized drainage area measurements of each of the USGS monitoring sites (Table 3). Using these data, the total watershed area of the Cape Fear River was apportioned into seven groupings; three upstream rivers and four areas draining directly to the estuary (Table 4). For each of these groupings one of the flow time histories was used with a scaling constant based upon the corresponding drainage areas. For the four areas draining directly to the estuary, the closest of the three model boundary time histories was used to quantify the flow. The area draining directly to the estuary was further subdivided into a total of fourteen point sources (Figure 11), with the contribution of each source determined according to the particular drainage area of that subwatershed (Table 5).

Table 3. USGS Monitoring Station Data

| Location | USGS <br> Station <br> Number | Drainage <br> Area (mi |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Cape Fear R at Lock \#1 near <br> Kelly | 02105769 | 5255 | Latitude <br> (deg, min, sec) | Longitude <br> (deg, min, sec) |
| Black River Near <br> Tomahawk | 02106500 | 676 | $34,24,15$ | $78,17,38$ |
| NE Cape Fear near <br> Chinquapin | 02108000 | 599 | $34,45,17$ | $78,17,21$ |
| Hood Creek near Leland | 02105900 | 21.6 | $34,49,40$ | $77,50,0$ |
| Black River near Currie | 02107544 | 1405 | $34,24,54$ | $78,08,37$ |
| Cape Fear at Navassa | 02107576 | 7060 | $34,15,36$ | $77,59,15$ |
| NE Cape Fear near Burgaw | 02108566 | 920 | $34,35,54$ | $77,52,31$ |
| NE Cape Fear near <br> Wilmington | 02108690 | 1669 | $34,15,32$ | $77,56,54$ |
| Cape Fear at Marker 12 | 02108820 | 9,000 | $34,04,35$ | $77,55,53$ |

Table 4. Subdivision of the Cape Fear River Watershed

| Area No. | Watershed | Drainage <br> Area (mi $\mathbf{2}^{2}$ | Flow Time <br> History Used <br> 1 |
| :---: | :--- | :---: | :--- |
| 2 | Cape Fear River at Model Boundary | 5,255 | CF @ L\&D 1 |
| 3 | Black River at Model Boundary | 1,405 | Black at Currie |
| 4 | NE Cape Fear at Model Boundary | 920 | NE CF @ Burgaw |
|  | Estuary draining to Cape Fear above NE Cape <br> Fear (upper estuary) | 270 | NE CF @ Burgaw |
| 5 | Estuary draining to Cape Fear below NE Cape <br> Fear (lower estuary | 250 | NE CF @ Burgaw |
| 6 | Estuary draining to Black River | 120 | Black @ Currie |
| 7 | Estuary draining to NE Cape Fear River | 749 | NE CF @ Burgaw |
|  |  | Total | 8,979 |
|  |  |  |  |

Table 5. Subwatersheds Draining Directly to the Estuary

| Point Source Number | Estuary Watershed Name | $\begin{gathered} \text { Drainage Area } \\ \left(\mathbf{m i}^{2}\right) \end{gathered}$ |
| :---: | :---: | :---: |
| 21 | NE Cape Fear, estuary 1 | 31 |
| 22 | NE Cape Fear, estuary 2 | 6 |
| 23 | NE Cape Fear, estuary 3 | 319 |
| 24 | NE Cape Fear, estuary 4 | 101 |
| 25 | NE Cape Fear, estuary 5 | 194 |
| 26 | NE Cape Fear, estuary 6 | 109 |
|  | Total | 760 |
| 27 | Black River, estuary 1 | 91 |
| 28 | Black River, estuary 2 | 30 |
|  | Total | 121 |
| 29 | Cape Fear, upper estuary 1 | 156 |
| 30 | Cape Fear, upper estuary 2 | 31 |
| 31 | Cape Fear, upper estuary 3 | 78 |
|  | Total | 265 |
| 32 | Cape Fear, lower estuary 1 | 61 |
| 33 | Cape Fear, lower estuary 2 | 164 |
| 34 | Cape Fear, lower estuary 3 | 22 |
|  | Total | 247 |



Figure 11. Watersheds Draining Directly into the Cape Fear River Estuary

### 3.1.2 Temperature and Concentration Specification

Specification of temperature and water quality inputs for the riverine sources utilized the monitoring programs described in the previous section. Temperature time histories were available at the Lower Cape Fear Monitoring Program sites on a monthly basis, and from the USGS continuous monitoring sites on a 15 -minute basis. A total of five different monitoring locations were used to specify the temperature time histories. When available, roughly from November 2003 to November 2004, the USGS continuous monitoring data were used. After examining the temporal variability present in the 15 -minute data, it was decided that daily time averaging would be sufficient for purposes of defining the temporal variation in the data. Since the model period (November 2003 - January 2006) extended beyond the period when the continuous monitoring data were available, each temperature time history also utilized data from the Lower Cape Fear River Program. All freshwater sources thus utilized two separate sources to specify temperatures and water quality concentrations (Table 6).

Table 6. Temperature Time History Specification

| Time <br> Series <br> No. | Location | Time Period Using <br> USGS Data and <br> Station Number | Lower Cape Fear <br> River Program <br> Station Used | Point Source <br> Numbers Using <br> This Location |
| :---: | :---: | :---: | :---: | :---: |
| 1 | Black River, Upstream <br> Model Boundary | $9 / 8 / 03-12 / 3 / 04$, <br> 02107544 | B210 | $19,27-28$ |
| 2 | Cape Fear River, Upstream <br> Model Boundary, | $9 / 8 / 03-12 / 3 / 04$, <br> 02105769 | NC11 | 6,18 |
| 3 | Navassa | $9 / 10 / 03-11 / 24 / 04$, <br> 02107576 | Nav | $1-5,9,10,13-17$, |
|  | NE Cape Fear River, | $9 / 10 / 03-12 / 2 / 04$, <br> 02108566 | NCF117 | $20,21-26,29-34$ |
| 4 | Upstream Model Boundary | $11 / 7 / 03-12 / 2 / 04$, <br> 02108820 | M18 | $7,8,11,12$ |
| 5 | Carolina Beach |  |  |  |

Water quality input specification consisted of quantifying both the flow and the time varying mass load for each source and for each water quality model constituent (Table 1). In EFDC, the mass loads from all point sources are quantified in the WQPSL.INP file. This one file gives the time varying mass loads for the riverine sources described in this section, and the point sources described in the next section for every model constituent. The term riverine sources is used rather generally here, and includes not only the three major river inputs to the estuary (Cape Fear, Northeast Cape Fear, Black Rivers), but also the loadings of water quality constituents from the tidal creeks and fringing wetland areas that discharge directly to the estuary. It should be noted here that this approach, which explicitly defines the loading from the entire watershed area draining to any portion of the estuary, is a departure from the method used in the previous modeling study of the Lower Cape Fear River estuary (Tetra Tech 2001), where the wetlands and tidal creeks did not contribute water or water quality constituents to the estuary.

For each scenario, multiple model runs were used to quantify the water quality impact of a particular change in the system (e.g. deepening the navigation channel, reducing wastewater treatment plant discharges, reducing nonpoint source loadings). Many of these scenario runs required changing the WQPSL.INP file, thus it was necessary to create many different versions (approximately 20) of this file. To simply specification of this file, an automated file generation system was created. This system used a set of Excel spreadsheets, one for each source or flow boundary, to quantify the source data time histories and the conversion of these data to mass loads for each constituent. The mass load time histories taken from each spreadsheet were then combined and reformatted to that required by EFDC using a Matlab script (make_wqpsl_capefear.m). The overall procedure for creating the WQPSL.INP is shown in Figure 12.

Use data interpolation and estimation to create a monitoring data set with no data gaps, enter data into Excel spreadsheet, one spreadsheet for each source


For each source, create a data conversion matrix to estimate each model constituent from the available parameters in the source data


For source data given as a concentration time history, multiply concentrations by flows to get mass loads


Collect mass load time histories and reformat, then write into WQPSL.INP file using Matlab script

Figure 12. Procedure for creating water quality mass load file, WQPSL.INP

For the riverine and watershed sources, the mass load time history for each of the 21 model constituents was estimated from a data set that combined information from the Lower Cape Fear River Monitoring Program, the DWQ ambient monitoring program, and the USGS continuous monitoring data. The stations used were the same ones as those used for
specification of the temperature time histories (Table 6). These data were used with data interpolation to create a daily concentration time history of the following eight water quality parameters:

1. dissolved oxygen (mg/L),
2. chlorophyll-a ( $\mu \mathrm{g} / \mathrm{L}$ ),
3. ammonia nitrogen ( $\mathrm{mg} / \mathrm{L}$ as N ),
4. total nitrogen ( $\mathrm{mg} / \mathrm{L}$ as N ),
5. nitrate + nitrite nitrogen $(\mathrm{mg} / \mathrm{L}$ as N$)$,
6. total phosphorus ( $\mathrm{mg} / \mathrm{L}$ as P ),
7. 5-day BOD (mg/L),
8. 20-day BOD (mg/L)

Daily flows at the model boundaries were also available from the USGS data set, as described in section 3.1. Together the flow and water quality information was used to create a daily load (g/s) for seven parameters (5-day and 20-day BOD were used together to estimate ultimate BOD loading). These data served as the basis for estimating the twenty-one model constituents in the EFDC water quality model.

An important consideration in specifying the time history of water quality constituent loading was to accurately quantify the quantity and degradation rate of the various sources of organic matter to the estuary. For instance, the degradation rate of detrital algal matter could be significantly different than that of wastewater discharges, which in itself could be significantly different than the degradation rate of the riverine and wetland sources of organic matter. To provide additional flexibility in specifying organic matter inputs from different categories of sources, two different dissolved organic carbon state variables were used. To avoid changing the EFDC code to accomplish this, the refractory particulate organic matter state variables (RPOC, RPON, RPOP) were used to model refractory dissolved organic matter (RDOC, RDON, RDOP) by specifying a settling velocity of $0.0 \mathrm{~m} / \mathrm{d}$. A second set of organic matter state variables (DOC, DON, DOP) was used to model the labile dissolved organic matter fraction. Different organic matter sources were modeled by using different fractions of the two dissolved organic matter state variables.

With these considerations, the following assumptions were used to estimate the loading time histories for all 21 state variables present in the EFDC model:

- the following 8 constituents were assumed to have zero loading (labile particulate carbon, nitrogen and phosphorus, particulate biogenic silica, available silica, chemical oxygen demand, total active metal, fecal coliform bacteria),
- the carbon to chlorophyll ratio in phytoplankton was assumed to be $60: 1$ ( $\mathrm{mg} \mathrm{C} / \mathrm{mg}$ chl),
- the phytoplankton biomass was assumed to be equally divided among the three algal constituents (diatoms, greens, blue-greens),
- the total organic nitrogen (TON) was calculated as: TON = total nitrogen - ammonia nitrogen - nitrate+nitrite nitrogen,
- total organic carbon in the EFDC model was estimated from ultimate BOD using the stoichiometric ratio for carbon oxidation ( 12 g carbon oxidized requires 32 g of oxygen).

From this total was subtracted the carbon associated with phytoplankton biomass, using measured chlorophyll-a concentrations and the assumed carbon:chlorophyll-a ratio.

- the Redfield ratio, 106:16:1 C:N:P molar ratios (Redfield 1934) was used to convert between dissolved organic phosphorus, nitrogen, and carbon,
- ammonia, nitrate, and dissolved oxygen loadings were estimated directly from the monitoring data,
- two separate dissolved organic matter pools (labile and refractory) were used to model various sources of organic matter (labile organic matter is shown as DOC, DON, and DOP in Figure 2), and
- the refractory dissolved carbon, nitrogen, and phosphorus constituents were included in the model by using corresponding refractory particulate (RPOC, RPON, and RPOP in Figure 2) constituents with an assumed zero settling velocity.

Using these assumptions it was possible to create a $7 \times 21$ parameter matrix that was used to create twenty-one loading time histories, one for each constituent, from the seven available water quality time histories. Matrix multiplication using the parameter matrix also converted each load to the proper units (kilograms per day). Matrix multiplication within Excel was used to do the conversions in an automated fashion once the monitoring time histories were created.

Different parameter matrices were used for the Cape Fear, Black, and Northeast Cape Fear, as the split between labile (first order decay rate $=0.15$ day $^{-1}$ ) and refractory (first order decay rate $=0.03$ day $^{-1}$ ) organic matter was different for these three water bodies. The split between labile and refractory organic carbon was determined from the average BOD20/BOD5 ratio, after correcting for the impact of nitrogenous BOD. A corrected and time-averaged BOD20/BOD5 ratio for each river was used to calculate a corresponding organic matter decay rate and the labile refractory split (Table 7). For instance, the Cape Fear River refractory/labile split was $64.5 \% / 34.5 \%$ based on the BOD5 and BOD20 measurements. For the Cape Fear River, this gave the parameter matrix shown in Table 8.

Table 7. Organic Matter Degradation Parameters for Each Riverine Source

| River | BOD20/BOD5 ratio | First Order Decay <br> Rate $\left(\right.$ day $^{-1}$ ) | Refractory/Labile Split <br> for Organic Matter |
| :---: | :---: | :---: | :---: |
| Cape Fear | 2.58 | 0.68 | $0.645 / 0.355$ |
| Black | 2.70 | 0.60 | $0.690 / 0.310$ |
| Northeast Cape Fear | 3.15 | 0.34 | $0.786 / 0.214$ |

### 3.2 Point Source Inputs

A similar procedure as that described in the previous section was used to define the loading time histories from the point sources that discharge directly to estuary. After reviewing discharge records provided by the NC Division of Water Quality, it was decided that a total of
twenty point sources discharged sufficient organic matter and/or nutrients to justify including them in the model of the Lower Cape Fear River Estuary. These point sources were distributed throughout the estuary both upstream and downstream of Wilmington (Figure 13). As the owner of several of these facilities has changed over the years, the twenty sources were uniquely identified using their DWQ permit number (Table 9). The following sections describe the method for estimating the flow, temperature, and water quality load time histories for these point sources.

### 3.2.1 Flow Specification

Specification of the flows for the twenty wastewater point sources was taken from the discharge monitoring reports, which were provided by NC Division of Water Quality. Flows were generally provided on a monthly basis in million gallons per day. The only changes made to these flows were to convert them from this unit to that used by the model ( $\mathrm{m}^{3} / \mathrm{s}$ ). Unlike the riverine sources, the wastewater data were not interpolated to a daily time step, but instead were left as monthly data.

### 3.2.2 Temperature and Concentration Specification

Temperature time histories were not available in the discharge monitoring reports but were required by the model. These time histories were estimated in the same way as the riverine sources. The five data sources available for the riverine inputs (Table 6) were used for the twenty wastewater sources. For each of these, the closest monitoring location was used to represent the temperature of the water for that source.

The EFDC model also requires a water quality load time history for each of the wastewater sources. This information is provided to the model along with the riverine sources in the WQPSL.INP file. The data format needed by the EFDC model for these two types of sources is identical. Because the data needs and input data were so similar, it was decided to create this information using basically the same procedure as that for the riverine sources (Figure 11).

Table 8. Parameter Matrix for Creating Cape Fear Point Source Loading Time History



Figure 13. Wastewater Treatment Plant Source Locations

Table 9. Description of Point Sources

| Source No. | NPDES <br> Permit No. | Name of Source | Category |
| :---: | :---: | :---: | :---: |
| 1 | NC0000663 | Dupont-Wilmington/Brunswick | WWTP |
| 2 | NC0000663 | Dupont-Wilmington/Brunswick | WWTP |
| 3 | NC0000663 | Dupont-Wilmington/Brunswick | WWTP |
| 4 | NC0001112 | Arteva Specialties-Wilmington | WWTP |
| 5 | NC0001112 | Arteva Specialties-Wilmington | WWTP |
| 6 | NC0003298 | International Paper | WWTP |
| 7 | NC0021334 | Southport, Town-WWTP/Southport | WWTP |
| 8 | NC0023256 | Carolina Beach, Town-WWTP | WWTP |
| 9 | NC0023965 | Wilmington-Northside WWTP | WWTP |
| 10 | NC0023973 | Wilmington-Southside WWTP | WWTP |
| 11 | NC0025763 | Kure Beach WWTP, Town Of | WWTP |
| 12 | NC0027065 | Archer Daniels Midland Co. | WWTP |
| 13 | NC0059234 | Takeda Chemical Products USA | WWTP |
| 14 | NC0059234 | Takeda Chemical Products USA \#2 | WWTP |
| 15 | NC0065676 | Leland Industrial Park WWTP | WWTP |
| 16 | NC0075540 | Belville, Town - WWTP | WWTP |
| 17 | NC0082295 | Fortron Industries | WWTP |
| 18 | NA | Cape Fear River | Upstream River |
| 19 | NA | Black River | Upstream River |
| 20 | NA | NE Cape Fear River | Upstream River |
| 21 | NA | NE Cape Fear, estuary 1 | Estuarine |
| 22 | NA | NE Cape Fear, estuary 2 | Estuarine |
| 23 | NA | NE Cape Fear, estuary 3 | Estuarine |
| 24 | NA | NE Cape Fear, estuary 4 | Estuarine |
| 25 | NA | NE Cape Fear, estuary 5 | Estuarine |
| 26 | NA | NE Cape Fear, estuary 6 | Estuarine |
| 27 | NA | Black River, estuary 1 | Estuarine |
| 28 | NA | Black River, estuary 2 | Estuarine |
| 29 | NA | Cape Fear, upper estuary 1 | Estuarine |
| 30 | NA | Cape Fear, upper estuary 2 | Estuarine |
| 31 | NA | Cape Fear, upper estuary 3 | Estuarine |
| 32 | NA | Cape Fear, lower estuary 1 | Estuarine |
| 33 | NA | Cape Fear, lower estuary 2 | Estuarine |
| 34 | NA | Cape Fear, lower estuary 3 | Estuarine |
| 35 | NC0086819 | Brunswick Co. WWTP | WWTP |
| 36 | NC0001228 | GE - Wilmington | WWTP |
| 37 | NC0003875 | Occidental Chemical | WWTP |

The only differences in the procedure used to create the loading time histories for the wastewater sources resulted from there being a slightly different set of water quality parameters in the discharge monitoring reports. Twenty-day BOD measurements were not available, but for three of the major sources (International Paper, Wilmington Southside WWTP, Wilmington Northside WWTP) there was a long-term BOD study undertaken. In addition, unlike the riverine water quality data set, some of the reported wastewater data were given as flow-weighted concentrations, whereas others were given as a load. For this reason, the conversion parameter matrix used to create the water quality load time histories varied from location to location, and the load calculations were done only when water quality concentrations were given. A complete set of the conversion parameter matrices is available in Appendix A along with the information showing which parameters for which sources required load calculation using the flow record.

Three of the twenty wastewater sources undertook additional monitoring to quantify the degradation rate of their organic matter. Two long-term BOD studies were done on the International Paper wastewater discharge. In addition, long-term BOD measurements were undertaken for the Wilmington Northside and Wilmington Southside municipal wastewater discharges. In all of these studies BOD was measured over approximately 100 days, while additional monitoring was done to measure changes in nitrogen compounds that might result if the waste contained an appreciable nitrogenous biochemical oxygen demand (NBOD).

These long-term BOD data were analyzed by separating the total BOD into carbonaceous and nitrogenous biochemical oxygen demands (CBOD and NBOD). The NBOD exerted was estimated by assuming that each $\mathrm{mg} / \mathrm{L}$ reduction in total Kjeldahl nitrogen in the waste would result in an oxygen demand of $4.57 \mathrm{mg} / \mathrm{L}$. The carbonaceous BOD was estimated as the total BOD exerted minus the NBOD exerted. The NBOD and CBOD were plotted as a function of time, and the following equation was used to fit a line to the CBOD vs. time data:

$$
\begin{equation*}
C B O D(t)=r B O D_{u}\left(1-e^{-k_{d r} t}\right)+r B O D_{u} \frac{k_{d r}}{k_{d r}-k_{d l}}\left(e^{-k_{d l} t}-e^{-k_{d r} t}\right)+l B O D_{u}\left(1-e^{-k_{d l} t}\right) \tag{1}
\end{equation*}
$$

where $\operatorname{CBOD}(\mathrm{t})$ is the carbonaceous BOD exerted at time $\mathrm{t}, \mathrm{rBOD} \mathrm{u}_{\mathrm{u}}$ and BODu are the ultimate BODs of the refractory and labile fractions of organic carbon, and $\mathrm{k}_{\mathrm{dl}}$ and $\mathrm{k}_{\mathrm{dr}}$ are the first-order decay rates of the refractory and labile organic carbon fractions. This equation assumes that the carbonaceous BOD exertion occurs as a two-step process where refractory CBOD is first degraded to labile CBOD, which then degrades while exerting an oxygen demand. This model is consistent with the organic carbon fate processes modeled in EFDC (see DOC and RPOC constituents, Figure 2).

For the International Paper waste, the two long-term BOD showed similar results. For the test done with wastewater collected on July 20, 2003 (Figure 14), the NBOD component was approximately $20 \%$, and the CBODu was approximately $90 \mathrm{mg} / \mathrm{L}$. In the wastewater collected on July 27, 2003 (Figure 15), the NBOD was similar to the earlier measurement, but the $\mathrm{CBOD}_{\mathrm{u}}$ was lower, at about $60 \mathrm{mg} / \mathrm{L}$.


Figure 14. Long-term BOD Measurement, International Paper, July 20, 2003.


Figure 15. Long-term BOD Measurement, International Paper, July 27, 2003.

In both tests of the International Paper long-term BOD, the predicted CBOD time history matched the observed data quite well (Figure 13 and 14). From these tests, the fraction of organic carbon to be refractory was found to be approximately $75 \%$, when using decay rates of labile and refractory organic carbon of 0.15 day $^{-1}$ and 0.03 day $^{-1}$ respectively (Table 10).

Table 10. Degradation Rates and Fractionation of International Paper and Wilmington Municipal Wastewaters

| Location and <br> Test | Labile First <br> Order Decay <br> Rate (day $^{-1}$ ) | Refractory <br> First Order <br> Decay Rate $^{\text {(day }^{-1} \text { ) }}$ | Labile <br> Ultimate <br> BOD <br> $(\mathbf{m g / L})$ | Refractory <br> Ultimate <br> BOD (mg/L) | Fraction of <br> Total Organic <br> Carbon that is <br> Refractory |
| :---: | :---: | :---: | :---: | :---: | :---: |
| IP, July 20, 2003 | 0.15 | 0.03 | 23 | 69 | $75 \%$ |
| IP, July 27, 2003 | 0.15 | 0.03 | 15.5 | 46.5 | $75 \%$ |
| Wilmington <br> Northside | 0.15 | 0.03 | 16.1 | 6.9 | $30 \%$ |
| Wilmington <br> Southside | 0.15 | 0.03 | 0.7 | 0.3 | $30 \%$ |

The results of the long-term BOD measurements of the two Wilmington wastewater discharges looked significantly different from one another, and also different from the long-term BOD measurement of the International Paper wastewater. The test using the sample from Wilmington Northside WWTP discharge (Figure 16) had significantly more NBOD than CBOD. The NBOD exerted was approximately $30 \mathrm{mg} / \mathrm{L}$ after 100 days, but the CBOD at this time was less than $5 \mathrm{mg} / \mathrm{L}$ (Figure 16).


Figure 16. Long-term BOD Measurement, Wilmington Northside WWTP.

The test using a sample from the Wilmington Southside WWTP also had a significantly higher NBOD than CBOD (Figure 17), but both of these values were significantly higher than that seen for the Northside wastewater. The long-term test of the Southside WWTP sample showed an NBOD of nearly $100 \mathrm{mg} / \mathrm{L}$, which was exerted within approximately ten days (Figure 17). The CBOD was lower, and took longer to be fully exerted. For both of these measurements, however, it was possible to model the CBOD exerted curve using the model equation (Eq. 1). Using these long-term BOD measurements, the organic matter in each of these wastewaters was fractionated into labile and refractory components (Table 10), with $30 \%$ of the Wilmington Northside and Southside wastewater assumed to be in the refractory component.


Figure 17. Long-term BOD Measurement, Wilmington Southside WWTP

### 3.3 Downstream Boundary Inputs

The southern boundary of the model grid, at the mouth of the Lower Cape Fear River near Southport, was specified as an elevation boundary (Figure 3). At this boundary there exist four cells ( $\mathrm{I}=30: 33, \mathrm{~J}=5$ ). For these cells, information specifying the water surface elevations, salinity and temperatures, and water quality concentrations were specified for the complete time period that was modeled (November 2003 - January 2006). The following sections describe how each of these boundary conditions was specified in the model.

### 3.3.1 Water Surface Elevation Specification

The downstream elevation boundary condition was specified as the sum of elevation variations due to tidal motions, which were specified with tidal harmonic data, and an elevation time history that was used to represent elevation variations due to weather events. For the tidal harmonic data, a radiation separation boundary condition was used that allows for the outgoing radiation of waves generated within the model region (Bennett, 1976; Blumberg \& Kantha, 1985). The tidal harmonics were determined from 6-min water surface elevation data from the NOAA station at Southport, North Carolina (Station No. 8659084) that were collected from August 1 - December 31, 2006. The data were downloaded from the NOAA's tides and current web site (tidesandcurrents.noaa.gov). The 6-min water surface elevation data were decomposed into tidal harmonic constituents using the FORTRAN program lsqhs.f that was provided with EFDC. The program produced a harmonic description of the elevation data that was dominated by the semi-diurnal $\left(\mathrm{M}_{2}, \mathrm{~S}_{2}, \mathrm{~N}_{2}\right)$ and diurnal $\left(\mathrm{O}_{1}\right)$ components (Table 11).

Table 11. Tidal Harmonic Constituent Data Used to Specify Open Boundary Elevation

| Tidal Component | Period (sec) | Amplitude (m) | Phase (sec) |
| :---: | :---: | :---: | :---: |
| M2 | 44714 | 0.6431 | -11019 |
| S2 | 43200 | 0.0994 | -14422 |
| N2 | 45570 | 0.1328 | -18722 |
| O1 | 92950 | 0.0776 | -43968 |
| M4 | 22357 | 0.0023 | 6837.7 |
| M6 | 14905 | 0.0045 | 6704.8 |
| K2 | 43082 | 0.0296 | 5368.2 |

It was expected that a significant amount of flushing of the system might occur during periods when the water surface elevations in the estuary would either rise or fall due to episodic events such as hurricanes and other major weather events. While the effects of flooding rains would be captured with the flow data, it was believed that the effects of storm surges on the open boundary elevation would not be represented either by the flow time history or by the tidal harmonic forcing. For this reason the open boundary elevation specification also included lowpass filtered elevation data. For this specification it was necessary to have a data set near the model's open boundary at Southport during the time period to be modeled (2003-2005). While a NOAA operated tide gauge was available at Wilmington (Station No. 8658120), it was decided that these data were better used as calibration data since the station is located well within the estuary. In addition, it was felt that using Wilmington for specification of tidal amplitudes might bias the specification of wetland areas within the estuary, which was done iteratively in such a way so as to match the observed attenuation of the tidal elevation signal. Instead, elevation data at Sunset Beach (Station No. 86599897), which is on the Atlantic coast near Southport, were used to for this purpose. Hourly data from January 2003 to December 2006 were taken from the NOAA web site. The hourly data were low-pass filtered using a Butterworth filter (Roberts and Roberts 1978) having a 50 -hour cutoff time period. The cutoff time period was selected to retain the majority of the elevation variations occurring over a time scale consistent with the passage of
storm events. It was found that a 50 -hour cutoff time period filtered out the tidal elevations sufficiently well while allowing longer period elevation variations to remain in the filtered data.

### 3.3.2 Temperature and Concentration Specification

The temperature boundary condition at the southern open boundary was specified in a fashion nearly identical to that of the river boundaries described earlier. A complete time history for the model time period was developed using a combination of sources including the Lower Cape Fear River Program and the continuous monitoring data collection performed by the USGS. The southern open boundary condition for temperature used the fifth time history created for the EFDC temperature time series input file TSER.INP. This time series was also used for some of the wetland sources (Table 6). It was created by combining data from the USGS continuous monitoring station at Marker 12 (Station No. 02108820) while this station was operating (11/7/03-12/2/04), and using the Lower Cape Fear River Program data from station M18 for the other times. The USGS data were used as daily time-averaged data. The LCFRP data were collected at monthly intervals, and were interpolated linearly by EFDC to obtain the necessary temperature time history.

To run the EFDC hydrodynamic model, a salinity time series was also needed for the surface and bottom layer at the open boundary. Since the USGS data station closest to the model boundary (Station No. 02108820, Marker 12) was well removed from the actual boundary at Southport, an estimation procedure was used to create the salinity time history. Near top and near bottom salinities were available for this site, and were used to represent the upper and lower layer salinity. Creation of the boundary condition involved a combination of time-averaging of the 15 -minute salinity data, and selection of a number of maximum values per day to represent the time history for that day. Time averaging and selection of the maximum value were used to as a means of estimating the salinity time history that might be expected at the mouth of the Cape Fear. Hourly time averages were calculated from the 15 -minute data. From this time history were taken $1,2,4$, or 8 maximum salinity values per day, with each value representing the maximum salinity for the appropriate fraction of the day. During model calibration each of these time histories was used and the calibration performance was assessed for each. It was found that the time history using two maximum values per day gave the best fit to the observed data and was therefore used to represent the boundary condition. Linear interpolation was used to calculate the salinities for the intermediate layers. During the time period when the USGS data were not available Lower Cape Fear River Program data from station SPD (Figure 6) was used to represent salinity for EFDC layer 5. Each layer was assumed to have a salinity that differed by 0.5 ppth from its neighbor. This layer-to-layer salinity difference was calculated by taking the difference between the mean salinity surface and bottom salinity values at the USGS Marker 12 station.

Several other water quality time histories were created for representing conditions at the model's southern boundary. Time histories were created for the following constituents:

- cyanophytes (blue-green algae)
- diatoms
- chlorophytes (green algae)
- orthophosphate phosphorus
- ammonia nitrogen
- nitrate+nitrite nitrogen
- dissolved oxygen

All of these time histories used data from the Lower Cape Fear River Program station M18. These data were available on a weekly to monthly time period, depending upon the season. All layers were assigned the same concentration. For the other constituents, a time-averaged value was set in the water quality control file WQ3DWC.INP using time-averaged or estimated values at station M18. The dissolved oxygen time history was estimated by assuming that the incoming water to the model region would be saturated with dissolved oxygen. The saturation concentration depended on water temperature and salinity (Chapra 1997). Water temperatures were taken from the temperature time series file. For this calculation only, the salinity was assumed to be 35 ppth .

### 3.4 Benthic Inputs

The EFDC water quality model includes a predictive sediment diagenesis model. For this study, however, a simpler prescriptive model was used in which benthic fluxes of dissolved oxygen and nutrients were specified as a time history. To better quantify the fluxes of dissolved oxygen between the water column and the benthos, the NC Division of Water Quality measured sediment oxygen demands (SODs) at five sites during the Summer and Fall of 2003 (Table 12).

Five sampling rounds were undertaken as part of the study, in which SOD was measured at one site within the estuary. The sites measured included sites in the upper and lower estuary, and in both the Northeast and Cape Fear Rivers (Figure 9). In general the modeling approach for using these data was to specify the SOD in the simplest manner possible that was sufficient for quantifying the impact of the benthos on the dissolved oxygen budget of the estuary. For this reason, these data were initially used by specifying a constant SOD rate in the water quality control file WQ3DWC.INP. During model calibration it was found that model prediction capability could be improved by varying the SOD as a function of water temperature. It was also found that model prediction of dissolved oxygen were improved by specifying a higher SOD for the Northeast Cape Fear as compared with the remainder of the estuary.

Temporal variation of SOD was considered in the following way. An SOD time history was created by developing a SOD vs. temperature curve that utilized the monitoring data to determine the reference SOD at $20^{\circ} \mathrm{C}$ and the rate at which SOD changes with changing temperature. Spatially averaged but temporally varying water temperatures were then estimated using the monitoring data. Together this information was used to specify the time variation in SOD, which was provided to the numerical model in the benthic flux input file. A non-zero nitrate flux to the sediments was also included, and was considered to vary according to temperature in the same way as SOD. The ratio of nitrate flux to SOD was determined during calibration by comparing model predicted and observed nitrate concentrations within the model region.

Table 12. Sediment Oxygen Demand (SOD) Measurements

| Date | Measurement Location | Latitude \& Longitude | Water Temp. $\left({ }^{\circ} \mathrm{C}\right)$ | SOD at ambient temp (g/m2/d) | SOD corrected to 20 deg C (g/m2/d) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 8/6/2003 | Prince George Creek, tributary to the Northeast Cape Fear River, station located 0.5 miles upstream from its confluence with the Northeast Cape Fear River | $\begin{gathered} \mathrm{N} 34^{\circ} 21^{\prime} \\ 29.11^{\prime \prime} \mathrm{W} 77^{\circ} \\ 57^{\prime} 7.55^{\prime \prime} \\ \hline \end{gathered}$ | 26.3 | 0.5189 | 0.3490 |
| 11/20/2003 | Northeast Cape Fear River, upstream from Wilmington near channel marker 4 | $\begin{gathered} \mathrm{N} 34^{\circ} 16^{\prime} 7.67^{\prime \prime} \\ \mathrm{W} 77^{\circ} 56{ }^{\prime} \\ 58.20^{\prime \prime} \\ \hline \end{gathered}$ | 15.9 | 0.1900 | 0.2460 |
| 10/29/2003 | Cape Fear River downstream from Wilmington near channel marker 61 | $\begin{gathered} \mathrm{N} 34^{\circ} 11^{\prime} \\ 39.30^{\prime \prime} \mathrm{W} 77^{\circ} \\ 57^{\prime} 28.24^{\prime \prime} \\ \hline \end{gathered}$ | 19.3 | 0.4440 | 0.4640 |
| 7/17/2003 | Cape Fear River downstream from Wilmington near channel marker 55 | $\begin{gathered} \mathrm{N} 34^{\circ} 08^{\prime} 7.8^{\prime \prime} \\ \mathrm{W} 77^{\circ} 56^{\prime} 46.3^{\prime \prime} \end{gathered}$ | 27.6 | 0.4679 | 0.2899 |
| 10/2/2003 | Town Creek, tributary to the Cape Fear River, station located 0.5 miles upstream from its confluence with the Cape Fear River. | $\begin{aligned} & \text { N34 }{ }^{\circ} 07^{\prime} 53^{\prime \prime} \\ & \text { W77 } \end{aligned}$ | 20.7 | 0.6951 | 0.6651 |

### 3.5 Meteorological Inputs

All three EFDC sub-models (hydrodynamics, temperature, water quality) require information on meteorological conditions in the modeled region. Several options exist as to which set of data is given to the model. In the Lower Cape Fear River application, these time histories were specified in the ASER.INP and WSER.INP files:

- atmospheric pressure in millibars,
- air temperature in degrees Celsius,
- relative humidity,
- rainfall (a conversion factor is available to handle various rainfall units)
- evaporation rate (a conversion factor is available to handle various evaporation units)
- short-wave solar radiation (given as Watts $/ \mathrm{m}^{2}$, various units can be used)
- cloud cover (given as sky fraction, various units can be used)
- wind speed in $\mathrm{m} / \mathrm{s}$
- wind direction (given as the "to" direction, 0 degrees $=$ to the north)

Data collected from the National Weather Service were used for all values except solar shortwave radiation. Data from the Wilmington International Airport were used and was assumed to represent weather conditions over the entire model region. Evaporation and precipitation were assumed to be in balance, and were therefore assumed both to be zero. Solar short-wave radiation was obtained from the National Renewable Energy Lab's National Solar Radiation Database. Hourly values of solar radiation at the Wilmington Airport (Station No. 723013) for years 2003, 2004, and 2005 were downloaded from the database's web site (www.nrel.gov/rredc/solar_data.html).

## 4. MODEL CALIBRATION

### 4.1 Description of the Calibration Time Period

A fourteen-month time period was used for model calibration (November 1, 2003 January 1, 2005). This period of time was chosen because it offered the best set of monitoring data for establishing model forcings and for assessing model performance. A second time period (January 1, 2005 - January 1, 2006) was used for model confirmation. The model confirmation runs are described in section 5. For model calibration, the first two months of the time period (November 1, 2003 - January 1, 2004) were used for "spinning up" the model to establish initial conditions. Comparison of model predictions to data from the various monitoring programs was done for the period from January 1, 2004 to January 1, 2005.

Calibration was accomplished first using only the hydrodynamic model. Hydrodynamic calibration consisted of comparing model predicted water surface elevations, water velocities, salinities, and temperatures to corresponding values from the monitoring program. The EFDC input files describing the model grid (CELL.INP, DXDY.INP, LXLY.INP) were created as part of the hydrodynamic model calibration. These files are included in this report as Appendix B. The model time step ( 20 seconds), the number and location of all boundary conditions, the locations for model output, and many other parameters needed for running the model were set in the EFDC control file (EFDC.INP). This file is included in this report as Appendix C. Once the hydrodynamic model was calibrated, water quality calibration was conducted.

The model calibration period had streamflows that were generally below average for the first half of the year, and generally above average for the latter half of the year (Figure 18). After a storm in early May 2004, streamflows remained below average until August. The remainder of 2004 and 2005 until the middle of April had streamflows that were near historical average values. Flows through the summer of 2005 were generally below average (Figure 18).

Dissolved oxygen (DO) concentrations at two stations that were in the impaired region and had continuous monitors installed in 2004 (Navassa and the Northeast Cape Fear at Wilmington) showed summertime (April 1 - October 31) dissolved oxygen concentrations that varied between 3 and $7.5 \mathrm{mg} / \mathrm{L}$ (Figure 19). At both of these sites the median DO concentration was below $4.5 \mathrm{mg} / \mathrm{L}$ and the $25^{\text {th }}$ percentile DO concentration was below $4.1 \mathrm{mg} / \mathrm{L}$. DO concentrations were generally lower at the Northeast Cape Fear site, and at this site were generally lower near the bottom of the water column (Figure 19).


Figure 18. Streamflows at the Cape Fear River at the Lock and Dam 1 Station During the Model Time Period

### 4.2 Hydrodynamic Model Calibration

Hydrodynamic model calibration was performed by comparing model predictions to observations at twenty-three different locations where data had been collected by the USGS continuous monitoring program, the Lower Cape Fear River Program, or by NOAA (Table 13). These twenty-one sites span the entire model region and all three rivers that make up the Lower Cape Fear River Estuary (Figure 20). Initially hydrodynamic calibration concerned specification of the model grid. The objective of this part of the calibration task was to locate and size all the channel and wetland cells, and to establish a bottom roughness for each of these cells. Once the model grid was specified hydrodynamic calibration continued by comparing model predictions to salinity and temperature observed data. Hydrodynamic calibration is described in more detail in the following sections.


Figure 19. Dissolved Oxygen Concentrations at the Navassa and Northeast Cape Fear at Wilmington Monitoring Stations

Table 13. Locations Used for Model Calibration

| EFDC Grid <br> Cell Used (I, J) | Station Name | River | Monitoring Data <br> Collected by |
| :---: | :--- | :--- | :--- |
| 4,84 | Lock and Dam 1, 02105769 | Cape Fear | USGS |
| 4,80 | NC11 | Cape Fear | LCFRP |
| 7,67 | Black River at Currie, <br> 02107544 | Black | LCFRP |
| 7,66 | B210 | Black | LCFRP |
| 4,60 | AC | Cape Fear | LCFRP |
| 4,25 | IC | Cape Fear | LCFRP |
| 43,97 | NECF at Burgaw, 02108566 | Northeast Cape Fear | USGS |
| 43,66 | NCF117 | Northeast Cape Fear | LCFRP |
| 44,27 | NCF6 | Northeast Cape Fear | LCFRP |
| 4,11 | Cape Fear at Navassa, <br> 02107576 | Cape Fear | USGS |
| 4,10 | Nav | Cape Fear | LCFRP |
| 22,81 | HB | Cape Fear | LCFRP |
| 43,15 | NECF at Wilmington, <br> 02108690 | Northeast Cape Fear | USGS |
| 18,72 | Brunswick River | Cape Fear | LCFRP |
| 24,68 | M61 | Cape Fear | LCFRP |
| 24,58 | M54 | Cape Fear | LCFRP |
| 25,49 | M42 | Cape Fear | LCFRP |
| 25,47 | M12, 02108820 | Cape Fear | USGS |
| 26,40 | M35 | Cape Fear | LCFRP |
| 23,21 | M23 | Cape Fear | LCFRP |
| 24,11 | M18 | Cape Fear | LCFRP |
| 23,76 | Wilmington Tide Gage, Station | Cape Fear | NOAA |
| 33,5 | Cape Fear at Southport, Station <br> 8658120 | Cape Fear | NOAA |
|  |  |  |  |

### 4.2.1 Modeled and Observed Water Surface Elevations

Hydrodynamic model calibration began by comparing model predictions to observed water surface elevations and depths. This work took advantage of the continuous monitoring data program operated by the USGS and tide gage data available from NOAA. Model predictions of tide varying elevations were compared to observed data at these seven locations:

1. Cape Fear River at Lock and Dam 1 (USGS 02105769)
2. Black River at Currie (USGS 02107544)
3. Northeast Cape Fear River at Burgaw (USGS 02108566)
4. Cape Fear River at Navassa (USGS 02107576)
5. Northeast Cape Fear River at Wilmington (USGS 02108690)
6. Cape Fear River at Marker 12 (USGS 02108690)
7. Cape Fear River at NOAA's Wilmington Tide Gage (NOAA 8658120)

The initial model grid used the Lower Cape Fear River Estuary grid created by Tetra Tech (2001). It was recognized in Tetra Tech's earlier study that additional work needed to be done on the model grid. Specifically, two of the recommendations from the earlier modeling effort were that additional information be gathered on the bathymetry within the estuary, and that additional work be done to quantify the effect of the riparian wetlands within the estuary. Both recommended areas of additional work were undertaken during the hydrodynamic model calibration. Twenty-one river cross-sections (Figure 9) were surveyed by the Division of Water Quality as part of the data collection phase of this modeling project (go to the LCFR model website at www.coe.uncc.edu/~jdbowen/LCFR to view cross-sections). All of this new bathymetric information was incorporated into the updated specification of the model grid.

Additional work was also done to specify the location and size of "wetland" cells that adjoin the main river channel. The overall strategy in determining wetland surface area was to use the information on the attenuation of the tidal amplitude to determine the distribution and overall area of the fringing marshes. Thus the wetland area was determined to some extent as a calibrated value. The wetting/drying routines within EFDC were not used, thus water depths had to be specified to keep the cells wet during spring tides.

This work relied upon wetland delineations performed by the NC Division of Coastal Management (NC Division of Coastal Management, 1999). In the model grid these cells were assumed to be shallower (depth = approximately 1 m ) and have a higher bottom roughness ( $\mathrm{z}_{0}=$ 0.04 m ) than the channel cells. An initial estimate of the location and size of these cells was meant to approximate the surface area of off-channel storage available in two of the wetland types (riverine swamps and saltwater marshes) identified by the Division of Coastal Management (Figure 5). With this distribution of wetland cells as a starting point a sensitivity analysis was performed to investigate how changing the total surface area of these cells would affect the attenuation of incoming tidal waves. Multipliers of $1,2,5,10$ were applied to the width of each marsh cell, and the effect on the M2 component of the tide was calculated for the seven locations listed earlier that span the model region. As expected, it was found that increasing the width of the marsh cells increased the attenuation of the tidal signal (Figure 21). Width ratios of 5 and 10 seemed to over attenuate the tidal signal in the oligohaline portion of the estuary. The width ratio of 2.0 seemed to give the best overall results, but this run showed inadequate damping in the Navassa and Wilmington areas. Three additional cases were tried that used various width ratios, and distributions of wetland cells (Figure 22). In the end, the case with a uniform width ratio of 2.0 , and the " v 1 " version of the wetland cell distribution (shown in green in Figure 22) was chosen as optimal. The final size and distribution of wetland cells has these wetland cells along each river above Wilmington, and also in the area between Wilmington and Town Creek (Figure 6).


Figure 20. Monitoring Stations Used for Comparing Monitoring Data to Hydrodynamic Model Predictions.

With the specification of the model grid set model predictions of elevation were examined during the entire 2004 calibration period and at shorter time periods within the calibration period. The EFDC input files that define the model grid (CELL.INP, DXDY.INP, and LXLY.INP) are available in Appendix B. For the model/data comparison of the full 2004 calibration period, model predictions were printed twice per day and compared with observations at six stations: three near the upstream extent of the model region (Cape Fear at Lock and Dam 1, Black at Currie, NE Cape Fear at Burgaw), two in the impaired region near Wilmington (Cape Fear at Navassa, NE Cape Fear at Wilmington), and one in the lower estuary region south of Wilmington (Marker 12). Of interest was how the model responded to the longer-term changes in water surface elevation that would likely lead to episodes of flushing from the estuary. These flushing events could be seen in the flow record (Figure 18) during late August, September, and October 2004. They would also be expected in the variation of water surface elevation that would occur during the transitions between neap and spring tides. The water elevations at the three most upstream locations show increased elevations during the high flow time period in both the observations and the model predictions (Figures 23 and 24). In general, the model was found to underpredict elevation increases at the Cape Fear and Northeast Cape Fear sites during periods of high river flow. The two sites around Wilmington and the lower estuary site (Navassa, NE Cape Fear at Wilmington, Marker 12) showed elevation differences from spring to neap tide that are well represented by the model (Figures 24 and 25).

To examine elevation fluctuations within a few tide cycles it was necessary to examine shorter periods within the overall calibration time period. Hourly model predictions of elevation were compared to observations at the six sites listed earlier plus one additional site, the NOAA tide gage at Wilmington. The observations at the six sites other than the NOAA tide gage relied upon water measured and logged by the continuous water quality monitors. These pressure transducers were non-vented, thus their water depth measurement were affected by changes in atmospheric measurement. In addition, it seems likely that when the instruments were serviced, they were not always placed back in the water at exactly the depth they occupied previously, which could be seen in the data record as sudden increases or decreases in depth every few weeks. These two factors limited how the observed depth data could be used for comparison to model predictions. Because of the nature of the observed data set, we did not attempt any statistical calibration evaluations of the model's water surface elevation predictions. Rather, we examined the model predictions for evidence that the model was correctly attenuating the tidal wave at various locations within the estuary, and correctly varying tidal ranges through the neap to spring cycle. Two time periods were found to have a sufficiently a long period without disruption of any of the depth data at the six sites under study. These two time periods were January 11-21, 2004 and June $16-23,2004$. Tidal constituents predicted by the model at several sites within the estuary were also compared to corresponding observed data (Table 14) to assess how well the model could match the attenuation and phase lag characteristics seen in the observed data set.


Figure 21. M2 Tidal Amplitude at Seven Stations for Various Wetland Cell Width Scenarios


Figure 22. M2 Tidal Amplitude at Seven Stations for Various Wetland Cell Size Scenarios


Figure 23. Model Predictions and Observed Water Surface Elevations During 2004 at Cape Fear R. Station Lock and Dam 1 and the Black R. at Currie



Figure 24. Model Predictions and Observed Water Surface Elevations During 2004 at NE Cape Fear at Burgaw and Cape Fear at Navassa


Figure 25. Model Predictions and Observed Water Surface Elevations During 2004 at NE Cape Fear at Wilmington and Cape Fear at Marker 12

Examination of the model predictions and observations during January 2004 showed that the model did a good job of simulating the attenuation of the tidal signal from lower to the upper portions of the estuary. For this comparison hourly snapshot model predictions were compared to hourly time-averaged observed data. The time-averaging of the observed data was necessary to reduce the size of the observed data files. Examination of the observed data revealed no systematic bias resulting from the averaging procedure.

The four stations in either the lower estuary (Marker 12) or the Wilmington area (Wilmington tide gage, NE Cape Fear at Wilmington, Navassa) had a close agreement between the observed and predicted tidal amplitudes (Figures 26 and 27). There does seem to be some drift in observed data at the Marker 12 station, and some disagreement in the tidal amplitude at these three stations, but overall the fit is good between model predictions and the observed data. All the stations predict well the timing of peak, although there is a consistent disagreement of about one hour between the timing of the peaks at Navassa (Figure 27). During the model calibration it was found that model predictions of the tidal phase at Navassa and farther upstream at Lock and Dam 1were relatively insensitive to changes in model parameters. Model predictions of tidal phase at this station were not as good as at the remaining three stations analyzed.

The three stations in the upper portions of the estuary (NE Cape Fear at Burgaw, Cape Fear at Lock and Dam1, Black at Currie) also show good agreement between the observed and predicted tidal amplitudes (Figures 28 and 29) during January 2004. As at the Cape Fear at Navassa, the model prediction of tidal phase at the Cape Fear at Lock and Dam1 misses the observed data by about one hour (Figure 29). Phase predictions are better at the Northeast at Burgaw (Figure 28) and the Black at Currie (Figure 29). At each location the model does a good job in simulating the attenuation of the tidal signal as it moves up into the estuary. Tidal ranges at the three upstream sites are less than 0.5 m (Figure 28 and 29), while the tidal range at Marker 12 is generally more than 1.5 m (Figure 26). In general the model slightly overpredicts attenuation in the Cape Fear and Northeast Cape Fear, and correctly simulates it in the Black River (Figures 28 and 29).

A shorter time period in June 2004 (June 16-23) also had an observed elevation data set that was suitable for comparison to model predictions. The same set of seven stations were analyzed by comparing hourly time histories of observed and model predicted water surface elevations. As during the other time period analyzed, the model predicted well the tidal phase at the Wilmington tide gage, at Marker 12, and at the Northeast Cape Fear at Wilmington stations, but underpredicted the tide lag at the Navassa station (Figures 30 and 31). The consistent underprediction of the time lag in the upper portion of the Cape Fear River may have something to do with the effect of the split in the river channel in the Brunswick River area. It is not expected that the relatively slight underprediction of time lag has a negative impact on the model's ability to simulate water quality conditions. As in January 2004, the model underpredicts somewhat the tidal range at Marker 12, but better represents the tidal variation at the three stations around Wilmington: the Wilmington tide gage, the Northeast Cape Fear at Wilmington, and Navassa (Figures 30 and 31).


Figure 26. Model Predictions and Observed Water Surface Elevations During January 2004 (Day $740=1 / 11 / 2004$ ) at Cape Fear R. Marker 12 and the Wilmington Tide Gage


Figure 27. Model Predictions and Observed Water Surface Elevations During January 2004 (Day $740=1 / 11 / 2004$ ) at NE Cape Fear at Wilmington, and the Cape Fear at Navassa.


Figure 28. Model Predictions and Observed Water Surface Elevations During January 2004 (Day $740=1 / 11 / 2004$ ) at NE Cape Fear at Burgaw.


Figure 29. Model Predictions and Observed Water Surface Elevations During January 2004 (Day $740=1 / 11 / 2004$ ) at the Cape Fear at Lock and Dam 1 and at the Black River at Currie.


Figure 30. Model Predictions and Observed Water Surface Elevations During June 2004 (Day $900=6 / 19 / 2004)$ at the Cape Fear at Marker 12 and the Wilmington Tide Gage.


Figure 31. Model Predictions and Observed Water Surface Elevations During June 2004 (Day $900=6 / 19 / 2004$ ) at the NE Cape Fear at Wilmington and the Cape Fear at Navassa.

Comparison of model predictions and observed water surface elevations at the three upstream locations (Northeast Cape Fear at Burgaw, Black at Currie, Cape Fear at Lock and Dam 1) show the model's ability to simulate the attenuation of the tidal signal as it moves through the estuary. All three locations show tidal amplitudes that are much reduced from the other more downstream stations (Figures 32 and 33). As for the January 2004 time period, the model overpredicts somewhat the attenuation at the Cape Fear station, but in this case also underpredicts slightly the tidal amplitude at the Black River site (Figure 33). Tidal phases are predicted well, although there exists again some error in the predicted phase at the Cape Fear site. Both the tidal phase and amplitude are predicted well at the Northeast Cape Fear site (Figure 32).

An additional examination of hydrodynamic model performance was performed by decomposing both observed data and model predictions into tidal frequency components. Tidal amplitudes and phases were calculated for both the model predictions and the observed data at seven stations scattered throughout the model region (Table 14). The numerical model does a good job in predicting the attenuation of the M2 tidal amplitude from the downstream location (e.g. Marker 12) to the upstream locations within the Northeast Cape Fear, Black, and Cape Fear Rivers. Errors in the predicted M2 tidal amplitude are $1-4 \mathrm{~cm}$ in the downstream locations and $1-8 \mathrm{~cm}$ in the upstream locations. Similar model performance is seen for the other tidal constituents (Table 14).

Table 14. Model Predicted and Observed Tidal Constituents at Seven Stations

| Cape Fear River at Lock and Dam1 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.065 | 0.011 | 0.054 | 15790 | -2907 | 18697 | 41.8\% |
| S2 | 0.006 | 0.004 | 0.002 | 10830 | 31689 | -20859 | -48.3\% |
| N2 | 0.003 | 0.041 | -0.038 | 10750 | 21200 | -10450 | -22.9\% |
| 01 | 0.021 | 0.069 | -0.048 | 85620 | 88520 | -2900 | -3.1\% |
| M4 | 0.015 | 0.022 | -0.007 | 9578 | 11804 | -2226 | -10.0\% |
| M6 | 0.005 | 0.013 | -0.008 | 10200 | 14209 | -4009 | -26.9\% |
| K1 | 0.016 | 0.032 | -0.016 | 48760 | 87623 | -38863 | -45.1\% |


| Black River at Currie |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.124 | 0.133 | -0.009 | 11070 | 11315 | -245 | -0.5\% |
| S2 | 0.012 | 0.009 | 0.003 | 8905 | 15073 | -6168 | -14.3\% |
| N2 | 0.013 | 0.023 | -0.010 | 3869 | 10255 | -6386 | -14.0\% |
| 01 | 0.022 | 0.041 | -0.019 | 80220 | 80123 | 97 | 0.1\% |
| M4 | 0.033 | 0.019 | 0.014 | 6393 | 777 | 5616 | 25.1\% |
| M6 | 0.005 | 0.007 | -0.002 | 13300 | 11755 | 1545 | 10.4\% |
| K1 | 0.015 | 0.046 | -0.031 | 42110 | 57624 | -15514 | -18.0\% |


| Northeast Cape Fear at Burgaw |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.090 | 0.168 | -0.078 | 25520 | 20265 | 5255 | 11.8\% |
| S2 | 0.013 | 0.036 | -0.023 | 28190 | 13525 | 14665 | 33.9\% |
| N2 | 0.018 | 0.091 | -0.073 | 21750 | 4167 | 17583 | 38.6\% |
| 01 | 0.008 | 0.114 | -0.106 | 72480 | 94049 | -21569 | -23.2\% |
| M4 | 0.031 | 0.020 | 0.011 | 18460 | 8662 | 9798 | 43.8\% |
| M6 | 0.024 | 0.012 | 0.012 | 11550 | 13397 | -1847 | -12.4\% |
| K1 | 0.026 | 0.117 | -0.091 | 69710 | 72991 | -3281 | -3.8\% |


| Cape Fear at Navassa |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Tidal Amplitude (m) |  | Tidal Phase (sec) |  |  |  |  |
| Comp- <br> onent | Model <br> Predicted | Observed | Difference | Model <br> Predicted | Observed | Difference | Diff. <br> (\% per.) |
| M2 | 0.557 | 0.562 | -0.005 | 41100 | 34391 | 6709 | $15.0 \%$ |
| S2 | 0.066 | 0.060 | 0.006 | 37580 | 32623 | 4957 | $11.5 \%$ |
| N2 | 0.081 | 0.123 | -0.042 | 33640 | 28586 | 5054 | $11.1 \%$ |
| O1 | 0.072 | 0.095 | -0.023 | 56840 | 55404 | 1436 | $1.5 \%$ |
| M4 | 0.026 | 0.031 | -0.005 | 13340 | 2786 | 10554 | $47.2 \%$ |
| M6 | 0.033 | 0.044 | -0.011 | 14210 | 8497 | 5713 | $38.3 \%$ |
| K1 | 0.073 | 0.076 | -0.003 | 19560 | 41820 | -22260 | $-25.8 \%$ |


| Northeast Cape Fear at Wilmington |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.548 | 0.589 | -0.041 | 41030 | 40571 | 459 | 1.0\% |
| S2 | 0.064 | 0.081 | -0.017 | 37360 | 39184 | -1824 | -4.2\% |
| N2 | 0.080 | 0.135 | -0.055 | 33300 | 37145 | -3845 | -8.4\% |
| 01 | 0.072 | 0.141 | -0.069 | 56930 | 48996 | 7934 | 8.5\% |
| M4 | 0.030 | 0.020 | 0.010 | 13650 | 6138 | 7512 | 33.6\% |
| M6 | 0.026 | 0.014 | 0.012 | 14440 | 14044 | 396 | 2.7\% |
| K1 | 0.076 | 0.121 | -0.045 | 19880 | 42858 | -22978 | -26.7\% |


| Cape Fear at Marker 12 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.646 | 0.669 | -0.023 | 36410 | 36955 | -545 | -1.2\% |
| S2 | 0.083 | 0.086 | -0.003 | 31750 | 31720 | 30 | 0.1\% |
| N2 | 0.105 | 0.028 | 0.077 | 28200 | 27412 | 788 | 1.7\% |
| 01 | 0.097 | 0.115 | -0.018 | 50370 | 83053 | -32683 | -35.2\% |
| M4 | 0.024 | 0.036 | -0.012 | 5872 | 23125 | 5872 | 26.3\% |
| M6 | 0.024 | 0.017 | 0.007 | 6672 | 10884 | -4212 | -28.3\% |
| K1 | 0.085 | 0.163 | -0.078 | 32120 | 35812 | -3692 | -4.3\% |


| Cape Fear at NOAA Tide Gage Wilmington |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tidal Amplitude (m) |  |  | Tidal Phase (sec) |  |  |  |
| Component | Model Predicted | Observed | Difference | Model Predicted | Observed | Difference | Diff. (\% per.) |
| M2 | 0.589 | 0.637 | -0.048 | 40000 | 40207 | -207 | -0.5\% |
| S2 | 0.068 | 0.084 | -0.016 | 36270 | 36576 | -306 | -0.7\% |
| N2 | 0.089 | 0.124 | -0.035 | 32350 | 34988 | -2638 | -5.8\% |
| 01 | 0.078 | 0.072 | 0.006 | 55550 | 55541 | 9 | 0.0\% |
| M4 | 0.022 | 0.053 | -0.031 | 12060 | 10217 | 1843 | 8.2\% |
| M6 | 0.020 | 0.002 | 0.018 | 13240 | 10066 | 3174 | 21.3\% |
| K1 | 0.079 | 0.090 | -0.011 | 18820 | 34973 | -16153 | -18.7\% |

### 4.2.2 Modeled and Observed Salinities

To assess the model's ability to simulate the transport of a conservative substance, twice daily predictions of salinity were compared to observed data at twenty-one sites that spanned the entire model region, and included all three rivers within the estuary (Figure 20). The observed data set included information collected by both the US Geological Survey and Lower Cape Fear River Program. The USGS data came from the in-situ water quality monitors that collected salinity data every fifteen minutes. For this model/data comparison, the USGS observed data were hourly averaged, and two data points per day were taken for the calibration data set. The Lower Cape Fear River Program data was collected at more stations but less frequently. The frequency of sampling varied from season to season. The summer sampling frequency was as often as weekly at some stations. Most of the Lower Cape Fear River Program stations were sampled at a monthly interval during the winter months.

While statistical measures of model fit used all the available data, as described above, here we show time history comparison of model predictions and observations only at the USGS in the lower parts of the estuary. The three USGS sites omitted here (Cape Fear at Lock and Dam 1, Black at Currie, Northeast Cape Fear at Burgaw) all had salinities, both observed and predicted that were always near zero. Two of the sites shown here are in the impaired region (Cape Fear at Navassa, Northeast Cape Fear at Wilmington), and both sites show an excellent agreement between the model predicted and observed salinity (Figure 34). At both sites the salinities are quite dynamic, being near zero during the high flow periods late February and March and also September, and increasing to 10 to 15 ppth in the intervening time periods. Swings in salinity of approximately 5 ppth happen in about a week's time in both June and July 2004. The model predictions track the changes seen in the observed data quite nicely. While rather difficult to see with this particular data presentation, it appears that the model underpredicts somewhat the stratification seen in the observed data. At the Northeast Cape Fear station, the observed data often shows surface salinities that are approximately two ppth lower than the bottom values, but model predictions show essentially no difference between the surface and bottom values (Figure 34).


Figure 32. Model Predictions and Observed Water Surface Elevations During June 2004 (Day $900=6 / 19 / 2004$ ) at the NE Cape Fear at Burgaw.



Figure 33. Model Predictions and Observed Water Surface Elevations During June 2004 (Day $897=6 / 16 / 2004)$ at the Black River at Currie and the Cape Fear R. at Lock and Dam 1.


Figure 34. Model Predictions (lines) and Observed (symbols) Salinities During 2004 at the Cape River at Navassa and the NE Cape Fear River at Wilmington.

Salinities at the one station downstream of Wilmington (Cape Fear at Marker 12) are even more dynamic than the two upstream stations. As for the two other stations, salinities approach zero during the high flow periods in Spring and Fall (Figure 35). In mid-May the salinities also approach zero, which does correspond with a high flow event that is seen upstream at the Cape Fear River model boundary earlier in the month (Figure 18). During the summer, a regular variation in salinity can be seen on approximately a two-week period, which likely corresponds to the variations due the neap/spring tide cycle. As at the other two sites the model agrees with the observed data to an excellent extent, showing all of the dynamic features seen in the observed data. Once again, however, it does appear that the model underpredicts somewhat the magnitude of salinity stratification in this lower part of the estuary (Figure 35).


Figure 35. Model Predictions (lines) and Observed (symbols) Salinities During 2004 at the Cape River at Marker 12.

As mentioned earlier, statistical measures of the model's ability to simulate salinity distributions used all of the available salinity observations. A total of 5308 model/data comparisons were made during the 2004 calibration period, with the following statistical measures used to quantify model fit to the data (see Thomann 1982 or Tetra Tech 2001 for explanation of the measures):

- Mean error (absolute and normalized by mean of observations)
- Mean absolute error (absolute and normalized by mean of observations)
- Root mean square error (absolute and normalized by mean of observations)
- Correlation $r^{2}$
- Bias corrected $\mathrm{r}^{2}$, also known as model efficiency

These comparisons indicated that the model slightly overpredicted salinity in the estuary by 0.112 ppth, which corresponds to a normalized mean error of $2.2 \%$ (Table 15). Two measures were used to quantify the typical magnitude of the prediction error. The root mean square error was 2.48 ppth and the mean absolute error (mean of the absolute value of each error) was 1.45 ppth (Table 15). The corresponding normalized errors for these two statistical measures of model fit were $50.1 \%$ and $29.1 \%$. The correlation $\mathrm{R}^{2}$ was used as an aggregate measure of model fit. For salinity the correlation $\mathrm{R}^{2}$ was $87.6 \%$, and the $\mathrm{R}^{2}$ corrected for model bias was $85.7 \%$. This level of calibration performance compares favorable with other hydrodynamic models of North Carolina estuaries (e.g. Bowen 2000).

Table 15. Calibration Statistics, Salinity

| Parameter | Value | Units |
| :---: | :---: | :---: |
| Mean Error (predicted observed) | 0.112 | ppth |
| Normalized Mean Error | 2.2 | \% |
| Root Mean Square Error | 2.48 | ppth |
| Normalized Root Mean Square Error | 50.1 | \% |
| Mean Absolute Error | 1.45 | ppth |
| Normalized Mean Absolute Error | 29.1 | \% |
| Correlation $\mathbf{R}^{2}$ | 87.6 | \% |
| Number of Model/Data Comparisons | 5308 | - |
| $\mathbf{R}^{\mathbf{2}}$ Corrected for Bias | 85.7 | \% |

Statistical measures of model fit were also calculated at each site where observed data were available. Each of the measures given in Table 15 were calculated at fourteen stations where salinity data were collected either by the USGS or as part of the Lower Cape Fear Program (Table 16). Mean errors were quite variable, with relatively large values in the Wilmington area, at stations such as the Northeast Cape Fear at Wilmington and Marker 54. Mean errors at sites upstream of Wilmington generally decreased. Both root-mean-square errors and mean absolute errors were generally more uniform, and were less than 2.0 ppth at most sites (Table 16).

Table 16. Model Calibration Statistics for Salinity at Monitoring Sites Within the Model Region

| Station | Mean Error (ppth) | \% | $\begin{aligned} & \text { RMSE } \\ & \text { (ppth) } \quad \text { \% } \end{aligned}$ |  | Mean Absolute Error (ppth) | \% | $\mathbf{R}^{\mathbf{2}}$ | No. data points |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CF L\&D1 | -0.05 | $100.0 \%$ | 0.05 | 100.0\% | 0.05 | 101.8\% | 0.4\% | 572 |
| Blk Currie | -0.03 | 100.0\% | 0.03 | 100.0\% | 0.03 | 102.1\% | 9.6\% | 653 |
| NE Burgaw | -0.05 | $\begin{array}{r} - \\ 100.0 \% \\ \hline \end{array}$ | 0.05 | 100.0\% | 0.05 | 102.2\% | 20.5\% | 689 |
| NCF6, NECF | 0.13 | 5.9\% | 2.51 | 116.6\% | 3.55 | 165.0\% | 0.0\% | 12 |
| USGS, <br> Navassa | 0.72 | 44.7\% | 1.42 | 88.4\% | 2.35 | 146.4\% | 47.4\% | 510 |
| CF @ Nav | 1.77 | 128.3\% | 2.15 | 155.7\% | 2.81 | 202.9\% | 27.9\% | 12 |
| USGS, NECF Wilm | -1.46 | -29.4\% | 1.94 | 39.1\% | 2.80 | 56.3\% | 68.9\% | 1439 |
| Brnswk Riv | 0.25 | 7.5\% | 1.74 | 52.2\% | 2.54 | 76.2\% | 30.3\% | 12 |
| $\begin{aligned} & \text { CF, Mrkr } \\ & 61 \end{aligned}$ | -0.04 | -0.6\% | 1.92 | 28.2\% | 2.52 | 37.0\% | 71.4\% | 13 |
| M54, CFM54 | 2.34 | 27.6\% | 2.61 | 30.7\% | 3.60 | 42.4\% | 79.7\% | 13 |
| M42, CFM42 | 2.43 | 22.8\% | 2.76 | 25.8\% | 3.35 | 31.4\% | 81.5\% | 13 |
| $\begin{aligned} & \text { USGS, CF } \\ & \text { Mrkr } 12 \end{aligned}$ | 1.76 | 13.6\% | 2.84 | 21.9\% | 3.61 | 27.9\% | 76.8\% | 1344 |
| $\begin{aligned} & \text { M35, } \\ & \text { CFM35 } \end{aligned}$ | 0.55 | 3.8\% | 2.12 | 14.3\% | 2.73 | 18.5\% | 75.9\% | 13 |
| $\begin{aligned} & \text { M23, } \\ & \text { CFM23 } \end{aligned}$ | -3.92 | -18.1\% | 4.80 | 22.1\% | 5.60 | 25.8\% | 43.9\% | 13 |

Scatter plots of observed vs. predicted salinity also give an indication of the relatively good fit of the model to the observed data. Little bias is seen in the predictions overall, with the points shown to be evenly distributed around the 45 -degree line indicating observation $=$ prediction (Figure 36). Model error seems not to vary with the magnitude of model prediction, as the magnitude of the error is similar for relatively low and high salinities. At the low salinity
values, there are many points where zero salinity is predicted but not observed or vice-versa, but there doesn't exist more of one type of error as opposed to the other. The magnitude of the residuals is significant, however, which may be due to the model's relatively poor ability at matching the temporal variations in the extent of salinity intrusion at certain times (see for example Figure 35).


Figure 36. Scatter Plot of Predicted Salinity (ppth, x-axis) vs. Corresponding Observed Salinity (ppth, y-axis).

Percentile plots of observed data and model predictions were also created for the salinity values in the estuary. These percentile plots are useful in assessing how the model predictions values and the corresponding observations are distributed relative to one another. Both model predictions and observed data have approximately $50 \%$ of the salinities at zero (Figure 37). At the higher salinity values the model both slightly under and over predicts salinity and the frequency of a particular salinity value. For instance, at the $90^{\text {th }}$ percentile value the observed salinity is 15 ppth. This same salinity value is seen in about only $85 \%$ of the model predictions. The lower percentile values, where the salinities are lower also show some error, but in this case the model predictions for a particular salinity value are slightly more frequent than seen in the observations. Over the entire range of salinities, however, there is not a very large difference between the observed and predicted salinities (Figure 37).


Figure 37. Percentile Plot of Observed and Model Predicted Salinities During the Calibration Period. The $y$-axis indicates the fraction of values below the corresponding salinity (ppth) indicated on the x -axis.

### 4.3 Temperature Model Calibration

The procedure for assessing the temperature model's calibration performance was very similar to the procedure for salinity. Observations from the twenty-one Lower Cape Fear River Program and USGS stations in the model region (Table 13) were compared. The USGS data were synthesized to produced two hourly-averaged temperatures per day. The Lower Cape Fear River program data were used without any data synthesis. While time histories comparisons were produced at all twenty-one stations, only six representative sites are shown here. Five of the six are the USGS sites where twice daily data are available. Three of the sites are near the upstream model boundary (Cape Fear at Lock and Dam 1, Black River at Currie, Northeast Cape

Fear at Burgaw), and two are in the mesohaline region of the estuary (Navassa, NE Cape Fear at Wilmington, Marker 12). One Lower Cape Fear River Program station from the Northeast Cape Fear River (NCF 117) is shown as well.

The model vs. data time history comparison at the Cape Fear station at Lock and Dam1 is typical of all of the temperature time histories. Model predictions follow the observed data to an excellent extent, with very little difference between model predictions and observed data (Figure 38). In the summertime, the model slightly overpredicts temperatures in the surface layer, but not so in the lower layers at this site. The seasonal variation in temperature from summer to winter is represented very well.

The two stations along the Northeast Cape Fear (Figure 39) and in the mesohaline portion of the estuary (Figure 40) also show an excellent fit between model predictions and observed data. Seasonal variations in temperature are nearly identical between model predictions and observations, and at only one site (Northeast Cape Fear at Wilmington) is there any observable difference between these temperatures. At this one site the model overpredicts summertime temperature by a few degrees.

This excellent degree of model fit is also seen in the other measures of model calibration. The normalized mean error in the model is less than $1 \%$, and the normalized root mean square error is only $5.8 \%$ (Table 17). The correlation $R^{2}$ is $97.9 \%$, indicating an excellent fit of the model to the observed data. Statistical measures of model fit were also calculated at individual stations (Table 18). All stations had calibration $R^{2}$ values above $97 \%$ and root mean square errors of less than $9 \%$ (Table 18). The scatter plot of observed data vs. corresponding model predictions (Figure 41) and the percentile plot (Figure 42) also indicate the excellent ability of the temperature model to match the observed temperatures.


Figure 38. Model Predicted and Observed Temperatures During 2004 at the Cape Fear River at Lock and Dam 1 and the Black River at Currie.


Figure 39. Model Predicted and Observed Temperatures During 2004 at the NE Cape Fear River at Burgaw and at Station NCF 117.


Figure 40. Model Predicted and Observed Temperatures During 2004 at the NE Cape Fear River at Wilmington and the Cape Fear River at Marker 12.

Table 17. Calibration Statistics, Temperature

| Parameter | Value | Units |
| :--- | :---: | :---: |
| Mean Error (predicted - <br> observed) | $-\mathbf{0 . 1 2}$ | Deg C |
| Normalized Mean Error | $\mathbf{- 0 . 6 1}$ | \% |
| Root Mean Square Error | $\mathbf{1 . 1 2}$ | Deg C |
| Normalized Root Mean Square <br> Error | 5.8 | \% |
| Mean Absolute Error | 0.81 | Deg C |
| Normalized Mean Absolute <br> Error | 4.2 | \% |
| Correlation R ${ }^{2}$ | $\mathbf{9 8 . 4}$ | \% |
| Number of Model/Data <br> Comparisons | $\mathbf{6 0 4 0}$ | - |
| R $^{2}$ Corrected for Bias | $\mathbf{9 7 . 9}$ | \% |

Table 18. Model Calibration Statistics for Temperature at Monitoring Sites Within the Model Region.

| Station | Mean Error (ppth) | \% | RMS (ppth) | \% | MAE (ppth) | \% | $\mathbf{R}^{\mathbf{2}}$ | No. data points |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CF L\&D1 | -0.21 | -1.1\% | 0.30 | 1.5\% | 0.58 | 3.0\% | 99.6\% | 573 |
| NC11, CF | -0.17 | -0.7\% | 0.59 | 2.7\% | 0.99 | 4.4\% | 98.3\% | 75 |
| Blk Currie | -0.31 | -1.7\% | 0.63 | 3.4\% | 1.01 | 5.4\% | 98.4\% | 653 |
| B210, Blk | -0.66 | -3.8\% | 0.93 | 5.3\% | 1.26 | 7.2\% | 97.9\% | 13 |
| AC- <br> CF@Acme | -0.44 | -2.0\% | 0.65 | 2.9\% | 0.88 | 3.9\% | 98.9\% | 75 |
| $\begin{aligned} & \text { IC, Cape } \\ & \text { Fear } \end{aligned}$ | 0.26 | 1.1\% | 0.87 | 3.9\% | 1.07 | 4.7\% | 99.1\% | 75 |
| NE Burgaw | 0.73 | 4.1\% | 1.23 | 6.8\% | 1.59 | 8.8\% | 96.9\% | 690 |
| NCF117 | -0.50 | -2.7\% | 1.49 | 8.2\% | 1.84 | 10.1\% | 96.3\% | 13 |
| NCF6, NECF | 0.09 | 0.4\% | 1.78 | 8.8\% | 2.07 | 10.3\% | 98.6\% | 47 |
| Navassa | 0.13 | 0.6\% | 0.76 | 3.8\% | 0.92 | 4.5\% | 99.4\% | 510 |
| CF @ Nav | 0.26 | 1.2\% | 0.87 | 3.8\% | 1.04 | 4.6\% | 99.4\% | 75 |
| $\begin{aligned} & \mathrm{HB}, \\ & \text { CF@HSB } \end{aligned}$ | 0.14 | 0.6\% | 1.06 | 4.7\% | 1.36 | 6.0\% | 98.6\% | 74 |
| NECF Wilm | -0.30 | -1.6\% | 1.16 | 6.1\% | 1.39 | 7.3\% | 99.2\% | 1462 |
| Brnswk Riv | -0.56 | -3.0\% | 0.92 | 4.9\% | 1.29 | 6.9\% | 98.9\% | 39 |
| CF, Mrkr 61 | -0.73 | -3.6\% | 1.24 | 6.1\% | 1.46 | 7.2\% | 99.0\% | 48 |
| M54, CFM54 | -0.63 | -3.1\% | 0.87 | 4.3\% | 1.10 | 5.4\% | 99.4\% | 48 |
| M42, CFM42 | -0.58 | -2.8\% | 0.85 | 4.1\% | 1.15 | 5.7\% | 99.0\% | 48 |
| CF Mrkr 12 | -0.25 | -1.3\% | 0.50 | 2.6\% | 0.66 | 3.4\% | 99.5\% | 1429 |
| M35, CFM35 | -0.58 | -2.8\% | 0.82 | 4.0\% | 1.15 | 5.6\% | 98.7\% | 48 |
| M23, <br> CFM23 | -0.40 | -2.0\% | 0.96 | 4.8\% | 1.22 | 6.1\% | 97.7\% | 45 |



Figure 41. Scatter Plot of Predicted Temperature ( ${ }^{\circ} \mathrm{C}$, x -axis) vs. Corresponding Observed Temperature ( ${ }^{\circ} \mathrm{C}, \mathrm{y}$-axis) During the 2004 Calibration Run.


Figure 42. Percentile Plot of Observed and Model Predicted Temperatures During the 2004 Calibration Period. The y-axis indicates the fraction of values below the corresponding temperature $\left({ }^{\circ} \mathrm{C}\right)$ indicated on the x -axis.

Overall the hydrodynamic model's level of calibration performance is as good, or even better than other hydrodynamic models of North Carolina estuaries (e.g. Bowen 2000). Each component of the hydrodynamic model (hydrodynamics, temperature, conservative transport) fits well the seasonal and event-related dynamics within the system. Examination of time histories in the impaired region indicate the hydrodynamic model is capable of simulating conditions in this part of the estuary. Based upon this evaluation, the hydrodynamic model seems suitable for use as part of the water quality evaluation.

### 4.4 Water Quality Model Calibration

Following the procedures described in section 2, the input files necessary to run the water quality model were quantified. Specifically, the loadings of all constituents from all water sources where flows were specified in the QSER.INP file were quantified in the WQPSL.INP file. A total of 37 sources of water were quantified this way (Table 9), which required quantifying loads for all 21 constituents (Table 1). These loads were calculated with a procedure that utilized conversion matrix tables that quantified how constituent loading was calculated from the available monitoring data. Twenty-three such conversion tables (Appendix A) were needed to quantify these sources. At the model southern open boundary, concentrations were specified either as a time history or in the water quality control file WQ3DWC.INP (see Appendix D).

During calibration, predicted dissolved oxygen concentrations were compared to observed data at the same twenty-one stations used for hydrodynamic model calibration (Table 14, Figure 19). The USGS continuous monitoring data were synthesized to give two hourly averaged dissolved oxygen values per day. The Lower Cape Fear River Program data were generally collected monthly during the non-summer months, and bi-weekly during the summer. These twice-daily data were combined with the LCFRP data to create an extensive calibration data set that had frequent measurements of DO at stations throughout the estuary and somewhat less frequent measures of other water quality constituents. During calibration, kinetic values in the water quality model were adjusted to improve the fit between model predicted water quality concentrations and observed data. Model/data comparisons were made of the following water quality measures:

- Dissolved oxygen (mg/L)
- Total chlorophyll-a ( $\mu \mathrm{g} / \mathrm{L}$ )
- Total phosphorus (mg/L as P)
- Ammonia ( $\mathrm{mg} / \mathrm{L}$ as N )
- Nitrite/nitrate (mg/L as N)

Multiple measures were used to quantify the model's fit to the observed data, just as was done during hydrodynamic calibration. As described earlier the following measures were analyzed for each model run during calibration, (see Thomann 1982 or Tetra Tech 2001 for explanation of the measures):

- Mean error (absolute and normalized by mean of observations)
- Mean absolute error (absolute and normalized by mean of observations)
- Root mean square error (absolute and normalized by mean of observations)
- Correlation $r^{2}$
- Bias corrected $\mathrm{r}^{2}$, also known as model efficiency

The calibration had the following multiple objectives

- Use water quality kinetics parameter values that fall within the expected range of values for temperature estuaries (Bowie et al. 1985),
- Produce predicted water quality concentrations that qualitatively agree with the expected seasonal and spatial variations for temperature estuaries,
- Produce predicted water quality concentrations that simultaneously minimized mean error, mean absolute error, and root mean square error
- Produce predicted water quality concentrations that maximum correlation $r^{2}$ and bias corrected $\mathrm{r}^{2}$,
- Produce predicted water quality concentrations with a cumulative frequency distribution that was a close approximation of the observed distribution over the full range of percentiles.

A phased approach to water quality calibration was undertaken. The calibration phasing was done in such a way that recognized the expected relative importance of processes in the estuary's dissolved oxygen budget, and the inherent dependencies between water quality constituents. Thus, special importance was given to maximizing the model's fit to observed dissolved oxygen concentrations. Calibration phasing was done as follows:

- Calibrate first the sediment oxygen demand to give a reasonably low mean error for dissolved oxygen concentrations,
- Calibrate the nitrification rate to give an acceptable value for the mean error for ammonia concentrations within the estuary,
- Calibrate the phytoplankton group specific maximum growth rates, temperature optima, and half-saturation constants for nutrient concentrations to give a reasonable seasonal pattern of total chlorophyll-a concentrations,
- Calibrate the phytoplankton stoichiometric ratios (C/Chl, C/N, C/P) to give acceptable mean errors for total P and total N concentrations
- Calibrate the benthic flux rates for nitrate/nitrite to give an acceptable value for the mean error of nitrite/nitrate concentrations in the water column,
- Calibrate again the sediment oxygen demand to give acceptable error statistics and cumulative frequency distributions for dissolved oxygen,
- Check the statistical measures of model fit for the remaining water quality constituents, and repeat the calibration procedure if warranted.

At the conclusion of this calibration procedure, a complete set of calibrated water quality kinetic parameters was produced for the base case calibration run. These water quality kinetic parameters were then used for all subsequent model confirmation and scenario testing model runs. All of these parameters except SOD are specified in EFDC's water quality control file (see file WQ3DWC.INP, Appendix D). Benthic fluxes are specified in a separate file (benflux_tser.inp, Appendix E).

Of particular interest in the water quality modeling was the point source loadings of oxygen demanding wastes. In the EFDC water quality model as applied here, three EFDC constituents present in the point source loadings accounted for the water-column oxygen demand, refractory dissolved organic carbon (RDOC - in table 1 RDOC is shown as RPOC), dissolved organic carbon (DOC), and ammonia. The RDOC load was found to come primarily from the Cape Fear River inflow, and to a lesser extent the estuary inflows (Figure 43). The

## RDOC Load (kg/D)



Figure 43. Average Daily Load of Refractory Dissolved Organic Carbon to the Model Region from Various Sources
source contribution to the RDOC loading was relatively small (Figure 43). Likewise, more than half of the DOC load came from the Cape Fear River inflow, while the other four categories of sources (Black River, NE Cape Fear River, estuary inflows, and point sources) were of roughly equal magnitude (Figure 44). The ammonia loading distribution was quite different, with point sources contributing more than half of the total loading, and only small contributions from the Black and Northeast Cape Fear rivers and the estuary inflows (Figure 45).

An additional sink of dissolved oxygen results from the transport of water-column oxygen to satisfy the sediment oxygen demand (SOD). As described in previous sections, an SOD measurement program conducted by the NC Division of Water Quality collected SOD measurements at five locations (Figure 9) during the Summer and Fall of 2003 (Table 13). Initially these data were used to quantify a time-averaged SOD rate that was applied over the entire model grid. During calibration it was found that this strategy resulted in overpredictions of dissolved oxygen concentration in the summer and underprediction during the winter. For this reason, it was decided to introduce a temperature dependence to the SOD specification. The SOD data were fit with the following temperature dependent model:

$$
\begin{equation*}
\operatorname{SOD}(t)=\operatorname{SOD}(20) \times \Theta^{T-20} \tag{2}
\end{equation*}
$$

DOC Load (kg/D)


Figure 44. Average Daily Load of Dissolved Organic Carbon to the Model Region from Various Sources

NH4 Load (kg/D)


Figure 45. Average Daily Load of Ammonia to the Model Region from Various Sources.
where $\operatorname{SOD}(\mathrm{t})$ and $\operatorname{SOD}(20)$ is the sediment oxygen demands at the desired temperature and at $20^{\circ} \mathrm{C}$, and $\Theta$ is an empirical temperature multiplier. When fit to this model (Figure 46), the SOD data produced a $\operatorname{SOD}(20)$ of $0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ and a temperature multiplier of 1.058 . Using this SOD model and spatially averaged observed temperatures, an SOD time history file was created for input to the model. Using this approach it was found during calibration that the model consistently overpredicted dissolved oxygen concentrations in all three of the Northeast Cape Fear stations. This overprediction was confined to these stations, which indicated that SOD values for these stations should be increased for the cells within the Northeast Cape Fear river basin. During calibration an additional factor was introduced that gave the ratio of SOD in the NE Cape Fear and the rest of the model region. This factor was determined during calibration to give the best fit with a value of 1.5 . In this way SODs were given temperature and spatial variation during the model


Figure 46. Natural Logarithm of Measured Sediment Oxygen Demand (y-axis) vs. Deviation from 20 Degree ${ }^{\circ} \mathrm{C}$ Measured Temperature (x-axis).

Time history comparisons of observed vs. predicted dissolved oxygen concentrations were used to assess water quality model calibration performance. Here we show these time histories at seven stations, the six USGS stations used previously, and one LCFRP station in the Northeast Cape Fear River (NCF117) that was presented earlier when showing temperature time
histories. In general the time history comparisons show an excellent calibration performance for dissolved oxygen. At the two of the USGS sites in the oligohaline sites (Cape Fear at Lock and Dam 1, Black at Currie), the model closely tracks the seasonal variation in dissolved oxygen, whereby the summertime DO concentrations are low and the wintertime DO concentrations are relatively high (Figure 47). As seen in the temperature time history, some density stratification seems to be present at the Black River site, which seems to produce a relatively low DO concentration in the bottom waters (Figure 47). The surface DO concentration value, however, follows the observed data closely throughout the 2004 calibration period (Figure 47).

The two oligohaline stations along the Northeast Cape Fear produced interesting model results. The upper site (NE Cape Fear at Burgaw) had model predictions and observations that agreed well with one another, but at the more downstream site (NCF 117), the model generally overpredicted the DO concentration (Figure 48). During the summer months at this station the observed DO concentrations were approximately four $\mathrm{mg} / \mathrm{L}$ while the model predictions were approximately six $\mathrm{mg} / \mathrm{L}$. A similar discrepancy was seen further downstream along the NE Cape Fear at the Wilmington site (Figure 49). Once again the observed data, this time from a USGS continuous water quality monitor showed DO concentration through the summer of approximately $4.0 \mathrm{mg} / \mathrm{L}$. The model predictions are higher, but are closer to the observed data than at station NCF 117. The overprediction of DO concentration here is approximately $1 \mathrm{mg} / \mathrm{L}$ (Figure 49).

At the other station in the impaired area (Cape Fear at Navassa), the model fits the observed DO concentration very well (Figure 49). Both summertime and wintertime conditions are predicted well, as are more short-term changes during the summer months. The observed concentrations at the station south of Wilmington (Marker 12), showed significantly more variation in dissolved oxygen (Figure 50) than the other two sites in the mesohaline portion of the estuary. Nonetheless, the model did a good job of simulating the dynamic changes in the DO concentrations that occurred throughout the 2004 calibration period (Figure 50). In general the model predictions are somewhat less variable than the observed data, but the model does do a particularly good job of simulating the temporal changes in the minimum dissolved oxygen concentrations that occur through the summer months (Figure 50).


Figure 47. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During 2004 at the Cape Fear River at Lock and Dam1 and the Black River at Currie.


Figure 48. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During 2004 at the NE Cape Fear River at Burgaw and Station NCF 117.


Figure 49. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During 2004 at the Cape Fear River at Navassa and the NE Cape Fear River at Wilmington.


Figure 50. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During 2004 at the Cape Fear River at Marker 12.

Statistical measures of model fit demonstrate the model's ability to simulate the temporal and spatial differences in DO concentration. As for the hydrodynamic model calibration, the calibration statistics are based on model/data comparisons at all twenty-one sites. The USGS sites use hourly average values that have been sampled at two times per day, at twelve hour intervals. The Lower Cape Fear River Program data comes from boat-based sampling that occurred at a variable frequency - as often as weekly during the summer months and stretching to monthly during the winter months. A total of 5,274 model/data comparisons were used to assess model calibration performance for dissolved oxygen. The mean error for dissolved oxygen concentration (predicted - observed) was $-0.06 \mathrm{mg} / \mathrm{L}$, which represented a mean error of $0.86 \%$ (Table 19). Root mean square errors and mean absolute errors were $0.90 \mathrm{mg} / \mathrm{L}$ and 0.69 $\mathrm{mg} / \mathrm{L}$ respectively, which gives normalized errors of $13.4 \%$ and $10.3 \%$ respectively. The correlation $\mathrm{R}^{2}$ for dissolved oxygen model predictions relative to the observed data was $85.4 \%$, or $83.5 \%$ when the correlation coefficient is corrected for model bias (Table 19). In general this is excellent calibration performance for an estuary dissolved oxygen model.

Table 19. Calibration Statistics, Dissolved Oxygen

| Parameter | Value | Units |
| :--- | :---: | :---: |
| Mean Error (predicted - <br> observed) | $\mathbf{0 . 0 0 5}$ | $\mathbf{g} / \mathbf{m}^{\mathbf{3}}$ |
| Normalized Mean Error | $\mathbf{0 . 0 6}$ | $\mathbf{\%}$ |
| Root Mean Square Error | $\mathbf{0 . 9 2}$ | $\mathbf{g} / \mathbf{m}^{3}$ |
| Normalized Root Mean <br> Square Error | $\mathbf{1 3 . 8}$ | $\mathbf{\%}$ |
| Mean Absolute Error | $\mathbf{0 . 7 3}$ | $\mathbf{g} / \mathbf{m}^{\mathbf{3}}$ |
| Normalized Mean Absolute <br> Error | $\mathbf{1 3 . 8}$ | $\mathbf{\%}$ |
| Correlation R 2 | $\mathbf{\%}$ |  |
| Number of Model/Data <br> Comparisons | $\mathbf{8 4 . 4}$ | - |
| $\mathbf{R}^{2}$ Corrected for Bias | $\mathbf{5 2 7 4}$ | $\mathbf{\%}$ |

Statistical measures of model fit were also calculated at individual stations (Table 20). The mean error statistics show that the majority of the stations in the Black and Cape Fear Rivers had predicted dissolved oxygen concentrations that were generally under predicted with respect to observed values, while the predictions in the Northeast Cape Fear were slightly above the observed values. Most stations had a mean error of less than $0.5 \mathrm{mg} / \mathrm{L}$. Root mean square errors were similar to that seen for the cumulative data ( $13.4 \%$, Table 19). Correlation $r^{2}$ values ranged from a low of $67.3 \%$ at station B210 to a high of $96.3 \%$ at Marker 61 in the Cape Fear. Correlation r 2 in the impaired region of the estuary near Wilmington were generally near or above $90 \%$ (Table 20).

The scatter plots of observed vs. predicted dissolved oxygen concentration also indicate the excellent agreement between model prediction and observation. There is little or no bias in the model predictions, and the residuals appear to be of uniform magnitude across the range of predicted dissolved oxygen concentrations (Figure 51). There are a few values in which the model badly overpredicts DO concentration, but compared to the more than 5,000 model/data comparisons, the number of badly predicted DO concentrations is very small (Figure 51).

Equally impressive is the close agreement between model predictions and observations as seen in the percentile plot. The distribution of DO concentrations in the model predictions matches almost exactly that seen in the observations (Figure 52). Of special interest is the fact the number of relatively low DO concentrations, say for example the percentage of values below $4.0 \mathrm{mg} / \mathrm{L}$ is near identical $(10 \%)$ for both the model predictions and the observed data. Only at the very lowest DO concentrations is there any discrepancy between the two distributions. At these lowest DO concentrations, it appears the observed data have slightly higher percentiles, although the magnitude of the differences (e.g. $0 \%$ vs. $3 \%$ at $3.0 \mathrm{mg} / \mathrm{L}$ ) is relatively small (Figure 52).

Table 20. Model Calibration Statistics for Dissolved Oxygen at Monitoring Sites Within the Model Region.

| Station | Mean Error (ppth) | \% | RMS (ppth) | \% | MAE (ppth) | \% | $\mathbf{R}^{\mathbf{2}}$ | No. data points |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CF @ L\&D1 | -0.18 | -2.3\% | 0.66 | 8.3\% | 0.89 | 11.2\% | 81.8\% | 509 |
| NC11, CF | -0.87 | $11.1 \%$ | 1.11 | 14.1\% | 1.31 | 16.6\% | 79.8\% | 75 |
| BIK. @ Currie | -0.62 | $10.4 \%$ | 0.88 | 14.7\% | 1.13 | 18.8\% | 85.8\% | 571 |
| B210, Blk | -1.43 | $19.9 \%$ | 1.59 | 22.2\% | 2.02 | 28.2\% | 66.9\% | 13 |
| CF @ Acme | -0.76 | 10.2\% | 0.96 | 12.8\% | 1.22 | 16.4\% | 81.7\% | 75 |
| IC, CF | -0.05 | -0.7\% | 0.61 | 9.8\% | 0.76 | 12.1\% | 90.4\% | 75 |
| NE Burgaw | -0.18 | -2.7\% | 0.82 | 12.1\% | 1.09 | 16.1\% | 85.3\% | 671 |
| NCF117 | 1.01 | 15.7\% | 1.05 | 16.4\% | 1.32 | 20.6\% | 92.7\% | 13 |
| NCF6, NECF | -0.71 | 11.1\% | 1.03 | 16.0\% | 1.30 | 20.3\% | 82.5\% | 47 |
| Navassa | 0.23 | 3.7\% | 0.57 | 9.3\% | 0.68 | 11.1\% | 93.7\% | 433 |
| CF @ Nav | -0.43 | -6.8\% | 0.68 | 10.9\% | 0.82 | 13.0\% | 92.3\% | 75 |
| CF @ HSB | -0.15 | -2.6\% | 0.61 | 10.1\% | 0.76 | 12.5\% | 91.5\% | 74 |
| NECF Wilm | 0.74 | 11.7\% | 0.78 | 12.2\% | 0.90 | 14.1\% | 95.6\% | 1297 |
| Brnswk Riv | -0.45 | -6.0\% | 0.73 | 9.7\% | 0.99 | 13.1\% | 89.3\% | 39 |
| CF @ M61 | 0.11 | 1.6\% | 0.46 | 6.8\% | 0.58 | 8.6\% | 95.7\% | 48 |
| CF @ M54 | -0.24 | -3.4\% | 0.62 | 8.9\% | 0.74 | 10.6\% | 90.0\% | 48 |
| CF @ M42 | -0.53 | -7.3\% | 0.80 | 11.0\% | 0.92 | 12.7\% | 86.0\% | 48 |
| CF Mrkr 12 | -0.16 | -2.3\% | 0.59 | 8.5\% | 0.74 | 10.8\% | 82.4\% | 1070 |
| CF @ M35 | -0.54 | -7.3\% | 0.79 | 10.7\% | 0.93 | 12.6\% | 83.4\% | 48 |
| CF @ M23 | -0.40 | -5.3\% | 0.62 | 8.2\% | 0.75 | 10.0\% | 84.5\% | 45 |



Figure 51. Scatter Plot of Predicted Dissolved Oxygen Concentrations (mg/L, x-axis) vs. Corresponding Observed Dissolved Oxygen Concentrations (mg/L, y-axis) for the 2004 Calibration Period.


Figure 52. Percentile Plot of Observed and Model Predicted Dissolved Oxygen Concentrations During the Calibration Period. The $y$-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

On the following pages time histories comparisons are shown for several other key constituents in EFDC's water quality model:

- Ammonia nitrogen
- Nitrate+nitrite nitrogen
- Total phosphorus
- Total chlorophyll-a

These constituents were chosen for presentation because they represent key information on how organic matter is cycling between organic and inorganic pools within the estuary. Time histories are shown for each of these constituents at the following six sites:

1. Cape Fear River at NC 11
2. Black River at B210
3. Northeast Cape Fear at NCF6
4. Cape Fear River at IC
5. Cape Fear River at Navassa
6. Cape Fear River at Marker 23

These six sites, all monitored by the Lower Cape Fear River Program span the entire model region. Each set of model predicted vs. observed data time histories shows the water quality model's capability to simulate water quality dynamics within the system.

Ammonia concentrations at the six sites (Figures 53-55) are generally less than $0.5 \mathrm{mg} / \mathrm{L}$ but are occasionally higher in the summertime. The monitoring data at certain sites (e.g. Black River site B210, and Northeast Cape Fear site NCF6) shows a relatively large amount of sample-to-sample variability. In general the monitoring data are more variable in time than the model predictions. There are no significant spatial or temporal trends in either the monitoring data or model predictions. For instance, the lowest concentrations were seen upstream in the Black River and downstream at Marker 23 (Figures 53 and 55).

Nitrate + nitrate (NOx) concentrations (Figures 56-58) are generally higher in the oligohaline portion of the estuary and are always higher than total phosphorus concentrations (Figure 59-61). The lowest NOx concentrations are at station M23 (Figure 58). The model does a good job of representing these spatial trends. There are no detectable temporal trends in the monitoring data for nitrate + nitrite. Total phosphorus (TP) concentrations show a similar spatial trend, with the highest total phosphorus concentrations in the oligohaline portion of the estuary and the lowest concentrations near the estuary mouth. As for ammonia, the monitoring data shows a large amount of sample variability in total phosphorus concentration. There is no detectable temporal trend either in the monitoring data or the model predictions, although there do seem to be episodic increases in TP concentrations. As for ammonia, the model predictions are less variable in time than the observed data (Figures 59-61).

Total chl-a concentrations are always than $10 \mu \mathrm{~g} / \mathrm{L}$, but do show some significant increases during summertime events (Figures 62-64). Model predictions also show temporal variability, with two distinct peaks occurring at each station, although interestingly, the timing of the peak differs from station to station. For instance at station M23 the peaks occur in the March and July (Figure 64), whereas in the Northeast Cape Fear the peaks occur in May and October. These differences in the timing of the chlorophyll peaks are probably related to differences in timing of nutrient supply between these two areas of the estuary. The overall level of variability in the observations and in the model predictions are similar to one another (Figures 62-64).

Overall the water quality model's calibration performance was considered acceptable for its use as part of a dissolved oxygen TMDL. Calibration performance for the DO model was excellent, with statistical measures of calibration performance indicating that the model does an excellent job of capturing the observed spatial and temporal variability in dissolved oxygen concentrations. Correlation $r^{2}$ values at individual sites and overall are comparable, if not better, than those expected for estuarine water quality models. The model's predictions of other water quality constituents are also good. While for these constituents, the model's capability to match observed spatial and temporal patterns is not as good as for dissolved oxygen, the model is able to match the overall concentration level seen across the estuary during the calibration period. Together this information provides confidence that the model is suitable for its intended use, to estimate the system's dissolved oxygen response to changes in the discharge of oxygen demanding waste.


Figure 53. Model Predicted (lines) and Observed (symbols) Ammonia Concentrations During 2004 at the Cape Fear River at Station NC 11 and the Black River at Station B210.


Figure 54. Model Predicted (lines) and Observed (symbols) Ammonia Concentrations During 2004 at the NE Cape Fear River at Station NCF6 and the Cape Fear River at Station IC.


Figure 55. Model Predicted (lines) and Observed (symbols) Ammonia Concentrations During 2004 at the Cape Fear River at Navassa and at Station M23.


Figure 56. Model Predicted (lines) and Observed (symbols) Nitrate + Nitrite Concentrations During 2004 at the Cape Fear River at Station NC 11 and the Black River at Station B210.


Figure 57. Model Predicted (lines) and Observed (symbols) Nitrate + Nitrite Concentrations During 2004 at the NE Cape Fear River at Station NCF6 and the Cape Fear River at Station IC.


Figure 58. Model Predicted (lines) and Observed (symbols) Nitrate + Nitrite Concentrations During 2004 at the Cape Fear River at Navassa and at Station M23.


Figure 59. Model Predicted (lines) and Observed (symbols) Total Phosphorus Concentrations During 2004 at the Cape Fear River at Station NC 11 and the Black River at Station B210.



Figure 60. Model Predicted (lines) and Observed (symbols) Total Phosphorus Concentrations During 2004 at the NE Cape Fear River at Station NCF6 and the Cape Fear River at Station IC.



Figure 61. Model Predicted (lines) and Observed (symbols) Total Phosphorus Concentrations During 2004 at the Cape Fear River at Navassa and at Station M23.


Figure 62. Model Predicted (lines) and Observed (symbols) Chl-a Concentrations During 2004 at the Cape Fear River at Station NC 11 and the Black River at Station B210.


Figure 63. Model Predicted (lines) and Observed (symbols) Chl-a Concentrations During 2004 at the NE Cape Fear River at Station NCF6 and the Cape Fear River at Station IC.


Figure 64. Model Predicted (lines) and Observed (symbols) Chl-a Concentrations During 2004 at the Cape Fear River at Navassa and at Station M23.

## 5. MODEL EVALUATION

### 5.1 Model Confirmation

A model confirmation run was conducted to test the calibrated model's ability to simulate water quality conditions during a time period that was not used for model calibration. The 2005 calendar year was used for a model confirmation run. The 2005 calendar year had several high flow events in January and April, but had a dry summer (Figure 18). A short period in early August had streamflows above long-term average values, but otherwise had flows that were consistently below average from mid-April until the end of the year. The Northeast Cape Fear did have one high flow event in October 2004 that was not seen in the larger watershed (Figure 18). This high flow event was the result of heavy rains from tropical storm Jeanne that hit the coastal area but not farther upstream in the watershed.

A significant challenge of the confirmation run was collecting the necessary data to run the model. For the hydrodynamic model, the water surface elevation data was not a problem as both the Sunset Beach and Southport elevation data were sufficient for creating the necessary model forcings. The flow time histories at the upstream boundary on the Cape Fear River used the same location (Lock and Dam 1) as for the calibration run. For the Black and Northeast Cape Fear Rivers, however, data from the two USGS gaging stations nearest the upstream boundary (Black at Currie, Northeast Cape Fear at Burgaw) were no longer available. Instead data from the stations farther upstream (Black at Tomahawk, Northeast Cape Fear at Chinqapin, see Table 3 and Figure 7 for location information) were used instead. The USGS continuous station at Marker 12 was also not available for establishing the downstream boundary condition during 2005. Instead, the monthly salinities at one of the Lower Cape Fear River Stations (M18) was used instead. There was also much less data that could be used for calibration. For instance, over $5000 \mathrm{model} /$ data comparisons were made when calibrating salinity, but for the 2005 confirmation run, only 96 monitored salinities were available. Because of the small amount of monitoring data available only two stations (Navassa, Marker 23) are shown here as time history plots.

Time histories of model predicted salinities and observed salinities during 2005 demonstrate that the model can simulate estuarine dynamics in the system during periods other than the calibration period. Model predicted salinities at two stations in the impaired region (Navassa and Marker 61) and at two other stations in the Northeast Cape Fear and the lower estuary (NCF6 and M35) generally follow the patterns seen in the observed data (Figures 65 and 66). In general the model predicts a somewhat higher salinity than what is observed, but the magnitude of the overprediction is not severe. In addition the model does seem capable of modeling the overall seasonal trends in salinity, and does seem to correctly freshen the estuary during high flow events. For instance early and late 2005 had relatively high flows, and the model correctly simulates the freshening of the estuary that occurs during these times (Figure 65 and Figure 66).


Figure 65. Model Predicted (lines) and Observed (symbols) Salinities During the 2005 Confirmation Run in the NE Cape Fear River at Station NCF6 and in the Cape Fear River at Navassa.


Figure 66. Model Predicted (lines) and Observed (symbols) Salinities During the 2005 Confirmation Run at the Cape Fear River at Marker 61 and at Station M35.

Compared to salinity observations, significantly more observed data were available for assessing the performance of the temperature submodel during the confirmation run. The input data files needed to run the temperature submodel (e.g. meteorological data) were also available for the full 2005 confirmation run. Here we show time history comparisons of model predicted vs. observed temperatures at four sites within the estuary; two sites near Wilmington (Navassa, NCF117), and two sites downstream of Wilmington (Marker 61) and near the mouth of the estuary (Marker 35). Overall, all sixteen of the Lower Cape Fear River Monitoring Program sites used during calibration (Table 12, Figure 20) also had temperature data that could be used for performing a model confirmation run.

At all of the sites examined, the model did a good job predicting temperature time histories. At the two sites near Wilmington, the temperature reaches its maximum value near $30^{\circ}$ C during August and drops to approximately $10^{\circ} \mathrm{C}$ by December 1 (Figure 67). A similar seasonal pattern is seen at the other sites (Figure 68), although it seems that the peak temperature arrives just a little later (mid-August) at these two sites. The model follows the observed seasonal trend very well. At the most downstream site shown here (M35, Figure 68), the peak temperature arrives even a little later than at the other sites. Just as at the other three sites, the peak temperature is approximately $30^{\circ} \mathrm{C}$, and occurs sometime during August. At this site temperatures fall slightly less rapidly than at the other three sites. By the end of the year, model predicted temperatures were just above $10^{\circ} \mathrm{C}$ (Figure 68).

The same procedure used to assess model fit to an observed data set is used for the confirmation run. A total of 780 model/data comparisons were made for the 2005 confirmation run. Predicted temperatures had a mean value that was $0.52^{\circ} \mathrm{C}$ below that observed. The normalized mean error was only $-2.3 \%$, which attests to the model's overall ability to predict temperatures (Table 20). Likewise other measures of calibration performance (e.g. root mean square error, correlation $\mathrm{R}^{2}$, Table 20) are all similar to corresponding values for the calibration run (Table 16). Likewise the scatter plot (Figure 69) and the percentile plot (Figure 70) for the confirmation run look very similar to that for the calibration run (Figures 41 and 42).


Figure 67. Model Predicted (lines) and Observed (symbols) Temperatures During the 2005 Confirmation Run for the NE Cape Fear River at Station NCF117 and the Cape Fear River at Navasssa.


Figure 68. Model Predicted (lines) and Observed (symbols) Temperatures During the 2005 Confirmation Run for the Cape Fear River at Marker 61 and at Station M35.

Table 20. Model Fit Statistics, Temperature, 2005 Confirmation Run

| Parameter | Value | Units |
| :--- | :---: | :---: |
| Mean Error (predicted - <br> observed) | $-\mathbf{0 . 5 2}$ | ${ }^{\mathbf{\circ}} \mathrm{C}$ |
| Normalized Mean Error | $\mathbf{- 2 . 3}$ | $\mathbf{\%}$ |
| Root Mean Square Error | $\mathbf{1 . 5 1}$ | ${ }^{\mathbf{\circ}} \mathbf{C}$ |
| Normalized Root Mean <br> Square Error | $\mathbf{6 . 7}$ | $\mathbf{\%}$ |
| Mean Absolute Error | $\mathbf{1 . 0 8}$ | ${ }^{\mathbf{o}} \mathbf{C}$ |
| Normalized Mean Absolute <br> Error | $\mathbf{4 . 8}$ | $\mathbf{\%}$ |
| Correlation R ${ }^{2}$ | $\mathbf{9 6 . 4}$ | $\mathbf{\%}$ |
| Number of Model/Data <br> Comparisons | $\mathbf{7 8 0}$ | - |
| $\mathbf{R}^{2}$ Corrected for Bias | $\mathbf{9 6 . 0}$ | $\mathbf{\%}$ |



Figure 69. Scatter Plot of Predicted Temperature ( ${ }^{\circ} \mathrm{C}$, x-axis) vs. Corresponding Observed Temperature $\left({ }^{\circ} \mathrm{C}, \mathrm{y}\right.$-axis) during the 2005 Confirmation Run.


Figure 70. Percentile Plot of Observed and Model Predicted Temperatures During the 2005 Confirmation Period. The y-axis indicates the fraction of values below the corresponding temperature $\left({ }^{\circ} \mathrm{C}\right)$ indicated on the x -axis.

The model also did an excellent job of simulating DO concentrations during the confirmation run. The two upstream stations examined using time history plots (NC11, B210) had minimum dissolved oxygen concentrations that occurred near early September. The model simulates well the timing and magnitude of the minimum DO concentration (Figure 71). Likewise the recovery that occurs in the subsequent months as the water cools is simulated well by the model. At both upstream sites the model predicts a rise to a DO concentration of about 10 $\mathrm{mg} / \mathrm{L}$ by December 2005. Observed DO concentrations at the upstream sites in late 2005 closely match the model predictions (Figure 71).


Figure 71. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During the 2005 Confirmation Run for the Cape Fear River at Station NC11 and the Black River at Station B210.

Like the calibration run, the model overpredicts somewhat the DO concentrations in the Northeast Cape Fear during the early summer months. The model does a good job, however, at predicting the minimum dissolved oxygen concentration of approximately $3 \mathrm{mg} / \mathrm{L}$ (Figure 72). The duration of low dissolved oxygen ( $<4 \mathrm{mg} / \mathrm{L}$ ) at this site is somewhat underpredicted by the model. The model does a better job predicting DO concentration at this site during late Fall, when the water is cooler and DO concentrations are higher. The model does a better job predicting DO concentrations at Navassa, as both the observed and model predicted minimum DO concentrations are below $4 \mathrm{mg} / \mathrm{L}$ (Figure 72). The model does an equally good job matching the observed data at the Brunswick River station and at station M35 (Figure 73). At station M35, the model nicely predicts the timing of the DO minimum concentration, but overpredicts slightly the DO concentration at this time. Both the descent to the minimum DO concentration and the recovery later in the Fall are nicely represented by the model (Figure 73).

Model performance statistics for dissolved oxygen are very similar to those for the calibration run. Again, a total of 780 monitored DO concentrations were compared to the corresponding model prediction at that time and place (Table 21). The normalized mean error was slightly higher as compared with the calibration run ( $-2.0 \%$ vs. $0.06 \%$, see Table 18), but the normalized root mean square error and normalized mean absolute error are both in the range of $10-15 \%$, which is quite good for an estuary DO model.

The excellent fit to observed DO data during the 2005 model confirmation run can also been seen in the scatter and percentile plots. The scatter plot has a fairly tight distribution (Figure 74) and little bias is evident. Likewise the percentile plot for the confirmation run (Figure 75) looks very similar to the corresponding plot for the calibration run (Figure 52).


Figure 72. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During the 2005 Confirmation Run for the NE Cape Fear River at Station NCF117 and the Cape Fear River at Navasssa.


Figure 73. Model Predicted (lines) and Observed (symbols) Dissolved Oxygen Concentrations During the 2005 Confirmation Run for the Cape Fear River at the Brunswick River Station and at Station M35.

Table 21. Model Fit Statistics, Dissolved Oxygen, 2005 Confirmation Run

| Parameter | Value | Units |
| :--- | :---: | :---: |
| Mean Error (predicted - <br> observed) | $\mathbf{- 0 . 1 3}$ | $\mathbf{m g} / \mathrm{L}$ |
| Normalized Mean Error | $\mathbf{- 2 . 0}$ | $\mathbf{\%}$ |
| Root Mean Square Error | $\mathbf{1 . 2 0}$ | $\mathbf{m g} / \mathrm{L}$ |
| Normalized Root Mean <br> Square Error | $\mathbf{1 8 . 2}$ | $\mathbf{\%}$ |
| Mean Absolute Error | $\mathbf{0 . 9 2}$ | $\mathbf{m g} / \mathrm{L}$ |
| Normalized Mean Absolute <br> Error | $\mathbf{1 4 . 1}$ | $\mathbf{\%}$ |
| Correlation R 2 | $\mathbf{\%}$ |  |
| Number of Model/Data <br> Comparisons | $\mathbf{7 3 . 8}$ | - |
| $\mathbf{R}^{2}$ Corrected for Bias | $\mathbf{7 3 . 0}$ | $\mathbf{\%}$ |



Figure 74. Scatter Plot of Predicted Dissolved Oxygen Concentrations (mg/L, x-axis) vs.
Corresponding Observed Dissolved Oxygen Concentrations ( $\mathrm{mg} / \mathrm{L}, \mathrm{y}$-axis) for the 2005 Confirmation Period.


Figure 75. Percentile Plot of Observed and Model Predicted Dissolved Oxygen Concentrations During the 2005 Confirmation Period. The y-axis indicates the fraction of values below the corresponding DO concentration $(\mathrm{mg} / \mathrm{L})$ indicated on the x -axis.

### 5.2 Sensitivity Testing

Model evaluation also included sensitivity testing. During calibration it was realized that the DO results were quite sensitive to the choice of the sediment oxygen demand. The SOD rate was determined during calibration to give a percentile plot that matched the observed data while giving good values for the numerical measures of calibration performance. Here we show how changing the SOD rate at $20^{\circ} \mathrm{C}$ from the calibrated value of $0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ to a value $25 \%$ higher $\left(0.5 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right)$ and a value $25 \%$ lower $\left(0.3 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right)$. While we examined the model results several different ways (e.g. time history plots, statistical measures of model fit), here we show only the percentile plot and the scatter plot for the high SOD and low SOD cases. These figures can be
compared to the corresponding scatter (Figure 51) and percentile plots (Figure 52) for DO using the calibrated SOD value of $0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$.

The most noticeable difference between the three runs is in the percentage of low DO concentrations. The scatter plots show more subtle differences. The three scatter plots for low, medium, and high SOD rate (Figures 51, 76, 77) are very similar in overall shape. The higher SOD seems to simply raise predicted DO concentrations without affecting the overall distribution, such that the predicted DO's are moved to the left on the figure (lower predicted DO) without changing the shape or tightness of the distribution. The percentile plots are also very similar, but there is a marked increase in the number of low DO value as the SOD rate is increased (Figures 52, 76, 77). For instance, with the SOD of $0.3 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ the percentage of values below $4 \mathrm{mg} / \mathrm{L}$ is approximately $7 \%$, while at an SOD rate of $0.5 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ the percentage of DO concentrations below is approximately $12 \%$. For this higher SOD rate, there is noticeable difference between the observed and model predicted distribution of data. This offset, where the model predicted frequency exceeds the observed data by approximately $5 \%$ exists in the lower range of the distribution, up to a percentile of approximately $65 \%$.

The sensitivity testing did show that the model results are sensitive to the calibrated SOD value. Additional runs were made to determine how changing the SOD rate might change the estuary's sensitivity to changes in organic matter loading. This set of runs, which examined if scenario results were sensitive to the SOD rate used is presented in the following section with the other scenario tests.


Figure 76. Scatter Plot and Percentile Plot, DO, $\mathrm{SOD}=0.3$


Figure 77. Scatter Plot and Percentile Plot, DO, $\mathrm{SOD}=0.5$

## 6. SCENARIO TESTING

### 6.1 Description of Base and Scenario Cases

A number of scenario tests were conducted to investigate the system's sensitivity to loading changes or to other possible changes in the system. For each scenario, model runs were conducted to estimate the time varying dissolved oxygen concentrations during the summer period of 2004 (April 1 - October 31) within the state designated impaired region of the estuary. For the base case run and each of the various scenario cases, twice-daily dissolved oxygen snapshots were recorded from all eight vertical layers at eighteen sites within the impaired region (Figure 78). This data collection method gave a total of $61,920 \mathrm{DO}$ concentration values for a particular model run ( 215 days x 18 stations x 8 layers per station x 2 snapshots per day). The distribution of these DO concentration values was then compared between the base case and the various scenario cases by making percentile plots. In the following sections, the following eight different scenarios are examined:

1. Eliminating wastewater point source loadings
2. Reducing river, creek, and wetland loadings
3. Changing wastewater loadings for various values of sediment oxygen demand
4. Reducing river, creek, and wetland loadings and sediment oxygen demand
5. Eliminating ammonia inputs from wastewater point sources
6. Increasing wastewater inputs to maximum permitted values
7. Deepening of the navigation channel
8. Changing Brunswick County wastewater loadings

The final section provides an analysis of the relative magnitudes of the processes that produce a dissolved oxygen deficit in the impaired region of the Lower Cape Fear River Estuary during the summer months.

### 6.2 Effect of Eliminating Wastewater Point Source Loadings

Three model runs were made to investigate the effect of completely turning off the discharges from the twenty wastewater treatment plants that discharge directly to the Lower Cape Fear River Estuary (see Table 9 for a list of wastewater point sources). These scenarios were run, not because completely eliminating pollutant discharges was considered achievable or advisable, but instead to investigate the system's sensitivity to WWTP loads. For this analysis, the river loads arriving at the model boundary, plus the creek and wetland inputs within the estuary were maintained at the base condition levels. To model the elimination of wastewater loads it was assumed that organic matter (refractory particulate carbon, nitrogen, and phosphorus; dissolved carbon, nitrogen, and phosphorus) and ammonia concentrations in the wastewater inputs would be decreased to $0.0 \mathrm{mg} / \mathrm{L}$.


Figure 78. Eighteen Locations Used to Determine Model Predicted Dissolved Oxygen Concentrations for the Scenario Tests

When all WWTP loads were eliminated, the $10^{\text {th }}$ percentile DO concentration increased by approximately $0.3 \mathrm{mg} / \mathrm{L}$, from 4.3 to $4.6 \mathrm{mg} / \mathrm{L}$ (Figure 79). In general the effect of eliminating the wastewater input had a larger impact on the lower dissolved oxygen concentrations than it did on the higher DO concentrations. Whereas the $10^{\text {th }}$ percentile DO
concentration increased by $0.3 \mathrm{mg} / \mathrm{L}$, the median value DO concentration increased by about $0.10 \mathrm{mg} / \mathrm{L}$, from 5.6 to $5.7 \mathrm{mg} / \mathrm{L}$ when all point source discharges were removed. The distribution of dissolved oxygen concentrations above $6.0 \mathrm{mg} / \mathrm{L}$ was only slightly changed ( $<0.1$ $\mathrm{mg} / \mathrm{L}$ ) by eliminating wastewater inputs of organic matter and ammonia (Figure 79). Selectively turning off individual WWTP loads, as expected, had a smaller effect. Turning off the International Paper WWTP had an effect that was roughly $2 / 3$ of that when all the plants were turned off. Turning off both Wilmington domestic wastewater treatment plants had an effect that was roughly $1 / 3$ of that when all discharges were turned off (Figure 79). It was also found that eliminating the wastewater point sources results in only a small decrease in the percentage of DO concentrations below $5.0 \mathrm{mg} / \mathrm{L}$. For the base case, with discharges from all twenty wastewater


Figure 79. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and Three Wastewater Treatment Plant Load Reduction Scenarios. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.
treatment plants, approximately $32 \%$ of the estimated DO concentrations in the impaired region during the summer months are below $5.0 \mathrm{mg} / \mathrm{L}$. When the oxygen demanding wastes from these point sources are completely eliminated, approximately $27 \%$ of summertime DO concentrations are below $5.0 \mathrm{mg} / \mathrm{L}$ (Figure 79).

### 6.3 Effect of Reducing River, Creek, and Wetland Loadings

The impact of changing the loadings that entered the model region from rivers, creeks, and wetlands was also investigated. Loading reductions of either $30 \%, 50 \%$, or $70 \%$ were assumed for the three river inputs (Cape Fear, Black, and Northeast Cape Fear), and from the fourteen creek and wetland inputs into the estuary. To reduce the loadings, flows were maintained at the base case levels, but concentrations were reduced by $30 \%, 50 \%$, or $70 \%$ from the base case values. Identical reduction percentages were assumed for the organic matter (refractory particulate carbon, nitrogen, and phosphorus; dissolved carbon, nitrogen, and phosphorus) and ammonia constituents. Loadings for the twenty wastewater point sources were maintained at the levels in the base case scenario. As with all scenarios, twice-daily DO concentration snapshots were analyzed for all eight model layers, at eighteen sites within the impaired region for 215 days during 2004 (April 1, 2004 - October 31, 2004).

The load reductions of riverine, creek, and wetland inputs were found to have a significant impact on the estimated dissolved oxygen concentrations during the summer months in the impaired region of the Lower Cape Fear River Estuary. At the $10^{\text {th }}$ percentile level, DO concentrations for the three load reduction scenarios increased by $0.2,0.3$, and $0.4 \mathrm{mg} / \mathrm{L}$ respectively, from $4.3 \mathrm{mg} / \mathrm{l}$ to either $4.5,4.6$, or $4.7 \mathrm{mg} / \mathrm{L}$ (Figure 80 ). Unlike the scenarios described in the previous section in which wastewater loading decreases were investigated, the level of increase in DO concentration was maintained at the higher percentiles when reductions in the river, creek, and wetland loadings were made. In fact, for this "clean river" scenario, the median DO concentration increased to even a greater extent than the $10^{\text {th }}$ percentile value, increasing from 5.6 to $5.85 \mathrm{mg} / \mathrm{L}$ for the $30 \%$ reduction (an increase of $0.25 \mathrm{mg} / \mathrm{L}$ ), and from 5.6 to $6.2 \mathrm{mg} / \mathrm{L}$ (an increase of $0.6 \mathrm{mg} / \mathrm{L}$ ) for the $70 \%$ reduction scenario (Figure 80 ). Nonetheless, even at this largest assumed loading reduction amount of $70 \%$, a significant fraction (approximately $19 \%$ ) of DO concentrations are below $5.0 \mathrm{mg} / \mathrm{L}$. For the base case, approximately $32 \%$ of DO concentrations during the summer months are below $5.0 \mathrm{mg} / \mathrm{L}$ (Figure 80).


Figure 80. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and two Scenarios that Reduce Loading from Rivers and Creeks by $30 \%, 50 \%$, and $70 \%$. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.4 Effect of Changing Sediment Oxygen Demand

Additional water quality model runs were conducted to investigate the sensitivity of the results to the choice of model kinetic parameters. Because of its relative importance in determining dissolved oxygen concentrations, the sensitivity analysis focused on the impact of various choices for the sediment oxygen demand. During calibration it was found that this parameter had a significant impact on dissolved oxygen concentrations in the estuary. In addition to the calibrated SOD value $\left(0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right.$ at $\left.20^{\circ} \mathrm{C}\right)$, runs were made at a lower value ( 0.3 $\mathrm{g} / \mathrm{m}^{2} / \mathrm{d}$ at $20^{\circ} \mathrm{C}$ ) and a higher value ( $0.5 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ at $20^{\circ} \mathrm{C}$ ). Model fit to the observed data set was compared for these three cases. All three cases (SOD @ $20 \operatorname{deg} \mathrm{C}=0.3,0.4,0.5$ ) had similar correlation $r^{2}$ values ( $83.4 \%$, and $84.5 \%$, and $83.4 \%$ ), but on average the low SOD case overpredicted DO concentrations by $0.11 \mathrm{mg} / \mathrm{L}$, while the high SOD case underpredicted DO concentrations by a similar amount $(0.09 \mathrm{mg} / \mathrm{L})$.

The impact of turning off the organic matter and ammonia loadings from all wastewater treatment plants was investigated for each of these three SOD cases. Each scenario produced a very similar change in DO concentration when all point sources were turned off (Figure 81). Even though DO concentrations were changed with the different SOD cases, in each scenario, turning off the sources increased DO concentrations by approximately $0.25 \mathrm{mg} / \mathrm{L}$ at the $10^{\text {th }}$ percentile level (Figure 81). Likewise the percentage of DO values below $5.0 \mathrm{mg} / \mathrm{L}$ decreased by a similar amount using the three different sediment oxygen demand values. For the base case SOD value $\left(0.4 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right.$ at $\left.20^{\circ} \mathrm{C}\right)$ turning off the wastewater sources decreased the percentage of DO values below $5.0 \mathrm{mg} / \mathrm{L}$ from $45 \%$ to $40 \%$. For the lower SOD value ( $0.3 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}$ at $20^{\circ} \mathrm{C}$ ), removing the wastewater sources decreased the percentage of values below $5.0 \mathrm{mg} / \mathrm{L}$ from $37 \%$ to $29 \%$. Using the higher SOD value $\left(0.5 \mathrm{~g} / \mathrm{m}^{2} / \mathrm{d}\right.$ at $\left.20^{\circ} \mathrm{C}\right)$ removing the wastewater sources decreased the percentage of values below $5.0 \mathrm{mg} / \mathrm{L}$ from $54 \%$ to $49 \%$ (Figure 81). The magnitude of change in the percentage of values below $5.0 \mathrm{mg} / \mathrm{L}$ changes only slightly for these three SOD values. It is expected that changes in other model parameters would have a qualitatively similar result. Although absolute concentrations are affected by changes in parameter values, relative changes in concentration are less affected by changing parameter values.


Figure 81. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and a No WWTP Loading Scenario for Three Different Values of Sediment Oxygen Demand. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.5 Cumulative Effect of Lower Sediment Oxygen Demand and Reduced Loadings from Rivers, Creeks, and Wetlands

It has been shown (section 6.3) that even a $70 \%$ loading reduction in the organic matter inputs from rivers, creeks, and wetlands sources is not enough to completely eliminate violations of the state water quality standard of $5.0 \mathrm{mg} / \mathrm{L}$ (see Figure 80), even though these sources constitute the majority of inputs of oxygen demanding wastes into the Lower Cape Fear Estuary (see Figures 42 and 43). In this scenario we examine what conditions would be necessary to produce summertime DO concentrations above $5.0 \mathrm{mg} / \mathrm{L}$. In addition, one limitation of the analysis done previously is that it ignores possible changes that might occur in the benthos if organic matter loadings were reduced. For instance, it is likely that a reduction of $30 \%$ or $50 \%$ or $70 \%$ in organic matter loading would in the long-term also result in lower sediment oxygen demands. The cumulative effect of decreasing both organic matter loading and sediment oxygen demand are examined in this scenario.

In this scenario a total of six additional runs were made and compared to the base case. As done previously, three runs assumed that the oxygen demanding wastes (organic matter and ammonia) from the three rivers (Cape Fear, Black, Northeast Cape Fear) and from the fourteen creek and wetland inputs to the estuary were reduced by $30 \%, 50 \%$, or $70 \%$. These reductions were made by holding steady the freshwater inflows while reducing the concentrations by the appropriate amount. For these three runs, wastewater inputs and the sediment oxygen demand were held at the values used in the base case. Three additional runs were made that assumed $30 \%, 50 \%$, or $70 \%$ reductions both in the oxygen demanding waste input and the sediment oxygen demands. These reductions in sediment oxygen demand were made by assuming a constant percentage reduction in sediment oxygen demand while retaining the assumed time variations in these values. As described in the calibration section of the report, sediment oxygen demand was assumed to vary seasonally, with higher values in the summer when the benthic layer is warmer and the biota more active as compared with the winter months.

As expected reducing the sediment oxygen demand along with the waste inputs had a larger effect than when only the waste inputs were reduced. For the case with a $30 \%$ reduction in SOD and loadings from rivers, creeks, and wetlands, the percentage of summertime DO concentrations below $5.0 \mathrm{mg} / \mathrm{L}$ was reduced to $22 \%$, as compared to $45 \%$ for the base case and $33 \%$ when only the oxygen demanding wastes are decreased (Figure 82). The $50 \%$ reduction case had an even lower rate of water quality violations, but these were not completely eliminated. With both SOD and oxygen demanding wastes decreased, approximately $7 \%$ of summertime DO concentrations in the impaired region are below $5.0 \mathrm{mg} / \mathrm{L}$, as compared to $27 \%$ when only the oxygen demanding wastes are decreased (Figure 82). There is also a large increase in the minimum predicted DO concentration for this case. The base case had a minimum predicted DO concentration of approximately $3.2 \mathrm{mg} / \mathrm{L}$, whereas the minimum when SOD and oxygen demanding wastes are reduced by $50 \%$ is approximately $4.6 \mathrm{mg} / \mathrm{L}$ (Figure 82). A decrease in SOD of $70 \%$ and a reduction in river load of $70 \%$, however, does almost completely eliminate dissolved oxygen concentrations below $5.0 \mathrm{mg} / \mathrm{L}$ (Figure 82). For this case, only about $1 \%$ of the predicted dissolved oxygen concentrations are below the water quality standard value.

## April through October Simulated Dissolved Oxygen Concentrations in the Impaired Area, Lower Cape Fear River Estuary: Clean Rivers Scenarios



Figure 82. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and two Scenarios that Reduce Loading from Rivers and Creeks by $30 \%$ and $50 \%$. The dashed lines show the corresponding changes in dissolved oxygen when SOD is also reduced by $30 \%$ or $50 \%$. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.6 Effect of Eliminating Ammonia Inputs from Wastewater Point Sources

A fifth scenario considered eliminating the ammonia loading for all wastewater treatment plants. As for the other scenarios, loading reductions were achieved by decreasing the effluent concentrations, while keep the discharges at that specified in the base case. For this scenario the ammonia concentration in all twenty wastewater point sources was set to zero. The wastewater loadings from all other constituents was held constant, as was the loading from the seventeen other point sources that provided contributions from the rivers, creeks, and wetlands that discharge directly to the estuary.

This scenario produced an increase in dissolved oxygen concentrations that was detectable but not dramatic. As expected, the magnitude of the increase was less than when both the organic matter and ammonia loadings were eliminated from wastewater sources. At the $10^{\text {th }}$ percentile level, dissolved oxygen concentrations increased by about $0.1 \mathrm{mg} / \mathrm{L}$, from about 4.3 $\mathrm{mg} / \mathrm{L}$ for the base case to about $4.4 \mathrm{mg} / \mathrm{L}$ when the ammonia loading was eliminated (Figure 83). The magnitude of the increase generally decreased for higher percentile values. For instance, the median dissolved oxygen concentration in the base case ( $5.6 \mathrm{mg} / \mathrm{L}$ ) increased by only about 0.05 $\mathrm{mg} / \mathrm{L}$, with concentrations in the upper part of the distribution ( $>6.0 \mathrm{mg} / \mathrm{L}$ ) increasing by an even lesser amount. A small, but detectable drop in percentage of DO concentrations less than $5.0 \mathrm{mg} / \mathrm{L}$ was seen when ammonia discharges from wastewater treatment plants was eliminated. In the base case approximately $32 \%$ of summertime DO concentrations in the impaired region are below the water quality standard of $5.0 \mathrm{mg} / \mathrm{L}$, whereas approximately $30 \%$ are below 5.0 $\mathrm{mg} / \mathrm{L}$ when ammonia discharges are eliminated (Figure 83).

April through October Simulated Dissolved Oxygen Concentrations in the Impaired Area, Lower Cape Fear River Estuary


Figure 83. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and a no Ammonia Loading Scenario. The y-axis indicates the fraction of values below the corresponding DO concentration (mg/L) indicated on the x -axis.

### 6.7 Effect of Increasing Wastewater Inputs to Maximum Permitted Values

The base case scenario used facility discharge monitoring records to establish the loadings of oxygen demanding wastes from each of the twenty wastewater point sources that discharge directly to the Lower Cape Fear Estuary. In several instances, the actual loadings were considerably below permitted maximum values. Of interest is how increasing all 20 wastewater inputs to the permitted maximum values would affect the dissolved oxygen conditions in the estuary. In this scenario, those permitted maximum discharges are assumed for each of the sources where that information is available.

Of the twenty wastewater point sources in the model, thirteen had maximum permitted discharges for either BOD, flow, or ammonia that exceeded the loadings in the base case (Table 23). For each of these sources, modifications were made to the EFDC water quality point source input file (WQPSL.INP) to simulate a continuous discharge at the maximum permitted value. The permit limits for these thirteen sources (Table 24) were provided by the NC Division of Water Quality (DWQ). In keeping with the DWQ's rules for these permits, winter limits were used for five months (January, February, March, November, December) and summer limits for the remaining seven months. In creating the scenario all other sources and parameters were held at the base case levels.

Table 23. Model Point Sources That Were Adjusted to Maximum Permitted Values

| Source <br> No. | NPDES <br> Permit No. | Name of Source |
| :---: | :---: | :--- |

The scenario with the thirteen wastewater sources at their permitted maximum values produced dissolved oxygen concentrations that were slightly lower than the corresponding base values by approximately $0.1 \mathrm{mg} / \mathrm{L}$ (Figure 84). This decrease of approximately $0.1 \mathrm{mg} / \mathrm{L}$ was seen throughout the distributions, with both the $10^{\text {th }}$ percentile DO concentration value and the
median value decrease by a similar amount. More low DO concentrations were predicted for the maximum permit case as opposed to the base case. The percentage of DO concentration below the water quality permit value of $5.0 \mathrm{mg} / \mathrm{L}$ increased from $32 \%$ to approximately $34 \%$ for the maximum permit case (Figure 84). Approximately $4 \%$ of the base case DO concentrations in the impaired region during the Summer were expected to be below $4.0 \mathrm{mg} / \mathrm{L}$, while the maximum permit case had approximately $7 \%$ below $4.0 \mathrm{mg} / \mathrm{L}$ (Figure 84).

Table 24. Total Maximum Monthly Permitted Values for the Point Sources in the Lower Cape Fear Estuary.

| Permit \# | Constituents | Outfall 1 |  | Outfall 2 |  | Outfall 3 |  | Outfall 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Monthly | Unit | Monthly | Unit | Monthly | Unit | Monthly | Unit |
| NC0000663 | Flow | 3.5 | MGD |  |  |  |  |  |  |
| 1 | BOD, 5-Day (20 Deg. C) | 493.3 | Ibs/day |  |  |  |  |  |  |
|  | Solids, Total Suspended | 786 | lb/day | 75 | mg/L |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| NC0001112 | Flow |  |  | 1.25 | MGD |  |  |  |  |
| 5 (2) | BOD, 5-Day (20 Deg. C) Summer |  |  | 215.7 | lb/day |  |  |  |  |
|  | Solids, Total Suspended | 30 | mg/L | 329 | lb/day |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| NC0003298 | Flow | 50 | MGD |  |  |  |  |  |  |
| 6 | BOD, 5-Day (20 Deg. C) Winter | 10,000 | Ibs/day |  |  |  |  |  |  |
|  | BOD, 5-Day (20 Deg. C) Summer | 5,000 | Ibs/day |  |  |  |  |  |  |
|  | Solids, Total Suspended | 65,720 | lbs/day | 30 | mg/L |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| NC0021334 | Flow | 0.8 | MGD |  |  |  |  |  |  |
| 7 | BOD, 5-Day (20 Deg. C) | 30 | mg/L |  |  |  |  |  |  |
|  | Solids, Total Suspended | 30 | mg/L |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| NC0023256 | Flow | 3 | MGD |  |  |  |  |  |  |
| 8 | BOD, 5-Day (20 Deg. C) Winter | 10 | mg/L |  |  |  |  |  |  |
|  | BOD, 5-Day (20 Deg. C) Summer | 5 | mg/L |  |  |  |  |  |  |
|  | Nitrogen, Ammonia Total (as N) winter | 4 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  | Nitrogen, Ammonia Total (as N) summer | 2 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| NC0023965 | Flow | 16 | MGD |  |  |  |  |  |  |
| 9 | BOD, 5-Day (20 Deg. C) | 30 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  | Solids, Total Suspended | 30 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  | Nitrogen, Ammonia Total (as N) winter | 2 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  | Nitrogen, Ammonia Total (as N) summer | 1 | $\mathrm{mg} / \mathrm{L}$ |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |



## April through October Simulated Dissolved Oxygen Concentrations in the Impaired Area, Lower Cape Fear River Estuary: Maximum Permitted Discharge Scenario



Figure 84. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and a Scenario Where all WWTP's are Set to the Their Maximum Permitted Discharge. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.8 Effect of Deepening of the Navigation Channel

An ongoing navigation channel deepening project is underway in the Lower Cape Fear River Estuary. This project is being conducted by the Wilmington district of the US Army Corps of Engineers. The Wilmington Harbor Project will deepen 37 miles of channel by 4 feet in the Cape Fear River up to the State Port of Wilmington, and into the Northeast Cape Fear River. Project information is available www.saw.usace.army.mil/wilmington-harbor/main.htm. Information from this site was used to modify the water depths given in the EFDC grid file DXDY.INP. Each grid cell in the navigation channel was deepened according to the plan described by the Army Corps of Engineers. No other changes were made from the base case values in defining the various input files needed to run the model.

The channel deepening had a noticeable, and slightly negative effect on dissolved oxygen concentrations during the summertime in the Lower Cape Fear River Estuary. Deepening of the channel slightly decreased dissolved oxygen concentrations. The reduction in dissolved oxygen concentrations was similar across the distribution of dissolved oxygen values. At the $10^{\text {th }}$ percentile value, the dissolved oxygen was decreased by about $0.2 \mathrm{mg} / \mathrm{L}$, from $4.3 \mathrm{mg} / \mathrm{L}$ to 4.1 $\mathrm{mg} / \mathrm{L}$. Near the median dissolved oxygen concentration, the dissolved oxygen decreased by about $0.1 \mathrm{mg} / \mathrm{L}$, from $5.6 \mathrm{mg} / \mathrm{L}$ to $5.5 \mathrm{mg} / \mathrm{L}$ (Figure 85). The effect of deepening the channel is also seen when comparing the frequency of DO concentrations below $5.0 \mathrm{mg} / \mathrm{L}$. For the base case the percentage below $5.0 \mathrm{mg} / \mathrm{L}$ was $32 \%$, for the deepened channel the percentage was approximately $34 \%$. Thus there is an effect with the channel deepening, which slightly lowers dissolved oxygen concentrations, but the magnitude of the effect is small ( $<0.2 \mathrm{mg} / \mathrm{L}$ difference, $>5 \%$ difference in DO concentrations below a particular value).


Figure 85. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and a Scenario Where the Navigation Channel is Deepened by Dredging. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.9 Effect of Changing Brunswick County Wastewater Loadings

Three model scenarios considered the impact of changing the wastewater flow of a single discharger (Brunswick County wastewater treatment plant). Based on the discharge monitoring reports for the model period (2003-2005), the base condition had a time-averaged flow of 0.38 MGD. The three alternate scenarios tested had time-averaged flows of 1.65, 4.65 and 15 MGD. These scenarios therefore represented the effect of changing the wastewater treatment flow by factors of 4.3, 12.1, and 39.1. In these alternate loading scenarios, each discharge flow was increased by a constant factor. All concentrations were assumed to be constant. Thus the wastewater loadings had identical patterns of time variation, but had time-averaged flows that equaled the values given above. As for the other scenarios, twice daily dissolved oxygen concentrations were collected and compared to the base case at eighteen sites within the impaired region, from all eight layers of the model and from every day within the summertime period (April 1 - October 31, 2004).

Despite the relatively large changes in the assumed wastewater flow, there were only very small differences in the predicted dissolved oxygen concentrations, for summertime conditions in the impaired region. Across the entire range of dissolved oxygen concentrations, the differences between the base case and each of the three scenarios was always less than 0.05 $\mathrm{mg} / \mathrm{L}$ (Figure 86). Thus there was essentially no effect of changing the wastewater flow. One limitation of this analysis is the assumption that loading would increase solely by increasing the flow. Since the organic matter concentrations were only slightly higher than the receiving waters, it is not surprising that changing the loading by holding effluent concentrations constant while increasing the flows had a very small effect on dissolved oxygen concentrations.


Figure 86. Percentile Plot of Model Predicted Dissolved Oxygen Concentrations During the Summer 2004 for the Base Case and Three Brunswick County WWTP Loading Scenarios. The y-axis indicates the fraction of values below the corresponding DO concentration ( $\mathrm{mg} / \mathrm{L}$ ) indicated on the x -axis.

### 6.10 Analysis of the Summer Dissolved Oxygen Deficit in the Impaired Region

One conclusion that results from the previous scenarios is that depressed summertime dissolved oxygen concentrations in the Lower Cape Fear River Estuary can possibly be the result from one or more processes. Sediment oxygen demand, water-column decomposition of organic matter, or oxidation of ammonia can all result in dissolved oxygen depletion. Furthermore, there are several sources of organic matter in the water column (e.g. wastewater point sources, riverine inputs, inputs from creeks and rivers). An important consideration in the analysis of water quality in the Lower Cape Fear River Estuary is the relative importance of these various sources and processes to the dissolved oxygen concentrations in the impaired region of the Estuary.

To investigate the relative importance of these sources and processes, an analysis was performed to quantitatively compare the rates of oxygen consumption in the water-column due to riverine, estuarine, and wastewater sources, and to compare this consumption to sediment oxygen demand. The analysis focused on three sites within the impaired region where monitoring data were available: Cape Fear at Navassa (see Figure 6, Station NAV), Northeast Cape Fear River at Wilmington (see Figure 7, USGS Station 02108690), and Cape Fear River at Horseshoe Bend (see Figure 6, Station HB). Twice daily model predicted dissolved oxygen concentration, temperature, and salinity values were recorded near the surface (layer 7 of 8 ) at these three stations. Temperature and salinity values were used to calculate dissolved oxygen saturation concentration (Chapra 1997), and then the DO concentrations were plotted as a fraction of the DO saturation concentration (Figure 87) for a fourteen-month period beginning December 1, 2003. At all three stations, the DO saturation fraction is highest during the winter months and lowest during the summer months, probably as a result of temperature related differences in the consumption rate of water-column dissolved oxygen. At all three stations DO saturation percentages peak at about $90 \%$ in early March but then decline steady until August, when DO saturation percentages are less than $60 \%$ at all three stations. Beginning in early April, DO saturation values go below $70 \%$ at all stations, and do go back above $70 \%$ until late October (Figure 87).

Clearly oxygen consumption and under-saturated DO concentrations are a persistent feature during the summer at these three stations within the impaired region of the Lower Cape Fear Estuary, but what is the relative importance of the processes that produce this under saturation of dissolved oxygen in the water-column? To investigate this question, a first-order sensitivity analysis was performed to determine the relative importance of these three processes:

1. Discharge of oxygen demanding wastes by the 20 wastewater treatment plants that discharge directly to the estuary,
2. Discharge of oxygen demanding wastes by the three rivers in the basin (Cape Fear, Black, Northeast Cape Fear) as well as the creek and wetlands areas that discharge directly to the estuary, and
3. Sediment oxygen demand.


Figure 87. Model Predicted Dissolved Oxygen Concentrations in the Surface Waters as a Percentage of Saturation Concentration at Two Cape Fear Locations (Navassa and Horseshoe Bend) and one NE Cape Fear Station (Wilmington).

Three separate runs were made to determine the sensitivity of the time-averaged DO concentration at the three test sites to changes in one of the processes listed above. In each run, a prescribed relative change (e.g. $10 \%$ of the base case) in one of the loadings or the sediment oxygen demand rate was made, while holding all other rates and parameters at the values set in the base case. These three runs were then compared to the base case and the relatively sensitivity of each process was determined by comparing the change in dissolved oxygen
concentration between the cases. The time-averaged mean oxygen depletion at each site was then apportioned based upon the relative sensitivity of the three processes.

There were slight differences in the magnitudes of the oxygen depletion attributable to the three different factors at the three different sites (Figure 88). In general, however, the dissolved oxygen deficit caused by SOD and the river loadings of organic matter were significantly greater than that caused by the WWTP loadings. Time averaged dissolved oxygen concentrations were slightly different at the three sites, ranging from $5.65 \mathrm{mg} / \mathrm{L}$ at Navassa to $5.45 \mathrm{mg} / \mathrm{L}$ at the Northeast Cape Fear at Wilmington site. At each site the time-averaged dissolved oxygen saturation concentration was approximately $8.2 \mathrm{mg} / \mathrm{L}$, thus each site had a time-averaged dissolved oxygen depletion of approximately $2.7 \mathrm{mg} / \mathrm{L}$. Of the three processes compared, the wastewater treatment plant discharges had by far the smallest dissolved oxygen depletion effect. At each wastewater treatment plants accounted for approximately $0.2 \mathrm{mg} / \mathrm{L}$ of the total dissolved oxygen depletion. This amount of depletion represents approximately 7 to $9 \%$ of the total depletion. Of the remaining DO depletion (approximately $2.5 \mathrm{mg} / \mathrm{L}$, which is more than $90 \%$ of the total), roughly similar amounts could be attributable to SOD or riverine loadings of organic matter. The exception to this was the Northeast Cape Fear at Wilmington site, where more of the depletion ( $56 \%$ vs. $37 \%$ ) was due to the SOD. These findings are consistent with earlier estimates of organic matter loading that show that rivers, creeks, and wetlands are the dominant sources of oxygen demanding wastes to the estuary (e.g. see Figures 43, 44, and 45).


Figure 88. Bar Graphs of Summer Season (April through October) Time-Averaged Model Predicted Dissolved Oxygen Concentrations (green) and the Dissolved Oxygen Deficit Caused by Wastewater Treatment Plants (black) Riverine Discharges (blue), and Sediment Oxygen Demand (pink) at Two Cape Fear Locations (Navassa and Horseshoe Bend) and one NE Cape Fear Station (Wilmington).

## 7. SUMMARY, DISCUSSION, AND CONCLUSIONS

The three-dimensional water quality model EFDC (Environmental Fluid Dynamics Code) was applied to simulate hydrodynamic and water conditions for the Lower Cape Fear River Estuary, North Carolina. The model region included the tidally affected portions of the Cape Fear, Black, and Northeast Cape Fear Rivers near Wilmington, North Carolina, and extended southward to the mouth of the Cape Fear River near Southport, North Carolina. A multi-agency monitoring program was used to gather the necessary data to run and calibrate the model. A fourteen-month period from November 2003 until January 2005 was used for model calibration. The model was created in such a way so that it separately accounted for the oxygen demanding loadings to the estuary from the three major rivers, tidal creeks and fringing wetlands within the estuary, wastewater treatment plant discharges, and sediment oxygen demand within the estuary. A phased multi-objective calibration procedure was implemented that sought to maximize the model's fit to multiple water quality constituents. The calibration phasing was accomplished in such a way that recognized the primary importance of dissolved oxygen prediction. The calibrated model simulated well the temporal and spatial patterns of dissolved oxygen concentration within the estuary. Model confirmation testing was performed for a twelve-month period from January 2005 to January 2006. The model's ability to simulate dissolved oxygen during the confirmation run was similar to that for the calibration period.

The model was used to investigate the effects of various organic matter and ammonia load reduction scenarios on the dissolved oxygen concentrations within the estuary. Various scenarios investigated the system response to reductions in wastewater and watershed loadings. Of the three primary constituents that contribute to oxygen depletion, only ammonia had a load that was contributed primarily from the wastewater sources. For organic matter loadings, the majority of the load was contributed by the incoming rivers and local tidal creeks and wetlands discharging directly to the estuary. Reducing these watershed loads had a significant impact on the dissolved concentrations within the estuary. Despite the sensitivity of the system to changes in these loadings, elimination of violations of the water quality standard for dissolved oxygen (5 $\mathrm{mg} / \mathrm{L}$ ) would require a $70 \%$ reduction both in riverine loadings but also a similar reduction in sediment oxygen demand.

Reducing wastewater loads generally had a smaller, but still noticeable impact on dissolved oxygen concentrations. Generally the magnitude of the increase in dissolved oxygen concentration followed the relative importance of the discharge to the overall loading to the estuary. An additional scenario tested the degree to which parameter variation, in this case sediment oxygen demand, could change the system's sensitivity to load reductions. This case demonstrated that the change in DO concentrations for a particular load reduction did not vary appreciably as the SOD used by the model was changed. Other scenarios investigated the system's sensitivity, with regard to dissolved oxygen concentrations to changes in the estuary that might result from dredging of the navigation channel or increases in wastewater treatment plant loadings up to the maximum permitted values. These two scenarios had a relatively minor effect on dissolved oxygen concentrations.

The model was designed to allow for relatively simple specification of additional scenarios. The results of the scenario tests, strategic choices made in creating the model, and results from previous modeling work on the estuary (e.g. Tetra Tech 2001), indicate that further refinement of the benthic flux model is justified. Both this study and the previous modeling effort demonstrate the significance of benthic fluxes to the overall dissolved oxygen budget. While special efforts were undertaken to rationally specify the temporal and spatial variation in SOD, the method by which benthic fluxes are prescribed rather than predicted from a sediment diagenesis model does limit the predictive ability of this model. For instance, it was recognized that reductions in river loading would probably also reduce sediment oxygen demand in the longterm, but with the prescribed SOD there was no way to predict the magnitude of the changes in SOD, or the time scale of those changes. Also, the monitoring data indicated that spatial differences in SOD probably exist within the estuary, but the DO monitoring data alone don't provide sufficient information to rationally prescribe the spatial variation in sediment quality.

The results of the analysis are qualitatively similar to that of the previous modeling study (Tetra Tech 2001), despite the conceptual differences in the two models. Both models indicate the importance of the wetland areas within the estuary, even though conceptualization of this portion of the water body differs between the two models. The previous model only considered the wetlands to be a sink of dissolved oxygen; consideration of organic matter loadings from these areas was not considered. In the model described here, the tidal creeks and wetlands were considered sinks of dissolved oxygen as well as a source of organic matter and freshwater. In the previous modeling work the "marsh effect" was estimated by completely removing these cells from the model, which had both water quality and circulation effects. In the modeling work described here, the effects of the wetlands were lumped together with that of the major rivers that drain to the estuary. These effects were quantified by varying the loadings from the all of the riverine and non-point sources, and then observing the resulting effects on dissolved oxygen concentrations within the estuary. With the analysis done this way, it is not possible to distinguish between the effect caused by the riverine sources and that from the wetlands and tidal creeks. One good aspect of this model, however, is that a scenario that looks separately at the effects of the rivers and the wetlands could be easily formulated and may be warranted in the future.

It is hoped that the model will be used as a tool for the rational water quality management of the system. The model was created with the objective of considering all sources of oxygen demanding waste to the estuary. A second objective was to quantify these sources in such way so that additional scenario testing would not be overly burdensome. Considering the complexity of the system, the importance of the natural resource, and the importance of the economy supported by the estuary's resources, further analysis in the future using this model seems warranted.

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| Parameters Used to Calculate Loads | Value | Units |
| :--- | ---: | :--- | Notes | Name | N7257 |
| :--- | :--- |
| Starting Day | days |
| Day $0=1 / 1 / 2002$ |  |
| Sample Day | 15.5 days |

Starting Day
Sample Day
37257 days Day $0=1 / 1 / 2002$
15.5 days

## Load Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no no no

|  | Each con | nstituent v | value is lin | nearly relat | ted to the | following | 7 measur | ements | a |  |  | concen | ntratio | flow. |  |  |  |  | g | factor | 䢒 | stituent |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BC | Bd | Bg | RPOC | LPOC | DOC | POP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | su |  | SA |  | COD | DO | TAM | FCB | Flow |  |
| 00300 - DO, Oxygen, Dissolved, g/m3 | 0.0 | 0.0 | 0.00 | 0.0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |  | 0.00 | 0.0 | 3.79 | 0.00 | 0.00 |  | . 00 |
| 00310 - BOD, 5-Day (20 Deg_ C) Ib/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.32 | 0.0 | 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 |  | 0.00 |  | 0.00 | 0.00 | . 00 | 0.0 | 0.00 |  | 0.00 |
| 00530 - Solids, Total Suspended, Ib/day | 0.00 | , 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 |  | 0.00 |
| 00600 - Nitrogen, Total (as N), lb/day | . 00 | 0.00 | 0.00 | . 00 | 0.00 | . 00 | 0.00 | 0.00 | 0.0157 | -0.0157 | 0.00 | 0.00 | 0.1134 | 0.00 | 0.3402 |  | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.0 |
| 00610 - Nitrogen, Ammonia Total (as N), Ib/day | 0.0 | 0.00 | 0.0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | -0.0157 | 0.015 | 0.00 | 0.00 | -0.1134 | 0.45 | -0.3402 |  | 0.00 |  | 0.00 | 0.0 | 0.00 | 0.0 | 0.00 |  | 0.0 |
| 00665 - Phosphorus, Total (as P), Ib/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 | 0.45 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.0 |  | 0.00 | 0.00 | 0.0 | 0.00 | 0.0 |  | 0.0 |
| 50050 - Flow, in conduit or thru treatment plant, <br> MGD | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 1.0 |

Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio | 0.4536 | 0.4536 | 0.4536 | 0.4536 | 0.4536 | 0.4536 | 0.4536 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BOD5 to BODU scaleup | 1.9 | 1.9 |  |  |  |  |  |  |
| C to O 2 molecular weight ratio (12/32) | 0.3745 | 0.3745 |  |  |  |  |  |  |
| kilograms per gram |  |  |  |  |  |  |  | 0.001 |
| m3 per million gallon |  |  |  |  |  |  |  | 3786.2 |
| fraction of non-ammonia nitrogen that is DON |  |  | 0.25 | 0.25 | 0.25 |  |  |  |
| fraction of non-ammonia nitrogen that is NOX |  |  |  |  |  |  | 0.75 |  |
| P to N mass ratio |  |  | 0.1389 | 0.1389 |  |  |  |  |
| LPOC/DOC split (see jdb_1485) | 0.0\% | 100.0\% |  |  |  |  |  |  |
| Fraction Remaining | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% | 100\% |
| Resulting Unit in Computed Data |  | kg/d | kg/d | kg/d | kg/d | kg/d | kg/d | kg/MG |

kg/d
$\mathrm{DON}=(\mathrm{TN}-\mathrm{NH} 4) * .25$
NO3 $=($ TN - NH4) $* 0.75$
DOP $=$ DON * (Redfield P to N mass ratio)
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |  |
| :---: | :---: | :---: | :---: |
| Name | Value | Units | Notes |
| Starting Day | 37257 | days | Day $0=1 / 1 / 2002$ |
| Sample Day | 15.5 | days | Noon on the 15th of the month |

## Load Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no no
no no
no
no
no
no
no
no
no
no
no
no
yes
no
no

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BC | Bd | 1 Bg | RPOC | LPOC | IDOC | RPOP | LPOP | DOP | PTO4 | RPON | ILPON | DON | NH4 | NO3 | SU | SA | COD | IDO | ITAM | FCB |
| 00300 - DO, Oxygen, Dissolved | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $\begin{gathered} 00310 \text { - BOD, 5-Day (20 Deg_C }) \\ \text { Ib/day } \end{gathered}$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 00530 - Solids, Total Suspended | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 00600 - Nitrogen, Total (as N) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0157 | -0.0157 | 0.00 | 0.00 | 0.11 | 0.00 | 0.34 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 00610 - Nitrogen, Ammonia Total (as N) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | -0.0157 | 0.0157 | 0.00 | 0.00 | -0.11 | 0.45 | -0.34 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 00665 - Phosphorus, Total (as P) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.45 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 50050 - Flow, in conduit or thru treatment plant |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DO * flow (g/m3 MGD) | 0.00 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 |

## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to lb ratio
BOD5 to BODU scaleup
C to O 2 molecular weight ratio (12/32)

| 0.4536 |
| ---: |
| 1.9 |
| 0.3745 |

kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
to N mass ratio
Fraction Remaining
$0.25 \quad 0.25$
0.25

### 0.13890 .1389

| 0.4536 | 0.4536 |
| :--- | :--- |


| $100 \%$ | $100 \%$ |
| :---: | :---: |

## Assumptions

DOC $=$ BODu * C to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4) * .25$
NO3 $=($ TN - NH4 $4 * * .75$
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |  |
| :---: | :---: | :---: | :---: |
| Name | Value | Units | Notes |
| Starting Day | 37257 | days | Day $0=1 / 1 / 2002$ |
| Sample Day | 15.5 | days | Noon on the 15th of the month |

## Load Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no no
no no
no
no
no
no
no
no
no
no
no
no
no
yes
no
no


## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to lb ratio
BOD5 to BODU scaleup
C to O 2 molecular weight ratio (12/32)

| 0.4536 |
| ---: |
| 1.9 |
| 0.3745 |

kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
to N mass ratio
Fraction Remaining
$0.25 \quad 0.25$
0.25

### 0.13890 .1389

| 0.4536 | 0.4536 |
| :--- | :--- |


| $100 \%$ | $100 \%$ |
| :---: | :---: |

## Assumptions

DOC $=$ BODu * C to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4) * .25$
NO3 $=($ TN - NH4 $4 * * .75$
PO4T = TP - DOP

## arameters Used to Calculate Loads

Name
Starting Day
Sample Day
$\begin{array}{cl}\text { Value } & \text { Units } \\ 37257 & \text { dotes } \\ \text { days } & \text { Day } 0=1 / 1 / 2002\end{array}$
15.5 days $\quad \begin{aligned} & \text { Doon on the } 15 \text { th of the month }\end{aligned}$

## Load Calculation Matrix (to get results in kg/day)

|  | Multiply by <br> no | no to | $\begin{aligned} & \text { get Daily } \\ & \text { no } \end{aligned}$ |  | no | no | yes | yes | yes | yes | yes | yes | yes | yes | yes | no |  | no | no | yes | no | no |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Each constituent concentration is linearly related to the following 8 measurements (6 loads, 1 conc. (DO), 1 flow) Each column gives weighting factors for each constituent |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | BC |  | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | SU |  | SA | COD | DO | TAM | IFCB | Flow |  |
| 00300 - DO, Oxygen, Dissolved, g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 |  | 0.00 |
| 00310 - BOD, 5-Day ( 20 Deg_C), lb/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00530 - Solids, Total Suspended, lib/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00600 - Nitrogen, Total (as N), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.1315 | -0.1315 | 0.00 | 0.00 | 0.95 | 0.00 | 2.84 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00610 - Nitrogen, Ammonia Total (as N), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | -0.1315 | 0.1315 | 0.00 | 0.00 | -0.95 | 3.79 | -2.84 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00665 - Phosphorus, Total (as P), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 50050 - Flow, in conduit or thru treatment | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 1.00 |


\section*{Parameters Used to Get Values Above (each coefficient is the product of the numbers below it) <br> | Kg to lb ratio | 0.4536 |  |
| :--- | :--- | ---: |
| BOD5 to BODU scaleup |  | 1.9 |
| C to O2 molecular weight ratio (12/32) |  | 0.3745 |}

kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX

| 0.001 | 0.001 |
| ---: | ---: |
| 3786.2 | 3786.2 |
| $\mathbf{0 . 2 5}$ | $\mathbf{0 . 2 5}$ |


| 0.001 | 0.001 | 0.001 |
| :---: | :---: | :---: |
| 3786.2 | 3786.2 | 3786.2 |
| $\mathbf{0 . 2 5}$ |  |  |


fraction of non-ammonia nitrogen that is NOX
$0.1389 \quad 0.1389$

## Fraction Remaining

$\mathrm{OC}=\mathrm{BODu}$ * C to O 2 molecular weight ratio
DON = (TN - NH4) *. 25
NO3 $=($ TN - NH4 $) * 0.75$
NO3 $=(T N-$ NH4 $4 * 0.75$
DOP $=$ DON $*$ (Redfield $P$ to $N$ mass ratio)
PO4T $=T P$ - DOP

## arameters Used to Calculate Loads

Name
Starting Day
Sample Day
$\begin{array}{cl}\text { Value } & \text { Units } \\ 37257 & \text { dotes } \\ \text { days } & \text { Day } 0=1 / 1 / 2002\end{array}$
15.5 days $\quad \begin{aligned} & \text { Doon on the } 15 \text { th of the month }\end{aligned}$

## Load Calculation Matrix (to get results in kg/day)

|  | Multiply by <br> no | no to | $\begin{aligned} & \text { get Daily } \\ & \text { no } \end{aligned}$ |  | no | no | yes | yes | yes | yes | yes | yes | yes | yes | yes | no |  | no | no | yes | no | no |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Each constituent concentration is linearly related to the following 8 measurements (6 loads, 1 conc. (DO), 1 flow) Each column gives weighting factors for each constituent |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | BC |  | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | SU |  | SA | COD | DO | TAM | IFCB | Flow |  |
| 00300 - DO, Oxygen, Dissolved, g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 |  | 0.00 |
| 00310 - BOD, 5-Day ( 20 Deg_C), lb/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00530 - Solids, Total Suspended, lib/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00600 - Nitrogen, Total (as N), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.1315 | -0.1315 | 0.00 | 0.00 | 0.95 | 0.00 | 2.84 |  | . 00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00610 - Nitrogen, Ammonia Total (as N), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | -0.1315 | 0.1315 | 0.00 | 0.00 | -0.95 | 3.79 | -2.84 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00665 - Phosphorus, Total (as P), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 50050 - Flow, in conduit or thru treatment | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 1.00 |


\section*{Parameters Used to Get Values Above (each coefficient is the product of the numbers below it) <br> | Kg to lb ratio | 0.4536 |  |
| :--- | :--- | ---: |
| BOD5 to BODU scaleup |  | 1.9 |
| C to O2 molecular weight ratio (12/32) |  | 0.3745 |}

kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX

| 0.001 | 0.001 |
| :---: | :---: |
| 3786.2 | 3786.2 |
| $\mathbf{0 . 2 5}$ |  |


| 0.001 | 0.001 | 0.001 |  |
| :---: | :---: | :---: | :---: |
| 3786.2 | 3786.2 | 3786.2 |  |
| $\mathbf{0 . 2 5}$ |  |  |  |
|  |  |  |  |


fraction of non-ammonia nitrogen that is NOX
$0.1389 \quad 0.1389$

## Fraction Remaining

$\mathrm{OC}=\mathrm{BODu}$ * C to O 2 molecular weight ratio
DON = (TN - NH4) *. 25
NO3 $=($ TN - NH4 $) * 0.75$
NO3 $=(T N-$ NH4 $4 * 0.75$
DOP $=$ DON $*$ (Redfield $P$ to $N$ mass ratio)
PO4T $=T P$ - DOP

## arameters Used to Calculate Loads

Name
Starting Day
Sample Day
Value Units Notes
37257 days
37257 days $\quad \begin{aligned} & \text { Notes } \\ & \text { Day } 0\end{aligned}=1 / 1 / 2002$
15.5 days Noon on the 15 th of the month

## Load Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
Each constituent concentration is linearly related to the following 8 measurements ( 6 loads, 1 conc. (DO), 1 flow) Each column gives weighting factors for each constituent

|  | Eac | uent | concentratio | dion is linealy | dPar related | DOC | following 8 | measure | ments | loads, 1 con | conc. (DO), | , 1 flow) | ach | umn gives | weighting | g factors | for each co | onstituent |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 00300 - DO, Oxygen, Dissolved, g/m3 | BC 0.00 |  | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH 4 | 1 NO | SU | SA | COD | DO | TAM | FCB | Flow | 0.00 |
| 00310 - BOD, 5-Day (20 Deg_ C), Ib/day | 0.00 | 0.00 | 0.00 | 0.78 | 0.00 | 0.26 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00530 - Solids, Total Suspended, Ib/day | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00600 - Nitrogen, Total (as N), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.00 | 0.03 | -0.13 | 0.71 | 0.00 | 0.24 | 0.00 | 2.84 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00610 - Nitrogen, Ammonia Total (as N), $\mathrm{g} / \mathrm{m} 3$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | -0.10 | 0.00 | -0.03 | 0.13 | -0.71 | 0.00 | -0.24 | 3.79 | -2.84 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 00665 - Phosphorus, Total (as P), g/m3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.79 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |
| 50050 - Flow, in conduit or thru treatment plant, mg/D | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 1.00 |

Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio | 0.4536 | 0.4536 |  |
| :---: | :---: | :---: | :---: |
| BOD5 to BODU scaleup | 6.11 | 6.11 |  |
| C to O 2 molecular weight ratio (12/32) | 0.3745 | 0.3745 |  |
| kilograms per gram |  |  | 0.001 |
| m3 per million gallon |  |  | 3786.2 |
| fraction of non-ammonia nitrogen that is ON |  |  | 0.25 |
| fraction of non-ammonia nitrogen that is NOX |  |  |  |
| P to N mass ratio |  |  | 0.1389 |
| LPOC/DOC split (see jdb_1485) | 75.0\% | 25.0\% | 75.0\% |
| Fraction Remaining | 100\% | 100\% | 100\% |


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.2 |
| $\mathbf{0 . 2 5}$ | $\mathbf{0 . 2 5}$ | $\mathbf{0 . 2 5}$ |


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.2 |
| $\mathbf{0 . 2 5}$ |  |  |


| $\mathbf{0 . 1 3 8 9}$ | 0.1389 |  |
| ---: | ---: | ---: |
| $25.0 \%$ |  | $75.0 \%$ |
| $100 \%$ | $100 \%$ | $100 \%$ |


| $25.0 \%$ |  |  |
| ---: | :---: | :---: |
| $100 \%$ | $100 \%$ | $100 \%$ |

$100 \%$
Assumptions
DOC $=\mathrm{BODu}$ $\mathrm{C}_{\mathrm{C}}$ to O 2 molecular weight ratio
DON $=(\text { TN }- \text { NH4 })^{*} .25$
DON $=(\text { TN }- \text { NH4 })^{*} .25$
NO3 $=($ TN - NH4 0.75
DOP $=$ DON * (Redfield $P$ to $N$ mass ratio)
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \text { Multiply by Flow to get Daily Load? } \\
& \text { no no no no no yes yes ye }
\end{aligned}
$$



## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to Ib ratio
BOD5 to BODU scaleup
C to O2 molecular weight ratio (12/32)
kilograms per gram
m 3 per million gallon
fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX
P to N mass ratio
Fraction Remaining
Assumptions
DOC $=$ BODu ${ }^{*} \mathrm{C}$ to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4)^{*} .25$
NO3 $=\left(\mathrm{TN}-\mathrm{NH}^{*}\right)^{*} 0.75$
DOP $=$ DON * (Redfield P to N mass ratio)
PO4T $=T \mathrm{TP}-$ DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \text { Multiply by Flow to get Daily Load? } \\
& \text { no no no no no yes yes ye }
\end{aligned}
$$



## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to Ib ratio
BOD5 to BODU scaleup
C to O2 molecular weight ratio (12/32)
kilograms per gram
m 3 per million gallon
fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX
P to N mass ratio
Fraction Remaining
Assumptions
DOC $=$ BODu ${ }^{*} \mathrm{C}$ to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4)^{*} .25$
NO3 $=\left(\mathrm{TN}-\mathrm{NH}^{*}\right)^{*} 0.75$
DOP $=$ DON * (Redfield P to N mass ratio)
PO4T $=T \mathrm{TP}-$ DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

```
Multiply by Flow to get Daily Load?
```

no no no yes no yes yes yes

Each constuent concentration is


| $00300-$ DO, Oxygen, Dissolved, |
| :---: |
| $\mathrm{q} / \mathrm{m} 3$ |



| 0.00 | 0.00 |
| :--- | :--- |
|  | 0.00 |
|  | 0.00 |
| 0.3315 | 0.0 |
| 0.315 | 0.00 |
| 3.79 | 0.00 |


| 0.00 | 0.00 |
| :--- | :--- |
| 0.00 | 0.00 |
| 0.00 | 0.00 |
| 0.00 | 0.00 |
| 0.00 | 0.00 |
| 0.00 | 0.00 |

0.00
ighting fa
no
no
yes


Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)
Kg to Ib ratio
BOD5 to BODU scaleup

C to O 2 molecular weight ratio (12/32)
kilograms per gram
m 3 per million gallon
fraction of non-ammonia nitrogen that is DON

| 3.51 |
| ---: | ---: |
| 0.3745 |
| 0.001 |
| 3786.2 |$\quad$| 3.51 |  |
| ---: | ---: |
| 0.3745 |  |
| 0.001 | 0.001 |
| 3786.2 | 3786.2 |


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.2 |
| 0 | 0.25 |  |


0.75

P to N mass ratio
RPOC/DOC split (see jdb_1485)
Fraction Remaining
Assumptions
DOC $=\mathrm{BODu}$ * C to O 2 molecular weight ratio
DON $=($ TN - NH 4$) * .25$
NO3 $=($ TN - NH4 $) * 0.75$
DOP $=$ DON * (Redfield $P$ to $N$ mass ratio
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \text { Multiply by Flow to get Daily Load? } \\
& \text { no no no yo no yes yes ye }
\end{aligned}
$$



Parameters Used to Get Values Above (each coefficient is the product of the numbers below it) $\qquad$
Kg to lb ratio
BOD5 to BODU scaleup
C to O 2 molecular weight ratio (12/32)
kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX


0.25

| $100.0 \%$ |  |  |
| ---: | :---: | :---: |
| $100 \%$ | $100 \%$ | $100 \%$ |

Rema
Assumptions
DOC $=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4)^{*} .25$
DOP $=$ DON * (Redfield $P$ to $N$ mass ratio)
POP = DON $=$ TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

```
Multiply by Flow to get Daily Load?
```

no no no yes no yes yes yes



## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to lb ratio
BOD5 to BODU scaleup
C to O 2 molecular weight ratio (12/32)
kilograms per gram
m3 per million gallon

| 1.9 |
| ---: |
| 0.3745 |
| 0.001 |
| 3786.2 |$\quad$| 1.9 |
| ---: |
| 0.3745 |
| 0.001 |
| 3786.2 |


| 0.001 | 0.001 |
| ---: | ---: |
| 3786.2 | 3786.2 |$\quad$| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.2 |

$$
\begin{array}{|r|}
\hline 0.001 \\
\hline 3786.2 \\
\hline
\end{array}
$$

fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX
$0.25 \quad 0.25$
0.25
0.75

LPOC/DOC split (see jdb_1485)
Fraction Remaining

|  |  | 0.1389 | 0.1389 |
| :---: | :---: | :---: | :---: |
| 0.0\% | 100.0\% |  |  |
| 100\% | 100\% | 100\% | 100\% |


| $100 \%$ | $100 \%$ | $100 \%$ |
| :--- | :--- | :--- |

Assumptions
DOC $=\mathrm{BODu}$ * C to O 2 molecular weight ratio
DON $=($ TN - NH 4$) * .25$
NO3 $=($ TN - NH4 $) * 0.75$
DOP $=$ DON * (Redfield $P$ to $N$ mass ratio
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :--- | :--- |
| Name | Value | Units | Notes | Starting Day | 37257 |
| :--- | :--- |
| days | Day $0=1 / 1 / 2002$ |
| Sample Day | 15.5 days |

## Load Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?

$$
\begin{aligned}
& \text { Hultiply by } \\
& \text { no low to get Daily Load? } \\
& \text { no no no no no no no no no }
\end{aligned}
$$

no no no no


Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio | 0.4536 |  |  | 0.4536 | 0.4536 | 0.4536 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BOD5 to BODU scaleup | 1.9 |  |  |  |  |  |  |
| C to O 2 molecular weight ratio (12/32) | 0.3745 |  |  |  |  |  |  |
| kilograms per gram |  | 0.001 | 0.001 |  |  |  | 0.00 |
| m3 per million gallon |  | 3786.2 | 3786.2 |  |  |  | 3786.2 |

fraction of non-ammonia nitrogen that is DON
fraction of non-ammonia nitrogen that is NOX
$3786.2 \quad 3786.2$
0.001
fraction of non-ammonia nitrogen that is NOX
0.5
0.25
0.75

## Fraction Remaining

100\%

| $100 \%$ | $100 \%$ |
| :---: | :---: |

## Assumptions

DOC $=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
$\mathrm{ON}=(\mathrm{TN}-\mathrm{NH} 4) * .25$
NO3 $=(\mathrm{TN}-\mathrm{NH} 4) * 0.75$
DOP $=1 / 2 \mathrm{TP}$
$\mathrm{PO} 4=1 / 2 \mathrm{TP}$

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \begin{array}{l}
\text { Multiply by } \\
\text { no low to get Daily Load? } \\
\text { no no no no no no } \\
\text { no no }
\end{array} \text { yes }
\end{aligned}
$$




## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

Kg to lb ratio
BOD5 to BODU scaleup
C to O 2 molecular weight ratio (12/32) $\square$
kilograms per gram
m3 per million gallon
fraction of non-ammonia nitrogen that is DON
raction of non-ammonia nitrogen that is NOX

| 0.001 | 0.001 |
| ---: | ---: |
| 3786.2 | 3786.2 |
| 0.25 | 0.25 |


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.2 |
| $\mathbf{0}$ |  |  |


| 0.001 |
| ---: |
| 3786.2 |

P to N mass ratio
Fraction Remaining
$\longrightarrow$

## $100 \%$

0.75

## Assumptions

DOC $=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
ON $=(\mathrm{TN}-\mathrm{NH} 4)^{*} .25$
$D \mathrm{P}=\mathrm{DON}$
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \begin{array}{l}
\text { Multiply by Flow to get Daily Load? } \\
\text { no no no no no no yes yo }
\end{array}
\end{aligned}
$$


Parameters Used to Get Values Above (each coefficient is the product of the numbers

| Kg to Ib ratio | 0.4536 |
| :--- | ---: |
| BOD5 to BODU scaleup | 1.9 |
| C to O2 molecular weight ratio (12/32) | 0.3745 |


| C to O2 molecular weight ratio (12/32) | 0.3745 |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| kilograms per gram |  | 0.001 | 0.001 |  |
| m3 per million gallon |  | \begin{tabular}{\|l|r|r|}
\hline
\end{tabular} | 0.001 | 0.001 |
| fraction of non-ammonia nitrogen that is DON | 0.00 |  |  |  |

fraction of non-ammonia nitrogen that is DON
$0.25 \quad 0.25$
0.25

fraction of non-ammonia nitrogen that is NOX
0.13890 .1389

P to N mass ratio

|  | $100 \%$ |
| :--- | :--- |

$\mathrm{OOC}=\mathrm{BODu}{ }^{*} \mathrm{C}$ to O 2 molecular weight ratio
DON = (TN - NH4) * 25
NO3 $=($ TN - NH4 $) * 0.75$
DOP $=$ DON * (Redfield P to N mass ratio)
PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \text { Multiply by Flow to get Daily Load? } \\
& \text { no no no no no yes yes ye }
\end{aligned}
$$



Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BOD5 to BODU scaleup $\quad 1.9$ |  |  |  |  |  |  |  |
| C to O 2 molecular weight ratio (12/32) | 0.3745 |  |  |  |  |  |  |
| kilograms per gram | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| m3 per million gallon | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 |
| fraction of non-ammonia nitrogen that is DON |  | $\begin{array}{ll}0.25 & 0.25\end{array}$ |  | 0.25 |  |  |  |
| fraction of non-ammonia nitrogen that is NOX |  |  |  |  |  | 0.75 |  |
| P to N mass ratio |  | 0.1389 | 0.1389 |  |  |  |  |
| Fraction Remaining | 100\% |  |  | 100\% | 100\% |  |  |

Assumptions

DOC $=\mathrm{BODu}$ * C to O 2 molecular weight ratio

(TN - NH4) *. 25

DOP $=$ DON * (Redfield P to N mass ratio)

PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Notes |
| :--- |
| Starting Day |
| Sample Day |

## Load Calculation Matrix (to get results in kg/day)

$$
\begin{aligned}
& \text { Multiply by Flow to get Daily Load? } \\
& \text { no no no no no yes yes ye }
\end{aligned}
$$



Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BOD5 to BODU scaleup $\quad 1.9$ |  |  |  |  |  |  |  |
| C to O 2 molecular weight ratio (12/32) | 0.3745 |  |  |  |  |  |  |
| kilograms per gram | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| m3 per million gallon | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 | 3786.2 |
| fraction of non-ammonia nitrogen that is DON |  | 0.25 | 0.25 | 0.25 |  |  |  |
| fraction of non-ammonia nitrogen that is NOX |  |  |  |  |  | 0.75 |  |
| P to N mass ratio |  | 0.1389 | 0.1389 |  |  |  |  |
| Fraction Remaining | 100\% |  |  | 100\% | 100\% |  |  |

[^0]PO4T = TP - DOP

| Parameters Used to Calculate Loads |  |  |
| :--- | :---: | :--- |
| Name | Value | Units | Notes | Starting Day | 37257 |
| :--- | :--- |
| days | Day $0=1 / 1 / 2002$ |
| Sample Day | 15.5 days |

## Load Calculation Matrix (to get results in kg/day)

> Multiply by Flow to get Daily Load?
no no no

$$
\text { Each constituent concentration is linearly related to the following } 8 \text { measurements ( } 6 \text { loads, } 1 \text { conc. (DO), } 1 \text { flow) Each column gives weighting factors for each constituent }
$$



## Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)



| Param |  |  |  |
| :---: | :---: | :---: | :---: |
| Name | Value | Units | Notes |
| Starting Day | 37257 | days | Day $0=1 / 1 / 2002$ |
| Sample Day |  |  | Noon on the 15th of the |

## _oad Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no
no no no no yes yes yes yes
yes yes
s yes
no
no
no
yes
no
no

|  | Eac |  |  |  |  |  |  |  |  |  |  | v) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Bd | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | su | SA | COD | DO | TAM | FCB | Flow |
| 00300 - DO, Oxygen, Dissolved, | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | $3.8 \mathrm{E}+00$ | 0.0E+00 | 0.0E+00 | .0E+00 |
| $\begin{aligned} & 00310-\text { BOD, } 5 \text {-Day (20 Deg_C), } \\ & \text { q/m3 } \end{aligned}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 2.7E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 |
| $\begin{gathered} 00530 \text { - Solids, Total Suspended, } \\ \mathrm{g} / \mathrm{m} 3 \end{gathered}$ | E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | E+00 | E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | +00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+ | 0.0E+00 | 0.0E+00 | 0 |
| 00600 - Nitrogen, Total (as N), | 0.0E+00 | E+0 | E+0 | 0.0E+00 | 0 | 0 | 0 | 0.0E+00 | 3E-01 | -1.3E-01 | 0.0E+00 | 0.0E+00 | 9.5E-01 | 0.0E+00 | $2.8 \mathrm{E}+00$ | 0.0E+00 | 0.0E+0 | 0.0E+00 | 0.0E+00 | 0.0E+ | E+00 | 0.0E+00 |
| 00610 - Nitrogen, Ammonia Total <br> (as N), $\mathrm{a} / \mathrm{m} 3$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | -1.3E-01 | 1.3E-01 | 0.0E+00 | 0.0E+00 | -9.5E-0 | 3.8E+00 | -2.8E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 |
| 00665 - Phosphorus, Total (as P), <br> $\mathrm{g} / \mathrm{m} 3$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 3.8E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | OE+ | $0.0 \mathrm{E}+00$ |
| 50050 - Flow, in conduit or thru treatment plant, $\mathrm{mg} / \mathrm{D}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 1.0E+00 |

Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio |  |
| :--- | ---: |
| BOD5 to BODU scaleup | 1.9 <br> C to O2 molecular weight ratio (12/32) |
| kilograms per gram | 0.3745 |

## 3 per million gallon

fraction of non-ammonia nitrogen that is DON


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.18 |
| 0.25 |  |  |

fraction of non-ammonia nitrogen that is NOX
0.25

P to N mass ratio
$0.1389 \quad 0.1389$
Fraction Remaining

## 100\%

| $100 \%$ | $100 \%$ |
| :--- | :--- |

## Assumptions

$\mathrm{DOC}=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4) * .25$
NO3 $=(\mathrm{TN}-\mathrm{NH} 4) *$.
$\mathrm{DOP}=\mathrm{DON} *$ (Redfield P to N mass ratio $)$
PO4T = TP - DOP

| Param |  |  |  |
| :---: | :---: | :---: | :---: |
| Name | Value | Units | Notes |
| Starting Day | 37257 | days | Day $0=1 / 1 / 2002$ |
| Sample Day |  |  | Noon on the 15th of the |

## _oad Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no
no no no no yes yes yes yes
yes no
no
no
no
no
no
yes no
no

|  | Eac |  |  |  |  |  |  |  |  |  |  | v) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Bd | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | su | SA | COD | DO | TAM | FCB | Flow |
| 00300 - DO, Oxygen, Dissolved, 9/m3 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 3.8E+00 | 0.0E+00 | 0.0E+00 | .0E+00 |
| $00310-B O D, 5-$ Day (20 Deg_ C), q/m3 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 2.7E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+0 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E | 0.0E+00 |
| 00530 - Solids, Total Suspended, lb/day | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 |
| 00600 - Nitrogen, Total (as N), | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 1.3E-01 | -1.3E-01 | 0.0E+00 | 0.0E+00 | -01 | 0.0E+00 | 3.4E-0 | OE+00 | 0.0E+00 | 0.0E+00 | 0.0E+ | 0.0E+ | 0.0E | .0E+00 |
| 00610 - Nitrogen, Ammonia Total <br> (as N), lb/day | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | -1.3E-01 | 1.3E-01 | 0.0E+00 | 0.0E+ | 1E-0 | 4.5E-0 | -3.4E-01 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+ | 0.0E+ | 0.0E+00 | .0E+00 |
| $\begin{gathered} 00665 \text { - Phosphorus, Total (as P), } \\ \mathrm{q} / \mathrm{m} 3 \end{gathered}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 3.8E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | E+ | $0.0 \mathrm{E}+00$ |
| 50050 - Flow, in conduit or thru treatment plant, $\mathrm{mg} / \mathrm{D}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 1.0E+00 |

Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)


Assumptions
DOC $=$ BODu C to O 2 molecular weight ratio
DON $=(\mathrm{TN}-\mathrm{NH} 4) * .25$

DOP $=$ DON * (Redfield $P$ to $N$ mass ratio)
PO4T = TP - DOP

| Param |  |  |  |
| :---: | :---: | :---: | :---: |
| Name | Value | Units | Notes |
| Starting Day | 37257 | days | Day $0=1 / 1 / 2002$ |
| Sample Day |  |  | Noon on the 15th of the |

## _oad Calculation Matrix (to get results in kg/day)

Multiply by Flow to get Daily Load?
no no no no
no no no no yes yes yes yes
yes yes
s yes
no
no
no
yes
no
no

|  | Eac |  |  |  |  |  |  |  |  |  |  | v) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Bd | Bg | RPOC | LPOC | DOC | RPOP | LPOP | DOP | PTO4 | RPON | LPON | DON | NH4 | NO3 | su | SA | COD | DO | TAM | FCB | Flow |
| 00300 - DO, Oxygen, Dissolved, | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | $3.8 \mathrm{E}+00$ | 0.0E+00 | 0.0E+00 | .0E+00 |
| $\begin{aligned} & 00310-\text { BOD, } 5 \text {-Day (20 Deg_C), } \\ & \text { q/m3 } \end{aligned}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 2.7E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 |
| $\begin{gathered} 00530 \text { - Solids, Total Suspended, } \\ \mathrm{g} / \mathrm{m} 3 \end{gathered}$ | E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | E+00 | E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | +00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+ | 0.0E+00 | 0.0E+00 | 0 |
| 00600 - Nitrogen, Total (as N), | 0.0E+00 | E+0 | E+0 | 0.0E+00 | 0 | 0 | 0 | 0.0E+00 | 3E-01 | -1.3E-01 | 0.0E+00 | 0.0E+00 | 9.5E-01 | 0.0E+00 | $2.8 \mathrm{E}+00$ | 0.0E+00 | 0.0E+0 | 0.0E+00 | 0.0E+00 | 0.0E+ | E+00 | 0.0E+00 |
| 00610 - Nitrogen, Ammonia Total <br> (as N), $\mathrm{a} / \mathrm{m} 3$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | -1.3E-01 | 1.3E-01 | 0.0E+00 | 0.0E+00 | -9.5E-0 | 3.8E+00 | -2.8E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 |
| 00665 - Phosphorus, Total (as P), <br> $\mathrm{g} / \mathrm{m} 3$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 3.8E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | OE+ | $0.0 \mathrm{E}+00$ |
| 50050 - Flow, in conduit or thru treatment plant, $\mathrm{mg} / \mathrm{D}$ | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 0.0E+00 | 1.0E+00 |

Parameters Used to Get Values Above (each coefficient is the product of the numbers below it)

| Kg to lb ratio |  |
| :--- | ---: |
| BOD5 to BODU scaleup | 1.9 <br> C to O2 molecular weight ratio (12/32) |
| kilograms per gram | 0.3745 |

## 3 per million gallon

fraction of non-ammonia nitrogen that is DON


| 0.001 | 0.001 | 0.001 |
| ---: | ---: | ---: |
| 3786.2 | 3786.2 | 3786.18 |
| 0.25 |  |  |

fraction of non-ammonia nitrogen that is NOX
0.25

P to N mass ratio
$0.1389 \quad 0.1389$
Fraction Remaining

## 100\%

| $100 \%$ | $100 \%$ |
| :--- | :--- |

## Assumptions

$\mathrm{DOC}=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
DON $=(\text { TN }- \text { NH } 4)^{*} .25$
NO3 $=\left(\mathrm{TN}-\right.$ NH4 $4 * 0 .{ }^{*} \mathrm{~F}$
$\mathrm{DOP}=\mathrm{DON} *$ (Redfield P to N mass ratio)
PO4T = TP - DOP

C cell.inp file, i columns and j rows, for Cape Fear River
C 1 2 3 4
C
1234567890123456789012345678901234567890123456
C
990000000000000000000000000000000000000000000000
980000000000000000000000000000000000000000099990
970000000000000000000000000000000000000000095590
960000000000000000000000000000000000000000095590
950000000000000000000000000000000000000000095590
940000000000000000000000000000000000000000095590
930000000000000000000000000000000000000000095590
920000000000000000000000000000000000000000095590
910000000000000000000000000000000000000000095590
900000000000000000000000000000000000000000095590
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510095595590000000000955555955900000000000095590
500095595590000000000955455455900000000000095590
490095595590000000000952555555900000000000095590

CELL.INP
0995590000000000000009555955555559000000000000
110955590000000000000009555155555559000000000000
10099559000000000000000935555555555900000000000
90095590000000000000009955555555559000000000000
80095590000000000000000955559955559000000000000
70095590000000000000000955299955559000000000000
60099990000000000000000999990955559000000000000
50000000000000000000000000000955559000000000000
40000000000000000000000000000999999000000000000
30000000000000000000000000000000000000000000000
20000000000000000000000000000000000000000000000
10000000000000000000000000000000000000000000000
C
C
C

0095595590000000000931555552900000000000095590 0095595590000000000993555559900000000000095590 0095595590000000000094555551999000000000095590 0095595590000000000093555155519999990000095590 0095595590000000000099555555551955590000095590 0095595590000000000099555535552999990000095590 0095595590000000000095555595559900000000095590 0095595590000000000099555595551900000000095590 0095595590000000000099555555555990000000095590 0095595590000000000994555523555190000000095590 0095995590000000009945555555555290000000095590 0095555990000000009455555555955990000000095590 0095999900000000094555555555519900000000095590 0095590000000000095555555525529000000000095590 0095590000000000095523555593519000000000095590 0095590000000000093555555594529000000000095590 0095590000000000099555555595519000000000095590 0095590000000000009555555555559000000000095590 0095590000000000009355555555559900000000095590 0095590000000000009935555555551900000000095590 0095590000000000999995555555529900000000095590 0095590000000009945555555555299000000000095590 0095590000000009599959555555999900000000095590 0095590000000009599959555555555900000000095590 0095590000000009590959555555599990000000095590 0095590000000009599954555555555590000000095590 0095590000000009319355555555555599000000095590 0095590000000009931955555555555559000000095590 0095590000000000995555555255555559900000995590 0095590000000000099355555455555551990000945590 0095590000000000009995555555555293199000955590 0095590000000000000093555555295555519900955590 0095590000000000000099555555193599451900955590 0095590000000000000009555555524594555900995990 0095590000000000000009555555255545999900099900 0095590000000000000009555355155559900000000000 0995590000000000000009555955555559000000000000 0955590000000000000009555155555559000000000000 0995590000000000000009355555555559000000000000 0095590000000000000009955555555559000000000000 0095590000000000000000955559955559000000000000 0095590000000000000000955299955559000000000000 0099990000000000000000999990955559000000000000 0000000000000000000000000000955559000000000000 0000000000000000000000000000999999000000000000 0000000000000000000000000000000000000000000000 000000000000000000000000000000000000000000000
10000000000000000000000000000000000000000000000 1234567890123456789012345678901234567890123456

CELL.INP

## C dxdy.inp file, in free format across columns



| 32 | 10 | 396.96 | 388.02 | 4.83 | -4.53 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 33 | 10 | 401.39 | 402.40 | 4.00 | -3.70 | 0.015 |
| 3 | 11 | 102.96 | 519.71 | 4.00 | -3.70 | 0.015 |
| 4 | 11 | 102.96 | 519.71 | 6.50 | -6.20 | 0.015 |
| 5 | 11 | 952.46 | 519.71 | 0.30 | 0.00 | 0.035 |
| 23 | 11 | 308.45 | 565.33 | 6.65 | -6.35 | 0.015 |
| 24 | 11 | 224.53 | 549.10 | 12.62 | -12.32 | 0.015 |
| 25 | 11 | 231.61 | 553.22 | 7.84 | -7.54 | 0.015 |
| 26 | 11 | 389.53 | 547.83 | 3.08 | -2.78 | 0.015 |
| 27 | 11 | 406.39 | 518.22 | 3.50 | -3.20 | 0.015 |
| 28 | 11 | 425.67 | 506.80 | 3.50 | -3.20 | 0.015 |
| 29 | 11 | 444.94 | 495.38 | 3.50 | -3.20 | 0.015 |
| 30 | 11 | 422.60 | 496.68 | 3.50 | -3.20 | 0.015 |
| 31 | 11 | 397.62 | 507.27 | 3.04 | -2.74 | 0.015 |
| 32 | 11 | 395.98 | 499.86 | 3.00 | -2.70 | 0.015 |
| 33 | 11 | 396.09 | 488.63 | 3.00 | -2.70 | 0.015 |
| 4 | 12 | 150.75 | 524.62 | 6.50 | -6.20 | 0.015 |
| 5 | 12 | 943.54 | 524.62 | 0.30 | 0.00 | 0.035 |
| 23 | 12 | 273.34 | 598.81 | 9.55 | -9.25 | 0.015 |
| 24 | 12 | 189.10 | 603.06 | 12.50 | -12.20 | 0.015 |
| 25 | 12 | 293.87 | 594.69 | 4.29 | -3.99 | 0.015 |
| 27 | 12 | 371.39 | 591.58 | 3.00 | -2.70 | 0.015 |
| 28 | 12 | 390.67 | 580.90 | 3.00 | -2.70 | 0.015 |
| 29 | 12 | 409.95 | 570.22 | 3.00 | -2.70 | 0.015 |
| 30 | 12 | 408.75 | 562.23 | 3.00 | -2.70 | 0.015 |
| 31 | 12 | 400.95 | 556.47 | 2.67 | -2.37 | 0.015 |
| 32 | 12 | 400.20 | 556.03 | 2.50 | -2. 20 | 0.015 |
| 33 | 12 | 402.23 | 563.44 | 2.50 | -2. 20 | 0.015 |
| 4 | 13 | 141.42 | 672.68 | 6.50 | -6.20 | 0.015 |
| 5 | 13 | 735.86 | 672.68 | 0.30 | 0.00 | 0.035 |
| 23 | 13 | 252.41 | 613.21 | 6.29 | -5.99 | 0.015 |
| 24 | 13 | 189.10 | 623.44 | 12.50 | -12.20 | 0.015 |
| 25 | 13 | 304.58 | 623.26 | 5.18 | -4.88 | 0.015 |
| 26 | 13 | 345.10 | 606.82 | 2.50 | -2. 20 | 0.015 |
| 27 | 13 | 345.10 | 606.82 | 2.50 | -2. 20 | 0.015 |
| 28 | 13 | 365.60 | 605.64 | 2.50 | -2. 20 | 0.015 |
| 29 | 13 | 351.93 | 606.43 | 2.50 | -2. 20 | 0.015 |
| 30 | 13 | 358.77 | 606.03 | 2.50 | -2. 20 | 0.015 |
| 31 | 13 | 393.13 | 597.25 | 2.50 | -2. 20 | 0.015 |
| 32 | 13 | 395.63 | 597.47 | 2.74 | -2.44 | 0.015 |
| 33 | 13 | 397.70 | 603.03 | 2.00 | -1.70 | 0.015 |
| 4 | 14 | 185.00 | 600.19 | 6.98 | -6.68 | 0.015 |
| 5 | 14 | 824.74 | 600.19 | 0.30 | 0.00 | 0.035 |
| 23 | 14 | 252.41 | 585.68 | 3.63 | -3.33 | 0.015 |
| 24 | 14 | 189.10 | 575.05 | 12.54 | -12.24 | 0.015 |
| 25 | 14 | 292.16 | 568.85 | 6.93 | -6.63 | 0.015 |
| 26 | 14 | 305.24 | 580.07 | 2.00 | -1.70 | 0.015 |
| 27 | 14 | 323.03 | 593.24 | 2.00 | -1.70 | 0.015 |
| 28 | 14 | 341.53 | 604.50 | 2.00 | -1.70 | 0.015 |
| 29 | 14 | 357.97 | 612.25 | 1.90 | -1.60 | 0.015 |
| 30 | 14 | 370.67 | 616.44 | 1.82 | -1.52 | 0.015 |
| 31 | 14 | 379.72 | 618.02 | 2.00 | -1.70 | 0.015 |
| 32 | 14 | 386.09 | 617.78 | 2.11 | -1.81 | 0.015 |
| 33 | 14 | 389.81 | 613.75 | 1.50 | -1.20 | 0.015 |
| 34 | 14 | 398.29 | 590.09 | 1.71 | -1.41 | 0.015 |


| 4 | 15 | 70.71 | 514.78 | 6.98 | -6.68 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 15 | 961.58 | 514.78 | 0.30 | 0.00 | 0.035 |
| 23 | 15 | 232.72 | 518.62 | 3.48 | -3.18 | 0.015 |
| 24 | 15 | 203.73 | 519.01 | 12.50 | -12.20 | 0.015 |
| 25 | 15 | 267.28 | 527.85 | 7.38 | -7.08 | 0.015 |
| 26 | 15 | 287.84 | 567.41 | 1.74 | -1.44 | 0.015 |
| 27 | 15 | 307.36 | 591.25 | 1.70 | -1.40 | 0.015 |
| 28 | 15 | 325.26 | 606.47 | 1.70 | -1.40 | 0.015 |
| 29 | 15 | 341.16 | 616.62 | 3.00 | -2.70 | 0.015 |
| 30 | 15 | 354.76 | 623.64 | 2.00 | -1.70 | 0.015 |
| 31 | 15 | 366.10 | 628.97 | 1.50 | -1.20 | 0.015 |
| 32 | 15 | 375.52 | 633.99 | 1.50 | -1.20 | 0.015 |
| 34 | 15 | 391.31 | 651.94 | 1.50 | -1.20 | 0.015 |
| 35 | 15 | 362.71 | 665.69 | 1.50 | -1.20 | 0.015 |
| 36 | 15 | 345.64 | 649.55 | 1.50 | -1.20 | 0.015 |
| 37 | 15 | 410.70 | 617.62 | 1.50 | -1.20 | 0.015 |
| 43 | 15 | 175.40 | 625.58 | 9.50 | -9.20 | 0.015 |
| 4 | 16 | 111.80 | 585.23 | 7.31 | -7.01 | 0.015 |
| 5 | 16 | 845.82 | 585.23 | 0.30 | 0.00 | 0.035 |
| 23 | 16 | 213.90 | 573.56 | 4.69 | -4.39 | 0.015 |
| 24 | 16 | 181.95 | 606.08 | 12.50 | -12.20 | 0.015 |
| 25 | 16 | 265.58 | 631.76 | 7.57 | -7.27 | 0.015 |
| 26 | 16 | 285.37 | 618.71 | 2.33 | -2.03 | 0.015 |
| 27 | 16 | 301.78 | 614.69 | 1.70 | -1.40 | 0.015 |
| 28 | 16 | 316.24 | 615.38 | 1.74 | -1.44 | 0.015 |
| 29 | 16 | 316.24 | 615.38 | 1.74 | -1.44 | 0.015 |
| 31 | 16 | 352.89 | 627.91 | 1.50 | -1.20 | 0.015 |
| 32 | 16 | 363.04 | 634.16 | 1.50 | -1.20 | 0.015 |
| 35 | 16 | 373.09 | 661.76 | 1.50 | -1.20 | 0.015 |
| 36 | 16 | 372.83 | 678.98 | 1.50 | -1.20 | 0.015 |
| 37 | 16 | 409.99 | 701.91 | 1.50 | -1.20 | 0.015 |
| 42 | 16 | 173.30 | 664.97 | 2.47 | -2.17 | 0.015 |
| 43 | 16 | 178.85 | 690.47 | 8.50 | -8.20 | 0.015 |
| 44 | 16 | 1200.00 | 690.47 | 0.30 | 0.00 | 0.035 |
| 4 | 17 | 152.40 | 619.05 | 6.00 | -5.70 | 0.015 |
| 5 | 17 | 1200.00 | 619.05 | 0.30 | 0.00 | 0.035 |
| 22 | 17 | 277.60 | 558.33 | 1.50 | -1.20 | 0.015 |
| 23 | 17 | 208.15 | 576.02 | 5.38 | -5.08 | 0.015 |
| 24 | 17 | 142.13 | 569.94 | 12.71 | -12.41 | 0.015 |
| 25 | 17 | 283.37 | 576.10 | 7.21 | -6.91 | 0.015 |
| 26 | 17 | 290.40 | 586.29 | 2.20 | -1.90 | 0.015 |
| 27 | 17 | 299.13 | 593.15 | 2.20 | -1.90 | 0.015 |
| 28 | 17 | 308.69 | 598.64 | 1.50 | -1.20 | 0.015 |
| 29 | 17 | 318.78 | 603.66 | 1.50 | -1.20 | 0.015 |
| 31 | 17 | 340.33 | 613.76 | 1.50 | -1.20 | 0.015 |
| 32 | 17 | 351.86 | 619.24 | 1.93 | -1.63 | 0.015 |
| 33 | 17 | 363.60 | 625.06 | 1.72 | -1.42 | 0.015 |
| 34 | 17 | 374.06 | 631.04 | 1.50 | -1.20 | 0.015 |
| 35 | 17 | 379.64 | 639.75 | 1.50 | -1.20 | 0.015 |
| 36 | 17 | 357.64 | 645.77 | 1.50 | -1.20 | 0.015 |
| 42 | 17 | 163.91 | 628.06 | 6.45 | -6.15 | 0.015 |
| 43 | 17 | 175.85 | 615.84 | 8.50 | -8.20 | 0.015 |
| 44 | 17 | 1200.00 | 615.84 | 0.30 | 0.00 | 0.035 |
| 4 | 18 | 103.08 | 689.66 | 5.00 | -4.70 | 0.015 |
| 5 | 18 | 1200.00 | 689.66 | 0.30 | 0.00 | 0.035 |


| 22 | 18 | 256.33 | 572.13 | 1.50 | -1.20 | 0.015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 23 | 18 | 208.47 | 572.13 | 6.95 | -6.65 | 0.015 |
| 24 | 18 | 142.13 | 569.94 | 12.50 | -12.20 | 0.015 |
| 25 | 18 | 294.34 | 571.37 | 6.83 | -6.53 | 0.015 |
| 26 | 18 | 292.72 | 574.87 | 3.07 | -2.77 | 0.015 |
| 27 | 18 | 295.64 | 579.22 | 2.87 | -2.57 | 0.015 |
| 28 | 18 | 301.17 | 583.82 | 1.99 | -1.69 | 0.015 |
| 29 | 18 | 308.49 | 588.38 | 1.50 | -1.20 | 0.015 |
| 30 | 18 | 317.43 | 592.80 | 1.50 | -1.20 | 0.015 |
| 31 | 18 | 328.33 | 597.05 | 1.50 | -1.20 | 0.015 |
| 32 | 18 | 342.06 | 600.95 | 1.50 | -1.20 | 0.015 |
| 34 | 18 | 382.56 | 595.98 | 1.50 | -1.20 | 0.015 |
| 35 | 18 | 369.92 | 536.08 | 1.50 | -1.20 | 0.015 |
| 42 | 18 | 147.00 | 548.72 | 4.15 | -3.85 | 0.015 |
| 43 | 18 | 181.02 | 574.22 | 8.50 | -8.20 | 0.015 |
| 44 | 18 | 1100.00 | 574.22 | 0.30 | 0.00 | 0.035 |
| 4 | 19 | 143.96 | 561.89 | 4.00 | -3.70 | 0.015 |
| 5 | 19 | 1100.00 | 561.89 | 0.30 | 0.00 | 0.035 |
| 20 | 19 | 1200.00 | 606.11 | 0.30 | 0.00 | 0.035 |
| 21 | 19 | 268.74 | 591.33 | 1.50 | -1.20 | 0.015 |
| 22 | 19 | 242.74 | 575.98 | 1.50 | -1.20 | 0.015 |
| 23 | 19 | 214.36 | 572.58 | 7.41 | -7.11 | 0.015 |
| 24 | 19 | 160.70 | 587.76 | 13.03 | -12.73 | 0.015 |
| 25 | 19 | 290.92 | 595.87 | 5.56 | -5.26 | 0.015 |
| 26 | 19 | 288.52 | 582.50 | 1.69 | -1.39 | 0.015 |
| 27 | 19 | 289.64 | 578.92 | 3.44 | -3.14 | 0.015 |
| 28 | 19 | 292.95 | 579.41 | 2.73 | -2.43 | 0.015 |
| 29 | 19 | 298.09 | 581.74 | 1.74 | -1.44 | 0.015 |
| 30 | 19 | 305.21 | 585.07 | 1.55 | -1.25 | 0.015 |
| 31 | 19 | 315.05 | 589.42 | 1.60 | -1.30 | 0.015 |
| 32 | 19 | 329.66 | 596.24 | 1.50 | -1.20 | 0.015 |
| 33 | 19 | 355.08 | 611.71 | 1.50 | -1.20 | 0.015 |
| 34 | 19 | 391.49 | 599.21 | 2.27 | -1.97 | 0.015 |
| 42 | 19 | 160.82 | 683.53 | 2.49 | -2.19 | 0.015 |
| 43 | 19 | 187.33 | 706.38 | 8.50 | -8.20 | 0.015 |
| 44 | 19 | 700.76 | 706.38 | 0.30 | 0.00 | 0.035 |
| 4 | 20 | 125.00 | 497.02 | 4.00 | -3.70 | 0.015 |
| 5 | 20 | 497.97 | 497.02 | 0.30 | 0.00 | 0.035 |
| 19 | 20 | 522.22 | 544.85 | 0.30 | 0.00 | 0.035 |
| 20 | 20 | 520.46 | 563.13 | 0.30 | 0.00 | 0.035 |
| 21 | 20 | 247.16 | 578.30 | 1.50 | -1.20 | 0.015 |
| 22 | 20 | 232.22 | 589.15 | 1.50 | -1.20 | 0.015 |
| 23 | 20 | 217.55 | 599.84 | 6.54 | -6.24 | 0.015 |
| 24 | 20 | 177.84 | 570.44 | 12.50 | -12.20 | 0.015 |
| 25 | 20 | 272.68 | 548.08 | 5.51 | -5.21 | 0.015 |
| 26 | 20 | 278.30 | 566.47 | 1.50 | -1.20 | 0.015 |
| 27 | 20 | 280.92 | 573.68 | 1.97 | -1.67 | 0.015 |
| 28 | 20 | 283.23 | 577.77 | 3.64 | -3.34 | 0.015 |
| 29 | 20 | 286.59 | 581.73 | 2.37 | -2.07 | 0.015 |
| 30 | 20 | 291.66 | 586.42 | 1.73 | -1.43 | 0.015 |
| 31 | 20 | 298.83 | 591.48 | 1.52 | -1.22 | 0.015 |
| 32 | 20 | 308.33 | 594.99 | 1.50 | -1.20 | 0.015 |
| 43 | 20 | 321.25 | 592.33 | 1.50 | -1.20 | 0.015 |
|  | 20 | 980.00 | 493.03 | 0.50 | -7.20 | 0.015 |
| 20.30 | 0.00 | 0.035 |  |  |  |  |
| 23 |  |  |  |  |  |  |

Appendix B

| 4 | 21 | 163.48 | 636.97 | 4.00 | -3.70 | 0.015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 5 | 21 | 777.12 | 636.97 | 0.30 | 0.00 | 0.035 |
| 18 | 21 | 543.88 | 559.19 | 0.30 | 0.00 | 0.035 |
| 19 | 21 | 924.48 | 545.94 | 0.30 | 0.00 | 0.035 |
| 21 | 21 | 238.38 | 558.01 | 1.50 | -1.20 | 0.015 |
| 22 | 21 | 226.86 | 564.47 | 1.50 | -1.20 | 0.015 |
| 23 | 21 | 214.85 | 568.55 | 5.19 | -4.89 | 0.015 |
| 24 | 21 | 161.25 | 588.38 | 12.50 | -12.20 | 0.015 |
| 25 | 21 | 279.99 | 595.49 | 6.85 | -6.55 | 0.015 |
| 26 | 21 | 274.83 | 580.73 | 3.94 | -3.64 | 0.015 |
| 27 | 21 | 272.02 | 577.38 | 2.00 | -1.70 | 0.015 |
| 28 | 21 | 270.92 | 579.52 | 2.75 | -2.45 | 0.015 |
| 29 | 21 | 271.63 | 585.15 | 2.84 | -2.54 | 0.015 |
| 30 | 21 | 274.38 | 594.03 | 1.50 | -1.20 | 0.015 |
| 31 | 21 | 279.21 | 607.06 | 1.50 | -1.20 | 0.015 |
| 32 | 21 | 285.25 | 625.55 | 1.50 | -1.20 | 0.015 |
| 33 | 21 | 307.14 | 640.88 | 1.50 | -1.20 | 0.015 |
| 43 | 21 | 159.14 | 686.53 | 6.50 | -6.20 | 0.015 |
| 44 | 21 | 1200.00 | 686.53 | 0.30 | 0.00 | 0.035 |
| 4 | 22 | 125.00 | 548.29 | 4.00 | -3.70 | 0.015 |
| 5 | 22 | 902.80 | 548.29 | 0.30 | 0.00 | 0.035 |
| 17 | 22 | 482.46 | 505.75 | 0.30 | 0.00 | 0.035 |
| 18 | 22 | 482.46 | 505.75 | 0.30 | 0.00 | 0.035 |
| 20 | 22 | 482.54 | 516.75 | 0.30 | 0.00 | 0.035 |
| 21 | 22 | 235.11 | 523.47 | 2.04 | -1.74 | 0.015 |
| 22 | 22 | 227.95 | 530.67 | 1.50 | -1.20 | 0.015 |
| 23 | 22 | 218.67 | 534.34 | 3.13 | -2.83 | 0.015 |
| 24 | 22 | 142.84 | 534.77 | 12.50 | -12.20 | 0.015 |
| 25 | 22 | 287.61 | 540.45 | 9.28 | -8.98 | 0.015 |
| 26 | 22 | 268.32 | 552.92 | 5.09 | -4.79 | 0.015 |
| 27 | 22 | 259.86 | 565.91 | 2.45 | -2.15 | 0.015 |
| 28 | 22 | 255.71 | 577.68 | 2.46 | -2.16 | 0.015 |
| 29 | 22 | 253.83 | 587.67 | 2.34 | -2.04 | 0.015 |
| 30 | 22 | 254.30 | 596.42 | 1.56 | -1.26 | 0.015 |
| 31 | 22 | 259.21 | 606.91 | 1.50 | -1.20 | 0.015 |
| 32 | 22 | 267.68 | 615.01 | 1.50 | -1.20 | 0.015 |
| 43 | 22 | 133.14 | 474.05 | 5.50 | -5.20 | 0.015 |
| 44 | 22 | 900.00 | 474.05 | 0.30 | 0.00 | 0.035 |
| 4 | 23 | 145.77 | 637.38 | 4.00 | -3.70 | 0.015 |
| 5 | 23 | 776.62 | 637.38 | 0.30 | 0.00 | 0.035 |
| 17 | 23 | 485.30 | 419.43 | 0.30 | 0.00 | 0.035 |
| 21 | 23 | 333.07 | 467.21 | 1.50 | -1.20 | 0.015 |
| 22 | 23 | 236.64 | 486.00 | 1.50 | -1.20 | 0.015 |
| 23 | 23 | 250.57 | 515.23 | 1.96 | -1.66 | 0.015 |
| 24 | 23 | 142.84 | 534.77 | 12.50 | -12.20 | 0.015 |
| 25 | 23 | 254.17 | 543.71 | 10.15 | -9.85 | 0.015 |
| 26 | 23 | 248.18 | 565.59 | 5.56 | -5.26 | 0.015 |
| 27 | 23 | 245.71 | 589.65 | 2.42 | -2.12 | 0.015 |
| 28 | 23 | 243.18 | 611.71 | 2.46 | -2.16 | 0.015 |
| 29 | 23 | 238.70 | 629.28 | 1.90 | -1.60 | 0.015 |
| 30 | 23 | 228.74 | 627.96 | 1.50 | -1.20 | 0.015 |
| 31 | 23 | 236.34 | 639.74 | 1.50 | -1.20 | 0.015 |
| 42 | 23 | 290.26 | 652.64 | 1.50 | -1.20 | 0.015 |
| 44 | 23 | 160.08 | 656.22 | 4.50 | -4.20 | 0.015 |
|  |  | 1300.00 | 656.22 | 0.30 | 0.00 | 0.035 |
| 23 |  |  |  |  |  | 0 |


| 4 | 24 | 175.00 | 654.31 | 4.00 | -3.70 | 0.015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 5 | 24 | 1000.00 | 654.31 | 0.30 | 0.00 | 0.035 |
| 17 | 24 | 552.30 | 324.88 | 0.30 | 0.00 | 0.035 |
| 21 | 24 | 329.94 | 434.94 | 1.50 | -1.20 | 0.015 |
| 23 | 24 | 255.88 | 468.59 | 1.56 | -1.26 | 0.015 |
| 24 | 24 | 160.70 | 490.37 | 12.50 | -12.20 | 0.015 |
| 25 | 24 | 216.28 | 532.57 | 11.08 | -10.78 | 0.015 |
| 26 | 24 | 229.52 | 582.31 | 5.39 | -5.09 | 0.015 |
| 27 | 24 | 238.30 | 621.92 | 3.02 | -2.72 | 0.015 |
| 28 | 24 | 243.31 | 654.04 | 2.01 | -1.71 | 0.015 |
| 29 | 24 | 244.31 | 677.03 | 1.50 | -1.20 | 0.015 |
| 43 | 24 | 201.56 | 638.85 | 4.00 | -3.70 | 0.015 |
| 44 | 24 | 1200.00 | 638.85 | 0.30 | 0.00 | 0.035 |
| 4 | 25 | 206.16 | 667.08 | 4.00 | -3.70 | 0.015 |
| 5 | 25 | 1200.00 | 667.08 | 0.30 | 0.00 | 0.035 |
| 17 | 25 | 546.10 | 363.99 | 0.30 | 0.00 | 0.035 |
| 21 | 25 | 228.08 | 481.03 | 1.50 | -1.20 | 0.015 |
| 23 | 25 | 220.85 | 548.87 | 2.20 | -1.90 | 0.015 |
| 24 | 25 | 146.29 | 578.54 | 12.50 | -12.20 | 0.015 |
| 25 | 25 | 218.41 | 602.23 | 10.98 | -10.68 | 0.015 |
| 26 | 25 | 229.55 | 626.21 | 5.65 | -5.35 | 0.015 |
| 27 | 25 | 241.98 | 657.39 | 2.68 | -2.38 | 0.015 |
| 28 | 25 | 253.89 | 694.36 | 1.65 | -1.35 | 0.015 |
| 29 | 25 | 267.85 | 730.59 | 1.74 | -1.44 | 0.015 |
| 30 | 25 | 237.07 | 769.45 | 1.89 | -1.59 | 0.015 |
| 31 | 25 | 280.18 | 793.17 | 1.50 | -1.20 | 0.015 |
| 43 | 25 | 237.17 | 617.45 | 4.00 | -3.70 | 0.015 |
| 44 | 25 | 801.68 | 617.45 | 0.30 | 0.00 | 0.035 |
| 4 | 26 | 201.56 | 627.00 | 4.00 | -3.70 | 0.015 |
| 5 | 26 | 1200.00 | 627.00 | 0.30 | 0.00 | 0.035 |
| 17 | 26 | 800.00 | 442.85 | 0.30 | 0.00 | 0.035 |
| 21 | 26 | 231.49 | 534.05 | 1.50 | -1.20 | 0.015 |
| 23 | 26 | 211.17 | 596.25 | 4.28 | -3.98 | 0.015 |
| 24 | 26 | 112.25 | 611.44 | 12.50 | -12.20 | 0.015 |
| 25 | 26 | 214.55 | 618.01 | 12.58 | -12.28 | 0.015 |
| 26 | 26 | 229.21 | 631.30 | 3.87 | -3.57 | 0.015 |
| 27 | 26 | 249.96 | 640.47 | 1.90 | -1.60 | 0.015 |
| 28 | 26 | 278.35 | 639.12 | 1.50 | -1.20 | 0.015 |
| 43 | 26 | 227.71 | 613.88 | 4.00 | -3.70 | 0.015 |
| 44 | 26 | 1200.00 | 613.88 | 0.30 | 0.00 | 0.035 |
| 4 | 27 | 180.28 | 631.59 | 4.00 | -3.70 | 0.015 |
| 5 | 27 | 1200.00 | 631.59 | 0.30 | 0.00 | 0.035 |
| 18 | 27 | 907.14 | 457.00 | 0.30 | 0.00 | 0.035 |
| 19 | 27 | 504.02 | 485.41 | 0.30 | 0.00 | 0.035 |
| 20 | 27 | 497.56 | 513.00 | 0.30 | 0.00 | 0.035 |
| 21 | 27 | 244.65 | 543.95 | 1.50 | -1.20 | 0.015 |
| 22 | 27 | 221.05 | 564.90 | 1.77 | -1.47 | 0.015 |
| 23 | 27 | 210.12 | 573.31 | 7.07 | -6.77 | 0.015 |
| 24 | 27 | 127.57 | 608.89 | 12.50 | -12.20 | 0.015 |
| 25 | 27 | 196.57 | 649.51 | 10.84 | -10.54 | 0.015 |
| 26 | 27 | 227.48 | 664.59 | 2.33 | -2.03 | 0.015 |
| 27 | 27 | 260.00 | 677.77 | 1.66 | -1.36 | 0.015 |
| 28 | 27 | 300.82 | 699.04 | 1.50 | -1.20 | 0.015 |
| 29 | 27 | 369.52 | 732.80 | 1.50 | -1.20 | 0.015 |
| 43 | 27 | 196.47 | 693.18 | 4.00 | -3.70 | 0.015 |
|  |  |  |  |  |  |  |


| 44 | 27 | 1200.00 | 693.18 | 0.30 | 0.00 | 0.035 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 28 | 180.28 | 758.22 | 4.00 | -3.70 | 0.015 |
| 5 | 28 | 1300.00 | 758.22 | 0.30 | 0.00 | 0.035 |
| 22 | 28 | 220.33 | 556.60 | 2.87 | -2.57 | 0.015 |
| 23 | 28 | 201.66 | 595.76 | 9.07 | -8.77 | 0.015 |
| 24 | 28 | 139.75 | 587.33 | 12.50 | -12.20 | 0.015 |
| 25 | 28 | 236.87 | 573.35 | 9.32 | -9.02 | 0.015 |
| 26 | 28 | 258.18 | 598.68 | 1.50 | -1.20 | 0.015 |
| 27 | 28 | 287.31 | 619.34 | 1.50 | -1.20 | 0.015 |
| 28 | 28 | 326.60 | 638.21 | 1.50 | -1.20 | 0.015 |
| 29 | 28 | 382.93 | 655.65 | 1.50 | -1.20 | 0.015 |
| 30 | 28 | 537.69 | 682.65 | 1.50 | -1.20 | 0.015 |
| 43 | 28 | 222.99 | 690.015 | 4.00 | -3.70 | 0.015 |
| 44 | 28 | 1200.00 | 690.015 | 0.30 | 0.00 | 0.035 |
| 4 | 29 | 125.00 | 703.58 | 4.00 | -3.70 | 0.015 |
| 5 | 29 | 1300.00 | 703.58 | 0.30 | 0.00 | 0.035 |
| 21 | 29 | 237.07 | 616.81 | 2.51 | -2.21 | 0.015 |
| 22 | 29 | 222.71 | 584.61 | 6.02 | -5.72 | 0.015 |
| 23 | 29 | 212.54 | 566.87 | 8.25 | -7.95 | 0.015 |
| 24 | 29 | 134.63 | 560.24 | 12.50 | -12.20 | 0.015 |
| 25 | 29 | 300.91 | 558.19 | 7.62 | -7.32 | 0.015 |
| 26 | 29 | 301.39 | 561.70 | 1.50 | -1.20 | 0.015 |
| 27 | 29 | 320.51 | 573.01 | 1.50 | -1.20 | 0.015 |
| 28 | 29 | 355.82 | 585.93 | 1.50 | -1.20 | 0.015 |
| 29 | 29 | 412.04 | 596.97 | 1.50 | -1.20 | 0.015 |
| 30 | 29 | 495.77 | 597.43 | 1.50 | -1.20 | 0.015 |
| 31 | 29 | 364.42 | 592.89 | 1.50 | -1.20 | 0.015 |
| 43 | 29 | 276.09 | 789.45 | 4.00 | -3.70 | 0.015 |
| 44 | 29 | 1300.00 | 789.45 | 0.30 | 0.00 | 0.035 |
| 4 | 30 | 191.64 | 580.28 | 4.00 | -3.70 | 0.015 |
| 5 | 30 | 1100.00 | 580.28 | 0.30 | 0.00 | 0.035 |
| 20 | 30 | 227.49 | 531.80 | 5.20 | -4.90 | 0.015 |
| 21 | 30 | 225.31 | 519.01 | 7.46 | -7.16 | 0.015 |
| 22 | 30 | 221.45 | 535.00 | 2.64 | -2.34 | 0.015 |
| 23 | 30 | 217.91 | 522.32 | 3.81 | -3.51 | 0.015 |
| 24 | 30 | 152.07 | 538.11 | 12.50 | -12.20 | 0.015 |
| 25 | 30 | 322.64 | 570.13 | 6.90 | -6.60 | 0.015 |
| 26 | 30 | 325.67 | 572.98 | 1.50 | -1.20 | 0.015 |
| 27 | 30 | 340.06 | 571.48 | 1.50 | -1.20 | 0.015 |
| 28 | 30 | 366.83 | 569.34 | 1.50 | -1.20 | 0.015 |
| 29 | 30 | 412.52 | 563.71 | 1.50 | -1.20 | 0.015 |
| 30 | 30 | 499.28 | 538.03 | 1.50 | -1.20 | 0.015 |
| 43 | 30 | 302.08 | 575.54 | 4.00 | -3.70 | 0.015 |
| 44 | 30 | 1100.00 | 575.54 | 0.30 | 0.00 | 0.035 |
| 4 | 31 | 176.71 | 666.13 | 4.00 | -3.70 | 0.015 |
| 5 | 31 | 1300.00 | 666.13 | 0.30 | 0.00 | 0.035 |
| 20 | 31 | 235.34 | 770.17 | 8.26 | -7.96 | 0.015 |
| 21 | 31 | 231.35 | 726.78 | 7.02 | -6.72 | 0.015 |
| 22 | 31 | 225.06 | 711.81 | 1.75 | -1.45 | 0.015 |
| 23 | 31 | 215.77 | 720.88 | 2.87 | -2.57 | 0.015 |
| 24 | 31 | 162.50 | 728.36 | 12.50 | -12.20 | 0.015 |
| 25 | 31 | 325.08 | 685.99 | 5.92 | -5.62 | 0.015 |
| 26 | 31 | 335.30 | 625.33 | 1.50 | -1.20 | 0.015 |
| 27 | 31 | 335.30 | 625.33 | 1.50 | -1.20 | 0.015 |
| 28 | 31 | 366.85 | 584.47 | 1.50 | -1.20 | 0.015 |


| 29 | 31 | 397.56 | 590.72 | 1.50 | -1.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 30 | 31 | 510.41 | 612.57 | 1.50 | -1.20 | 0.015 |
| 43 | 31 | 214.71 | 733.55 | 4.00 | -3.70 | 0.015 |
| 44 | 31 | 1400.00 | 733.55 | 0.30 | 0.00 | 0.035 |
| 4 | 32 | 90.14 | 548.29 | 4.00 | -3.70 | 0.015 |
| 5 | 32 | 1080.00 | 548.29 | 0.30 | 0.00 | 0.035 |
| 20 | 32 | 235.44 | 611.75 | 6.75 | -6.45 | 0.015 |
| 21 | 32 | 228.95 | 601.48 | 6.05 | -5.75 | 0.015 |
| 22 | 32 | 223.75 | 592.29 | 4.71 | -4.41 | 0.015 |
| 23 | 32 | 219.85 | 583.35 | 5.29 | -4.99 | 0.015 |
| 24 | 32 | 152.15 | 554.19 | 12.50 | -12.20 | 0.015 |
| 25 | 32 | 331.01 | 536.81 | 5.62 | -5.32 | 0.015 |
| 26 | 32 | 334.64 | 550.15 | 1.50 | -1.20 | 0.015 |
| 28 | 32 | 434.64 | 600.15 | 1.50 | -1.20 | 0.015 |
| 29 | 32 | 399.67 | 618.24 | 1.50 | -1.20 | 0.015 |
| 30 | 32 | 399.67 | 618.24 | 1.50 | -1.20 | 0.015 |
| 43 | 32 | 131.24 | 627.00 | 4.00 | -3.70 | 0.015 |
| 44 | 32 | 1200.00 | 627.00 | 0.30 | 0.00 | 0.035 |
| 4 | 33 | 176.78 | 637.38 | 4.00 | -3.70 | 0.015 |
| 5 | 33 | 1200.00 | 637.38 | 0.30 | 0.00 | 0.035 |
| 19 | 33 | 249.63 | 512.48 | 3.93 | -3.63 | 0.015 |
| 20 | 33 | 226.63 | 497.43 | 8.50 | -8.20 | 0.015 |
| 21 | 33 | 220.16 | 496.84 | 3.39 | -3.09 | 0.015 |
| 22 | 33 | 217.70 | 490.62 | 1.50 | -1.20 | 0.015 |
| 23 | 33 | 222.09 | 479.36 | 2.31 | -2.01 | 0.015 |
| 24 | 33 | 134.82 | 467.11 | 12.50 | -12.20 | 0.015 |
| 25 | 33 | 327.40 | 478.91 | 6.58 | -6.28 | 0.015 |
| 26 | 33 | 322.66 | 511.83 | 1.50 | -1.20 | 0.015 |
| 28 | 33 | 399.13 | 544.22 | 1.50 | -1.20 | 0.015 |
| 29 | 33 | 399.13 | 544.22 | 1.50 | -1.20 | 0.015 |
| 30 | 33 | 338.54 | 550.21 | 1.50 | -1.20 | 0.015 |
| 43 | 33 | 143.96 | 508.94 | 4.00 | -3.70 | 0.015 |
| 44 | 33 | 1000.00 | 508.94 | 0.30 | 0.00 | 0.035 |
| 4 | 34 | 106.07 | 686.04 | 4.00 | -3.70 | 0.015 |
| 5 | 34 | 1300.00 | 686.04 | 0.30 | 0.00 | 0.035 |
| 19 | 34 | 255.29 | 434.56 | 8.39 | -8.09 | 0.015 |
| 20 | 34 | 232.33 | 468.82 | 6.01 | -5.71 | 0.015 |
| 21 | 34 | 218.89 | 489.13 | 1.50 | -1.20 | 0.015 |
| 22 | 34 | 211.57 | 502.76 | 1.50 | -1.20 | 0.015 |
| 23 | 34 | 210.35 | 512.93 | 2.55 | -2.25 | 0.015 |
| 24 | 34 | 123.31 | 519.82 | 12.50 | -12.20 | 0.015 |
| 25 | 34 | 279.83 | 533.34 | 7.64 | -7.34 | 0.015 |
| 26 | 34 | 296.11 | 568.27 | 1.50 | -1.20 | 0.015 |
| 28 | 34 | 459.92 | 646.01 | 1.50 | -1.20 | 0.015 |
| 29 | 34 | 401.90 | 617.07 | 1.50 | -1.20 | 0.015 |
| 30 | 34 | 404.01 | 618.20 | 1.50 | -1.20 | 0.015 |
| 43 | 34 | 200.06 | 752.08 | 4.00 | -3.70 | 0.015 |
| 44 | 34 | 1400.00 | 752.08 | 0.30 | 0.00 | 0.035 |
| 4 | 35 | 158.11 | 707.74 | 4.00 | -3.70 | 0.015 |
| 5 | 35 | 1400.00 | 707.74 | 0.30 | 0.00 | 0.035 |
| 19 | 35 | 280.38 | 507.76 | 6.02 | -5.72 | 0.015 |
| 20 | 35 | 237.79 | 518.99 | 2.58 | -2.28 | 0.015 |
| 21 | 35 | 214.96 | 549.70 | 1.50 | -1.20 | 0.015 |
| 22 | 35 | 198.12 | 589.87 | 1.50 | -1.20 | 0.015 |
| 23 | 35 | 182.35 | 636.76 | 3.34 | -3.04 | 0.015 |

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| 24 | 35 | 134.54 | 661.85 | 12.50 | -12.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25 | 35 | 186.65 | 662.57 | 12.50 | -12.20 | 0.015 |
| 26 | 35 | 244.99 | 662.91 | 3.73 | -3.43 | 0.015 |
| 27 | 35 | 403.62 | 643.10 | 1.50 | -1.20 | 0.015 |
| 28 | 35 | 459.92 | 646.01 | 1.50 | -1.20 | 0.015 |
| 29 | 35 | 434.88 | 689.59 | 1.50 | -1.20 | 0.015 |
| 30 | 35 | 405.87 | 741.94 | 1.50 | -1.20 | 0.015 |
| 43 | 35 | 154.03 | 642.83 | 4.00 | -3.70 | 0.015 |
| 44 | 35 | 1200.00 | 642.83 | 0.30 | 0.00 | 0.035 |
| 4 | 36 | 182.00 | 594.24 | 4.00 | -3.70 | 0.015 |
| 19 | 36 | 271.75 | 432.19 | 3.24 | -2.94 | 0.015 |
| 20 | 36 | 233.35 | 466.32 | 4.99 | -4.69 | 0.015 |
| 21 | 36 | 211.96 | 484.98 | 1.50 | -1.20 | 0.015 |
| 22 | 36 | 198.55 | 493.62 | 1.79 | -1.49 | 0.015 |
| 23 | 36 | 191.37 | 490.80 | 4.34 | -4.04 | 0.015 |
| 24 | 36 | 127.57 | 512.62 | 12.50 | -12.20 | 0.015 |
| 25 | 36 | 158.41 | 553.88 | 9.76 | -9.46 | 0.015 |
| 26 | 36 | 200.015 | 580.32 | 4.10 | -3.80 | 0.015 |
| 27 | 36 | 476.04 | 637.67 | 1.50 | -1.20 | 0.015 |
| 28 | 36 | 486.04 | 637.67 | 1.50 | -1.20 | 0.015 |
| 29 | 36 | 478.18 | 665.86 | 1.50 | -1.20 | 0.015 |
| 30 | 36 | 398.21 | 716.00 | 1.50 | -1.20 | 0.015 |
| 43 | 36 | 164.92 | 626.26 | 4.00 | -3.70 | 0.015 |
| 44 | 36 | 1200.00 | 626.26 | 0.30 | 0.00 | 0.035 |
| 4 | 37 | 150.00 | 500.00 | 4.00 | -3.70 | 0.015 |
| 5 | 37 | 750.42 | 375.00 | 4.00 | -3.70 | 0.015 |
| 6 | 37 | 627.00 | 201.56 | 4.00 | -3.70 | 0.015 |
| 7 | 37 | 400.00 | 200.00 | 4.00 | -3.70 | 0.015 |
| 20 | 37 | 227.35 | 480.86 | 10.46 | -10.16 | 0.015 |
| 21 | 37 | 214.29 | 487.70 | 8.12 | -7.82 | 0.015 |
| 22 | 37 | 207.59 | 495.04 | 4.70 | -4.40 | 0.015 |
| 23 | 37 | 209.05 | 500.69 | 5.79 | -5.49 | 0.015 |
| 24 | 37 | 126.00 | 504.00 | 12.50 | -12.20 | 0.015 |
| 25 | 37 | 194.49 | 505.31 | 6.39 | -6.09 | 0.015 |
| 26 | 37 | 177.01 | 506.07 | 3.08 | -2.78 | 0.015 |
| 27 | 37 | 151.51 | 520.49 | 1.50 | -1.20 | 0.015 |
| 28 | 37 | 269.56 | 510.51 | 1.50 | -1.20 | 0.015 |
| 30 | 37 | 287.10 | 513.43 | 1.50 | -1.20 | 0.015 |
| 31 | 37 | 458.65 | 517.80 | 1.50 | -1.20 | 0.015 |
| 43 | 37 | 217.83 | 647.65 | 4.00 | -3.70 | 0.015 |
| 44 | 37 | 1200.00 | 647.65 | 0.30 | 0.00 | 0.035 |
| 4 | 38 | 152.07 | 682.37 | 4.00 | -3.70 | 0.015 |
| 7 | 38 | 254.95 | 471.70 | 4.00 | -3.70 | 0.015 |
| 8 | 38 | 1399.20 | 471.70 | 0.30 | 0.00 | 0.035 |
| 21 | 38 | 209.45 | 602.09 | 2.96 | -2.66 | 0.015 |
| 22 | 38 | 205.82 | 591.46 | 4.32 | -4.02 | 0.015 |
| 23 | 38 | 203.09 | 582.59 | 5.79 | -5.49 | 0.015 |
| 24 | 38 | 126.00 | 581.00 | 12.50 | -12.20 | 0.015 |
| 25 | 38 | 199.45 | 582.66 | 6.18 | -5.88 | 0.015 |
| 26 | 38 | 178.19 | 587.13 | 2.68 | -2.38 | 0.015 |
| 27 | 38 | 151.51 | 590.49 | 1.50 | -1.20 | 0.015 |
| 28 | 38 | 269.56 | 587.51 | 1.50 | -1.20 | 0.015 |
| 29 | 38 | 262.32 | 582.20 | 1.50 | -1.20 | 0.015 |
| 30 | 38 | 334.06 | 587.41 | 1.50 | -1.20 | 0.015 |
| 31 | 38 | 396.13 | 598.20 | 1.50 | -1.20 | 0.015 |

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| 32 | 38 | 328.00 | 605.56 | 1.50 | -1.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 43 | 38 | 195.26 | 811.25 | 4.00 | -3.70 | 0.015 |
| 44 | 38 | 1400.00 | 811.25 | 0.30 | 0.00 | 0.035 |
| 4 | 39 | 155.24 | 869.60 | 4.00 | -3.70 | 0.015 |
| 5 | 39 | 999.00 | 869.60 | 0.30 | 0.00 | 0.035 |
| 7 | 39 | 151.33 | 615.55 | 4.00 | -3.70 | 0.015 |
| 8 | 39 | 1072.22 | 615.55 | 0.30 | 0.00 | 0.035 |
| 22 | 39 | 202.11 | 562.70 | 1.50 | -1.20 | 0.015 |
| 23 | 39 | 201.60 | 575.19 | 3.88 | -3.58 | 0.015 |
| 24 | 39 | 126.00 | 580.00 | 12.50 | -12.20 | 0.015 |
| 25 | 39 | 184.55 | 578.93 | 7.32 | -7.02 | 0.015 |
| 26 | 39 | 192.10 | 577.31 | 3.26 | -2.96 | 0.015 |
| 27 | 39 | 205.95 | 579.39 | 1.50 | -1.20 | 0.015 |
| 28 | 39 | 234.43 | 585.48 | 1.50 | -1.20 | 0.015 |
| 29 | 39 | 268.40 | 588.73 | 1.50 | -1.20 | 0.015 |
| 30 | 39 | 317.31 | 584.27 | 1.50 | -1.20 | 0.015 |
| 31 | 39 | 388.41 | 567.45 | 1.50 | -1.20 | 0.015 |
| 32 | 39 | 328.00 | 605.56 | 1.50 | -1.20 | 0.015 |
| 43 | 39 | 145.77 | 617.45 | 4.00 | -3.70 | 0.015 |
| 44 | 39 | 1068.92 | 617.45 | 0.30 | 0.00 | 0.035 |
| 4 | 40 | 91.24 | 723.30 | 4.00 | -3.70 | 0.015 |
| 5 | 40 | 999.00 | 723.30 | 0.30 | 0.00 | 0.035 |
| 7 | 40 | 202.24 | 637.89 | 4.00 | -3.70 | 0.015 |
| 8 | 40 | 1200.00 | 637.89 | 0.30 | 0.00 | 0.035 |
| 23 | 40 | 201.50 | 605.00 | 4.29 | -3.99 | 0.015 |
| 24 | 40 | 126.50 | 605.00 | 12.50 | -12.20 | 0.015 |
| 25 | 40 | 179.16 | 608.68 | 6.49 | -6.19 | 0.015 |
| 26 | 40 | 194.24 | 612.52 | 1.61 | -1.31 | 0.015 |
| 27 | 40 | 211.95 | 610.54 | 1.59 | -1.29 | 0.015 |
| 28 | 40 | 234.60 | 604.35 | 2.32 | -2.02 | 0.015 |
| 29 | 40 | 265.08 | 593.56 | 1.71 | -1.41 | 0.015 |
| 30 | 40 | 311.23 | 575.37 | 1.50 | -1.20 | 0.015 |
| 31 | 40 | 396.48 | 534.45 | 1.50 | -1.20 | 0.015 |
| 43 | 40 | 168.00 | 592.22 | 4.00 | -3.70 | 0.015 |
| 44 | 40 | 1114.46 | 592.22 | 0.30 | 0.00 | 0.035 |
| 4 | 41 | 102.96 | 652.76 | 4.00 | -3.70 | 0.015 |
| 5 | 41 | 900.00 | 652.76 | 0.30 | 0.00 | 0.035 |
| 7 | 41 | 125.00 | 450.69 | 4.00 | -3.70 | 0.015 |
| 8 | 41 | 732.21 | 450.69 | 0.30 | 0.00 | 0.035 |
| 23 | 41 | 201.50 | 707.00 | 7.60 | -7.30 | 0.015 |
| 24 | 41 | 114.00 | 707.12 | 12.50 | -12.20 | 0.015 |
| 25 | 41 | 192.42 | 694.87 | 4.82 | -4.52 | 0.015 |
| 26 | 41 | 206.01 | 670.20 | 1.50 | -1.20 | 0.015 |
| 28 | 41 | 239.39 | 622.55 | 1.50 | -1.20 | 0.015 |
| 29 | 41 | 264.90 | 601.93 | 2.61 | -2.31 | 0.015 |
| 30 | 41 | 310.28 | 591.34 | 1.50 | -1.20 | 0.015 |
| 31 | 41 | 430.95 | 576.15 | 1.66 | -1.36 | 0.015 |
| 43 | 41 | 168.00 | 733.64 | 4.00 | -3.70 | 0.015 |
| 44 | 41 | 899.62 | 733.64 | 0.30 | 0.00 | 0.035 |
| 4 | 42 | 127.48 | 571.18 | 4.00 | -3.70 | 0.015 |
| 5 | 42 | 750.00 | 571.18 | 0.30 | 0.00 | 0.035 |
| 7 | 42 | 134.63 | 536.77 | 4.00 | -3.70 | 0.015 |
| 8 | 42 | 614.79 | 536.77 | 0.30 | 0.00 | 0.035 |
| 22 | 42 | 278.08 | 735.33 | 2.40 | -2.10 | 0.015 |
| 23 | 42 | 189.50 | 707.00 | 7.26 | -6.96 | 0.015 |


| 24 | 42 | 116.22 | 682.03 | 12.52 | -12.22 | 0.015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 25 | 42 | 224.51 | 647.62 | 3.31 | -3.01 | 0.015 |
| 26 | 42 | 228.81 | 627.93 | 1.50 | -1.20 | 0.015 |
| 28 | 42 | 248.37 | 588.73 | 1.50 | -1.20 | 0.015 |
| 29 | 42 | 266.94 | 566.00 | 2.54 | -2.24 | 0.015 |
| 30 | 42 | 300.32 | 534.15 | 1.50 | -1.20 | 0.015 |
| 43 | 42 | 134.63 | 770.15 | 4.00 | -3.70 | 0.015 |
| 44 | 42 | 856.98 | 770.15 | 0.30 | 0.00 | 0.035 |
| 4 | 43 | 130.00 | 715.89 | 4.00 | -3.70 | 0.015 |
| 5 | 43 | 999.00 | 715.89 | 0.30 | 0.00 | 0.035 |
| 7 | 43 | 125.00 | 419.08 | 4.00 | -3.70 | 0.015 |
| 8 | 43 | 787.44 | 419.08 | 0.30 | 0.00 | 0.035 |
| 23 | 43 | 189.50 | 616.05 | 5.46 | -5.16 | 0.015 |
| 24 | 43 | 124.23 | 616.05 | 12.50 | -12.20 | 0.015 |
| 25 | 43 | 248.72 | 606.75 | 6.05 | -5.75 | 0.015 |
| 26 | 43 | 245.29 | 595.90 | 1.50 | -1.20 | 0.015 |
| 27 | 43 | 247.72 | 588.13 | 1.50 | -1.20 | 0.015 |
| 28 | 43 | 255.01 | 581.60 | 1.89 | -1.59 | 0.015 |
| 29 | 43 | 267.24 | 575.40 | 2.11 | -1.81 | 0.015 |
| 30 | 43 | 284.14 | 568.12 | 1.50 | -1.20 | 0.015 |
| 31 | 43 | 295.31 | 569.01 | 2.05 | -1.75 | 0.015 |
| 43 | 43 | 150.33 | 741.69 | 4.00 | -3.70 | 0.015 |
| 44 | 43 | 889.86 | 741.69 | 0.30 | 0.00 | 0.035 |
| 4 | 44 | 183.44 | 649.04 | 4.00 | -3.70 | 0.015 |
| 5 | 44 | 900.00 | 649.04 | 0.30 | 0.00 | 0.035 |
| 7 | 44 | 201.56 | 340.035 | 4.00 | -3.70 | 0.015 |
| 8 | 44 | 970.47 | 340.035 | 0.30 | 0.00 | 0.035 |
| 23 | 44 | 183.45 | 569.01 | 2.93 | -2.63 | 0.015 |
| 24 | 44 | 126.62 | 556.71 | 12.50 | -12.20 | 0.015 |
| 25 | 44 | 251.81 | 600.98 | 5.83 | -5.53 | 0.015 |
| 26 | 44 | 251.29 | 602.14 | 1.50 | -1.20 | 0.015 |
| 27 | 44 | 253.29 | 602.81 | 1.50 | -1.20 | 0.015 |
| 23 | 45 | 45 | 193.07 | 682.91 | 4.29 | -3.99 |


| 44 | 45 | 998.14 | 661.23 | 0.30 | 0.00 | 0.035 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 4 | 46 | 131.53 | 634.11 | 4.00 | -3.70 | 0.015 |
| 5 | 46 | 900.00 | 634.11 | 0.30 | 0.00 | 0.035 |
| 7 | 46 | 145.77 | 382.43 | 4.00 | -3.70 | 0.015 |
| 8 | 46 | 862.90 | 382.43 | 0.30 | 0.00 | 0.035 |
| 22 | 46 | 290.60 | 624.83 | 1.50 | -1.20 | 0.015 |
| 23 | 46 | 250.88 | 591.85 | 3.24 | -2.94 | 0.015 |
| 24 | 46 | 101.83 | 611.92 | 12.50 | -12.20 | 0.015 |
| 25 | 46 | 268.16 | 659.53 | 7.89 | -7.59 | 0.015 |
| 26 | 46 | 255.09 | 683.69 | 2.72 | -2.42 | 0.015 |
| 27 | 46 | 248.08 | 710.82 | 3.29 | -2.99 | 0.015 |
| 28 | 46 | 225.96 | 719.10 | 1.50 | -1.20 | 0.015 |
| 43 | 46 | 201.56 | 665.68 | 4.00 | -3.70 | 0.015 |
| 44 | 46 | 991.46 | 665.68 | 0.30 | 0.00 | 0.035 |
| 4 | 47 | 122.58 | 677.66 | 4.00 | -3.70 | 0.015 |
| 5 | 47 | 900.00 | 677.66 | 0.30 | 0.00 | 0.035 |
| 7 | 47 | 230.49 | 654.31 | 4.00 | -3.70 | 0.015 |
| 8 | 47 | 1008.70 | 654.31 | 0.30 | 0.00 | 0.035 |
| 22 | 47 | 286.12 | 586.77 | 1.50 | -1.20 | 0.015 |
| 23 | 47 | 275.88 | 649.52 | 2.20 | -1.90 | 0.015 |
| 24 | 47 | 116.22 | 683.86 | 12.50 | -12.20 | 0.015 |
| 25 | 47 | 236.80 | 697.98 | 9.06 | -8.76 | 0.015 |
| 26 | 47 | 240.86 | 723.27 | 3.57 | -3.27 | 0.015 |
| 27 | 47 | 241.82 | 746.05 | 2.00 | -1.70 | 0.015 |
| 43 | 47 | 111.80 | 538.52 | 4.00 | -3.70 | 0.015 |
| 44 | 47 | 612.79 | 538.52 | 0.30 | 0.00 | 0.035 |
| 4 | 48 | 133.14 | 719.81 | 4.00 | -3.70 | 0.015 |
| 5 | 48 | 900.00 | 719.81 | 0.30 | 0.00 | 0.035 |
| 7 | 48 | 195.26 | 529.74 | 4.00 | -3.70 | 0.015 |
| 8 | 48 | 622.95 | 529.74 | 0.30 | 0.00 | 0.035 |
| 21 | 48 | 303.16 | 601.49 | 1.50 | -1.20 | 0.015 |
| 22 | 48 | 287.20 | 642.65 | 1.50 | -1.20 | 0.015 |
| 23 | 48 | 267.57 | 687.69 | 1.81 | -1.51 | 0.015 |
| 24 | 48 | 114.19 | 698.39 | 12.50 | -12.20 | 0.015 |
| 25 | 48 | 238.08 | 691.62 | 8.72 | -8.42 | 0.015 |
| 26 | 48 | 239.88 | 701.04 | 2.10 | -1.80 | 0.015 |
| 27 | 48 | 241.58 | 710.49 | 1.80 | -1.50 | 0.015 |
| 28 | 48 | 251.32 | 724.31 | 1.50 | -1.20 | 0.015 |
| 43 | 48 | 175.00 | 760.35 | 4.00 | -3.70 | 0.015 |
| 44 | 48 | 1200.00 | 760.35 | 0.30 | 0.00 | 0.035 |
| 4 | 49 | 166.51 | 691.47 | 4.00 | -3.70 | 0.015 |
| 5 | 49 | 900.00 | 691.47 | 0.30 | 0.00 | 0.035 |
| 23 | 49 | 103.08 | 477.62 | 4.00 | -3.70 | 0.015 |
| 24 | 49 | 690.93 | 477.62 | 0.30 | 0.00 | 0.035 |
| 27 | 49 | 49 | 291.88 | 671.42 | 1.50 | -1.20 |


| 5 | 50 | 900.00 | 678.12 | 0.30 | 0.00 | 0.035 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 50 | 160.08 | 427.93 | 4.00 | -3.70 | 0.015 |
| 8 | 50 | 771.15 | 427.93 | 0.30 | 0.00 | 0.035 |
| 21 | 50 | 321.20 | 721.48 | 1.50 | -1.20 | 0.015 |
| 22 | 50 | 302.75 | 713.41 | 1.50 | -1.20 | 0.015 |
| 23 | 50 | 289.30 | 714.29 | 1.70 | -1.40 | 0.015 |
| 24 | 50 | 126.00 | 717.64 | 12.50 | -12.20 | 0.015 |
| 25 | 50 | 285.19 | 713.74 | 6.35 | -6.05 | 0.015 |
| 26 | 50 | 250.09 | 712.00 | 1.50 | -1.20 | 0.015 |
| 27 | 50 | 222.49 | 715.63 | 1.50 | -1.20 | 0.015 |
| 28 | 50 | 222.49 | 715.63 | 1.50 | -1.20 | 0.015 |
| 43 | 50 | 127.48 | 480.88 | 4.00 | -3.70 | 0.015 |
| 44 | 50 | 686.24 | 500.00 | 0.30 | 0.00 | 0.035 |
| 4 | 51 | 167.63 | 659.24 | 4.00 | -3.70 | 0.015 |
| 5 | 51 | 900.00 | 678.12 | 0.30 | 0.00 | 0.035 |
| 7 | 51 | 127.48 | 395.28 | 4.00 | -3.70 | 0.015 |
| 8 | 51 | 834.85 | 395.28 | 0.30 | 0.00 | 0.035 |
| 21 | 51 | 340.15 | 764.48 | 1.50 | -1.20 | 0.015 |
| 22 | 51 | 304.35 | 721.37 | 1.50 | -1.20 | 0.015 |
| 23 | 51 | 276.82 | 683.32 | 2.45 | -2.15 | 0.015 |
| 24 | 51 | 138.50 | 663.05 | 12.50 | -12.20 | 0.015 |
| 25 | 51 | 300.43 | 656.66 | 5.25 | -4.95 | 0.015 |
| 27 | 51 | 242.78 | 620.65 | 1.50 | -1.20 | 0.015 |
| 28 | 51 | 242.78 | 620.65 | 1.50 | -1.20 | 0.015 |
| 43 | 51 | 141.42 | 509.90 | 4.00 | -3.70 | 0.015 |
| 44 | 51 | 970.78 | 509.90 | 0.30 | 0.00 | 0.035 |
| 4 | 52 | 175.00 | 550.015 | 4.00 | -3.70 | 0.015 |
| 5 | 52 | 600.00 | 550.015 | 0.30 | 0.00 | 0.035 |
| 7 | 52 | 167.71 | 648.56 | 4.00 | -3.70 | 0.015 |
| 8 | 52 | 1017.64 | 648.56 | 0.30 | 0.00 | 0.035 |
| 21 | 52 | 344.26 | 732.63 | 1.50 | -1.20 | 0.015 |
| 22 | 52 | 300.63 | 725.85 | 1.50 | -1.20 | 0.015 |
| 23 | 52 | 263.39 | 716.70 | 2.88 | -2.58 | 0.015 |
| 24 | 52 | 151.00 | 711.00 | 12.50 | -12.20 | 0.015 |
| 25 | 52 | 313.52 | 689.92 | 3.01 | -2.71 | 0.015 |
| 27 | 52 | 264.23 | 622.67 | 1.50 | -1.20 | 0.015 |
| 28 | 52 | 264.23 | 622.67 | 1.50 | -1.20 | 0.015 |
| 43 | 52 | 115.00 | 601.02 | 4.00 | -3.70 | 0.015 |
| 44 | 52 | 1100.00 | 601.02 | 0.30 | 0.00 | 0.035 |
| 4 | 53 | 191.64 | 659.34 | 4.00 | -3.70 | 0.015 |
| 5 | 53 | 900.00 | 659.34 | 0.30 | 0.00 | 0.035 |
| 7 | 53 | 145.77 | 728.87 | 4.00 | -3.70 | 0.015 |
| 8 | 53 | 905.52 | 728.87 | 0.30 | 0.00 | 0.035 |
| 20 | 53 | 737.40 | 618.99 | 0.30 | 0.00 | 0.035 |
| 21 | 53 | 339.51 | 639.88 | 2.65 | -2.35 | 0.015 |
| 22 | 53 | 304.13 | 665.46 | 1.50 | -1.20 | 0.015 |
| 23 | 53 | 275.04 | 696.38 | 1.78 | -1.48 | 0.015 |
| 24 | 53 | 151.56 | 699.79 | 12.50 | -12.20 | 0.015 |
| 25 | 53 | 324.27 | 693.28 | 4.56 | -4.26 | 0.015 |
| 26 | 53 | 282.14 | 712.66 | 1.50 | -1.20 | 0.015 |
| 27 | 53 | 252.88 | 744.61 | 1.50 | -1.20 | 0.015 |
| 28 | 53 | 252.88 | 744.61 | 1.50 | -1.20 | 0.015 |
| 43 | 53 | 99.25 | 578.66 | 4.00 | -3.70 | 0.015 |
| 44 | 53 | 1100.00 | 578.66 | 0.30 | 0.00 | 0.035 |
| 4 | 54 | 140.00 | 542.31 | 4.00 | -3.70 | 0.015 |


| 5 | 54 | 750.00 | 542.31 | 0.30 | 0.00 | 0.035 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 54 | 160.08 | 585.77 | 4.00 | -3.70 | 0.015 |
| 8 | 54 | 1126.72 | 585.77 | 0.30 | 0.00 | 0.035 |
| 20 | 54 | 742.66 | 409.16 | 0.30 | 0.00 | 0.035 |
| 21 | 54 | 742.66 | 409.16 | 0.30 | 0.00 | 0.035 |
| 24 | 54 | 126.00 | 584.53 | 12.50 | -12.20 | 0.015 |
| 25 | 54 | 316.45 | 637.57 | 3.83 | -3.53 | 0.015 |
| 26 | 54 | 290.22 | 674.67 | 1.50 | -1.20 | 0.015 |
| 27 | 54 | 275.62 | 672.76 | 1.50 | -1.20 | 0.015 |
| 43 | 54 | 111.80 | 570.09 | 4.00 | -3.70 | 0.015 |
| 44 | 54 | 1100.00 | 570.09 | 0.30 | 0.00 | 0.035 |
| 4 | 55 | 100.00 | 602.08 | 4.00 | -3.70 | 0.015 |
| 5 | 55 | 800.00 | 602.08 | 0.30 | 0.00 | 0.035 |
| 7 | 55 | 176.78 | 807.00 | 4.00 | -3.70 | 0.015 |
| 8 | 55 | 613.38 | 807.00 | 0.30 | 0.00 | 0.035 |
| 20 | 55 | 702.32 | 439.24 | 0.30 | 0.00 | 0.035 |
| 21 | 55 | 702.32 | 439.24 | 0.30 | 0.00 | 0.035 |
| 24 | 55 | 107.62 | 739.19 | 12.50 | -12.20 | 0.015 |
| 25 | 55 | 301.07 | 803.61 | 8.09 | -7.79 | 0.015 |
| 26 | 55 | 296.45 | 926.21 | 1.50 | -1.20 | 0.015 |
| 27 | 55 | 292.41 | 1095.30 | 1.50 | -1.20 | 0.015 |
| 43 | 55 | 100.00 | 471.70 | 4.00 | -3.70 | 0.015 |
| 44 | 55 | 524.70 | 500.00 | 0.30 | 0.00 | 0.035 |
| 4 | 56 | 87.46 | 728.87 | 4.00 | -3.70 | 0.015 |
| 5 | 56 | 900.00 | 728.87 | 0.30 | 0.00 | 0.035 |
| 7 | 56 | 167.71 | 627.00 | 4.00 | -3.70 | 0.015 |
| 8 | 56 | 789.48 | 627.00 | 0.30 | 0.00 | 0.035 |
| 20 | 56 | 2600.00 | 507.24 | 0.30 | 0.00 | 0.035 |
| 21 | 56 | 322.59 | 545.33 | 1.69 | -1.39 | 0.015 |
| 22 | 56 | 334.89 | 591.26 | 1.50 | -1.20 | 0.015 |
| 23 | 56 | 368.54 | 648.71 | 1.50 | -1.20 | 0.015 |
| 24 | 56 | 118.73 | 692.69 | 12.50 | -12.20 | 0.015 |
| 25 | 56 | 314.73 | 720.21 | 7.40 | -7.10 | 0.015 |
| 26 | 56 | 317.38 | 748.33 | 1.50 | -1.20 | 0.015 |
| 27 | 56 | 286.57 | 765.79 | 1.50 | -1.20 | 0.015 |
| 43 | 56 | 111.80 | 602.08 | 4.00 | -3.70 | 0.015 |
| 44 | 56 | 1100.00 | 602.08 | 0.30 | 0.00 | 0.035 |
| 4 | 57 | 77.62 | 499.52 | 4.00 | -3.70 | 0.015 |
| 5 | 57 | 700.00 | 499.52 | 0.30 | 0.00 | 0.035 |
| 7 | 57 | 134.63 | 783.02 | 4.00 | -3.70 | 0.015 |
| 8 | 57 | 632.16 | 783.02 | 0.30 | 0.00 | 0.035 |
| 20 | 57 | 2600.00 | 596.43 | 0.30 | 0.00 | 0.035 |
| 21 | 57 | 304.70 | 620.97 | 1.50 | -1.20 | 0.015 |
| 22 | 57 | 300.81 | 660.73 | 1.53 | -1.23 | 0.015 |
| 23 | 57 | 300.55 | 706.79 | 1.58 | -1.28 | 0.015 |
| 24 | 57 | 154.55 | 725.96 | 12.50 | -12.20 | 0.015 |
| 25 | 57 | 313.78 | 735.26 | 6.06 | -5.76 | 0.015 |
| 26 | 57 | 274.30 | 777.83 | 1.50 | -1.20 | 0.015 |
| 43 | 57 | 70.71 | 626.50 | 4.00 | -3.70 | 0.015 |
| 44 | 57 | 790.10 | 626.50 | 0.30 | 0.00 | 0.035 |
| 4 | 58 | 100.00 | 670.30 | 4.00 | -3.70 | 0.015 |
| 5 | 58 | 900.00 | 670.30 | 0.30 | 0.00 | 0.035 |
| 7 | 58 | 145.77 | 649.04 | 4.00 | -3.70 | 0.015 |
| 8 | 58 | 762.66 | 649.04 | 0.30 | 0.00 | 0.035 |
| 20 | 58 | 2600.00 | 728.30 | 0.30 | 0.00 | 0.035 |


| 21 | 58 | 263.49 | 728.27 | 1.50 | -1.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 22 | 58 | 263.96 | 690.15 | 1.50 | -1.20 | 0.015 |
| 23 | 58 | 257.00 | 644.65 | 2.32 | -2.02 | 0.015 |
| 24 | 58 | 171.71 | 625.42 | 12.50 | -12.20 | 0.015 |
| 25 | 58 | 272.94 | 617.26 | 3.98 | -3.68 | 0.015 |
| 26 | 58 | 245.19 | 571.37 | 1.50 | -1.20 | 0.015 |
| 43 | 58 | 125.00 | 760.35 | 4.00 | -3.70 | 0.015 |
| 44 | 58 | 1200.00 | 760.35 | 0.30 | 0.00 | 0.035 |
| 4 | 59 | 90.00 | 606.71 | 4.00 | -3.70 | 0.015 |
| 5 | 59 | 800.00 | 606.71 | 0.30 | 0.00 | 0.035 |
| 7 | 59 | 145.77 | 571.18 | 4.00 | -3.70 | 0.015 |
| 8 | 59 | 866.62 | 571.18 | 0.30 | 0.00 | 0.035 |
| 20 | 59 | 2400.00 | 640.74 | 0.30 | 0.00 | 0.035 |
| 21 | 59 | 241.35 | 640.74 | 1.50 | -1.20 | 0.015 |
| 22 | 59 | 253.27 | 629.91 | 1.50 | -1.20 | 0.015 |
| 23 | 59 | 152.32 | 630.25 | 1.50 | -1.20 | 0.015 |
| 24 | 59 | 157.54 | 606.79 | 12.50 | -12.20 | 0.015 |
| 25 | 59 | 260.57 | 570.62 | 4.56 | -4.26 | 0.015 |
| 26 | 59 | 282.40 | 557.67 | 1.84 | -1.54 | 0.015 |
| 27 | 59 | 349.95 | 531.00 | 1.50 | -1.20 | 0.015 |
| 43 | 59 | 125.00 | 548.29 | 4.00 | -3.70 | 0.015 |
| 44 | 59 | 1000.00 | 548.29 | 0.30 | 0.00 | 0.035 |
| 4 | 60 | 90.00 | 629.36 | 4.00 | -3.70 | 0.015 |
| 5 | 60 | 900.00 | 629.36 | 0.30 | 0.00 | 0.035 |
| 7 | 60 | 106.07 | 617.45 | 4.00 | -3.70 | 0.015 |
| 8 | 60 | 801.68 | 617.45 | 0.30 | 0.00 | 0.035 |
| 21 | 60 | 1000.00 | 520.78 | 0.30 | 0.00 | 0.035 |
| 22 | 60 | 233.22 | 520.78 | 1.50 | -1.20 | 0.015 |
| 23 | 60 | 152.32 | 509.49 | 1.50 | -1.20 | 0.015 |
| 24 | 60 | 147.15 | 532.78 | 12.50 | -12.20 | 0.015 |
| 25 | 60 | 276.16 | 553.69 | 5.14 | -4.84 | 0.015 |
| 26 | 60 | 293.30 | 534.01 | 1.50 | -1.20 | 0.015 |
| 27 | 60 | 317.05 | 518.21 | 1.97 | -1.67 | 0.015 |
| 28 | 60 | 287.74 | 494.07 | 1.50 | -1.20 | 0.015 |
| 43 | 60 | 167.71 | 742.04 | 4.00 | -3.70 | 0.015 |
| 44 | 60 | 667.08 | 742.04 | 0.30 | 0.00 | 0.035 |
| 4 | 61 | 100.00 | 618.47 | 4.00 | -3.70 | 0.015 |
| 5 | 61 | 900.00 | 618.47 | 0.30 | 0.00 | 0.035 |
| 7 | 61 | 103.08 | 682.37 | 4.00 | -3.70 | 0.015 |
| 8 | 61 | 725.42 | 682.37 | 0.30 | 0.00 | 0.035 |
| 21 | 61 | 1100.00 | 556.15 | 0.30 | 0.00 | 0.035 |
| 22 | 61 | 219.82 | 556.16 | 1.50 | -1.20 | 0.015 |
| 23 | 61 | 276.79 | 533.25 | 1.50 | -1.20 | 0.015 |
| 24 | 61 | 153.97 | 499.11 | 12.50 | -12.20 | 0.015 |
| 25 | 61 | 285.54 | 478.25 | 3.83 | -3.53 | 0.015 |
| 27 | 61 | 292.58 | 472.78 | 1.50 | -1.20 | 0.015 |
| 28 | 61 | 291.38 | 485.85 | 1.73 | -1.43 | 0.015 |
| 29 | 61 | 291.38 | 485.85 | 1.73 | -1.43 | 0.015 |
| 43 | 61 | 145.77 | 637.38 | 4.00 | -3.70 | 0.015 |
| 44 | 61 | 517.74 | 637.38 | 0.30 | 0.00 | 0.035 |
| 4 | 62 | 100.00 | 701.78 | 4.00 | -3.70 | 0.015 |
| 5 | 62 | 1000.00 | 701.78 | 0.30 | 0.00 | 0.035 |
| 7 | 62 | 103.08 | 581.49 | 4.00 | -3.70 | 0.015 |
| 8 | 62 | 851.26 | 581.49 | 0.30 | 0.00 | 0.035 |
| 21 | 62 | 1200.00 | 618.07 | 0.30 | 0.00 | 0.035 |


| 22 | 62 | 243.07 | 618.06 | 1.50 | -1.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23 | 62 | 227.88 | 569.34 | 2.05 | -1.75 | 0.015 |
| 24 | 62 | 135.47 | 532.57 | 12.50 | -12.20 | 0.015 |
| 25 | 62 | 329.32 | 490.13 | 4.66 | -4.36 | 0.015 |
| 27 | 62 | 284.02 | 390.73 | 1.50 | -1.20 | 0.015 |
| 28 | 62 | 301.69 | 313.49 | 1.88 | -1.58 | 0.015 |
| 29 | 62 | 301.69 | 313.49 | 1.88 | -1.58 | 0.015 |
| 43 | 62 | 175.00 | 654.31 | 4.00 | -3.70 | 0.015 |
| 44 | 62 | 504.34 | 654.31 | 0.30 | 0.00 | 0.035 |
| 4 | 63 | 87.32 | 754.80 | 4.00 | -3.70 | 0.015 |
| 5 | 63 | 900.00 | 754.80 | 0.30 | 0.00 | 0.035 |
| 7 | 63 | 103.08 | 682.37 | 4.00 | -3.70 | 0.015 |
| 8 | 63 | 725.42 | 682.37 | 0.30 | 0.00 | 0.035 |
| 21 | 63 | 1600.00 | 814.38 | 0.30 | 0.00 | 0.035 |
| 22 | 63 | 228.48 | 814.38 | 1.50 | -1.20 | 0.015 |
| 23 | 63 | 213.04 | 738.12 | 1.88 | -1.58 | 0.015 |
| 24 | 63 | 113.19 | 673.54 | 12.50 | -12.20 | 0.015 |
| 25 | 63 | 319.79 | 579.71 | 3.70 | -3.40 | 0.015 |
| 26 | 63 | 288.26 | 494.28 | 1.50 | -1.20 | 0.015 |
| 27 | 63 | 285.07 | 428.07 | 1.50 | -1.20 | 0.015 |
| 28 | 63 | 309.26 | 350.09 | 1.57 | -1.27 | 0.015 |
| 43 | 63 | 150.00 | 626.50 | 4.00 | -3.70 | 0.015 |
| 44 | 63 | 526.74 | 626.50 | 0.30 | 0.00 | 0.035 |
| 4 | 64 | 120.42 | 596.41 | 4.00 | -3.70 | 0.015 |
| 5 | 64 | 800.00 | 596.41 | 0.30 | 0.00 | 0.035 |
| 7 | 64 | 100.00 | 694.62 | 4.00 | -3.70 | 0.015 |
| 8 | 64 | 712.62 | 694.62 | 0.30 | 0.00 | 0.035 |
| 22 | 64 | 1200.00 | 622.28 | 0.30 | 0.00 | 0.035 |
| 23 | 64 | 203.11 | 622.29 | 2.19 | -1.89 | 0.015 |
| 24 | 64 | 158.75 | 622.29 | 12.50 | -12.20 | 0.015 |
| 25 | 64 | 234.32 | 619.28 | 5.12 | -4.82 | 0.015 |
| 26 | 64 | 271.12 | 595.81 | 1.50 | -1.20 | 0.015 |
| 27 | 64 | 297.73 | 556.52 | 1.63 | -1.33 | 0.015 |
| 28 | 64 | 310.33 | 522.29 | 1.50 | -1.20 | 0.015 |
| 43 | 64 | 106.07 | 664.27 | 4.00 | -3.70 | 0.015 |
| 44 | 64 | 496.78 | 664.27 | 0.30 | 0.00 | 0.035 |
| 4 | 65 | 122.98 | 659.34 | 4.00 | -3.70 | 0.015 |
| 5 | 65 | 900.00 | 659.34 | 0.30 | 0.00 | 0.035 |
| 7 | 65 | 103.08 | 673.15 | 4.00 | -3.70 | 0.015 |
| 8 | 65 | 735.34 | 673.15 | 0.30 | 0.00 | 0.035 |
| 22 | 65 | 1200.00 | 625.13 | 0.30 | 0.00 | 0.035 |
| 23 | 65 | 192.29 | 625.13 | 1.99 | -1.69 | 0.015 |
| 24 | 65 | 165.40 | 638.07 | 12.50 | -12.20 | 0.015 |
| 25 | 65 | 198.69 | 660.10 | 7.73 | -7.43 | 0.015 |
| 26 | 65 | 283.05 | 662.64 | 1.83 | -1.53 | 0.015 |
| 27 | 65 | 200.00 | 520.00 | 1.83 | -1.53 | 0.015 |
| 43 | 65 | 145.77 | 675.46 | 4.00 | -3.70 | 0.015 |
| 44 | 65 | 488.56 | 675.46 | 0.30 | 0.00 | 0.035 |
| 4 | 66 | 147.14 | 686.48 | 4.00 | -3.70 | 0.015 |
| 5 | 66 | 900.00 | 686.48 | 0.30 | 0.00 | 0.035 |
| 7 | 66 | 125.00 | 800.39 | 4.00 | -3.70 | 0.015 |
| 8 | 66 | 618.44 | 800.39 | 0.30 | 0.00 | 0.035 |
| 19 | 66 | 1000.00 | 536.71 | 0.30 | 0.00 | 0.035 |
| 20 | 66 | 215.46 | 536.71 | 6.00 | -5.70 | 0.015 |
| 21 | 66 | 303.07 | 538.75 | 6.00 | -5.70 | 0.015 |


| 22 | 66 | 252.31 | 551.62 | 6.00 | -5.70 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23 | 66 | 195.15 | 561.99 | 6.00 | -5.70 | 0.015 |
| 24 | 66 | 152.01 | 547.83 | 6.00 | -5.70 | 0.015 |
| 25 | 66 | 189.45 | 547.32 | 6.00 | -5.70 | 0.015 |
| 43 | 66 | 111.80 | 626.50 | 4.00 | -3.70 | 0.015 |
| 44 | 66 | 526.74 | 626.50 | 0.30 | 0.00 | 0.035 |
| 4 | 67 | 150.08 | 678.40 | 4.00 | -3.70 | 0.015 |
| 5 | 67 | 900.00 | 678.40 | 0.30 | 0.00 | 0.035 |
| 7 | 67 | 111.80 | 570.09 | 4.00 | -3.70 | 0.015 |
| 8 | 67 | 868.28 | 570.09 | 0.30 | 0.00 | 0.035 |
| 18 | 67 | 1400.00 | 742.71 | 0.30 | 0.00 | 0.035 |
| 19 | 67 | 210.42 | 755.70 | 6.00 | -5.70 | 0.015 |
| 20 | 67 | 239.33 | 774.61 | 6.00 | -5.70 | 0.015 |
| 23 | 67 | 600.00 | 495.63 | 0.30 | 0.00 | 0.035 |
| 24 | 67 | 164.97 | 521.85 | 12.50 | -12.20 | 0.015 |
| 25 | 67 | 163.50 | 534.32 | 6.54 | -6.24 | 0.015 |
| 43 | 67 | 140.00 | 551.45 | 4.00 | -3.70 | 0.015 |
| 44 | 67 | 598.42 | 551.45 | 0.30 | 0.00 | 0.035 |
| 4 | 68 | 108.17 | 617.41 | 4.00 | -3.70 | 0.015 |
| 5 | 68 | 900.00 | 617.41 | 0.30 | 0.00 | 0.035 |
| 7 | 68 | 100.00 | 715.89 | 4.00 | -3.70 | 0.015 |
| 8 | 68 | 691.44 | 715.89 | 0.30 | 0.00 | 0.035 |
| 18 | 68 | 1500.00 | 798.26 | 0.30 | 0.00 | 0.035 |
| 19 | 68 | 400.00 | 798.26 | 6.00 | -5.70 | 0.015 |
| 24 | 68 | 177.00 | 762.64 | 12.50 | -12.20 | 0.015 |
| 43 | 68 | 166.88 | 590.64 | 4.00 | -3.70 | 0.015 |
| 44 | 68 | 558.72 | 590.64 | 0.30 | 0.00 | 0.035 |
| 4 | 69 | 95.53 | 584.23 | 4.00 | -3.70 | 0.015 |
| 5 | 69 | 800.00 | 584.23 | 0.30 | 0.00 | 0.035 |
| 7 | 69 | 134.63 | 800.39 | 4.00 | -3.70 | 0.015 |
| 8 | 69 | 618.44 | 800.39 | 0.30 | 0.00 | 0.035 |
| 17 | 69 | 653.04 | 757.99 | 0.30 | 0.00 | 0.035 |
| 18 | 69 | 1500.00 | 757.99 | 0.30 | 0.00 | 0.035 |
| 19 | 69 | 400.00 | 757.99 | 6.00 | -5.70 | 0.015 |
| 23 | 69 | 386.50 | 486.34 | 0.30 | 0.00 | 0.035 |
| 24 | 69 | 185.63 | 510.73 | 12.50 | -12.20 | 0.015 |
| 43 | 69 | 167.71 | 583.63 | 4.00 | -3.70 | 0.015 |
| 44 | 69 | 565.42 | 583.63 | 0.30 | 0.00 | 0.035 |
| 4 | 70 | 97.08 | 301.04 | 4.00 | -3.70 | 0.015 |
| 5 | 70 | 400.00 | 301.04 | 0.30 | 0.00 | 0.035 |
| 7 | 70 | 160.08 | 687.84 | 4.00 | -3.70 | 0.015 |
| 8 | 70 | 719.64 | 687.84 | 0.30 | 0.00 | 0.035 |
| 17 | 70 | 650.92 | 760.46 | 0.30 | 0.00 | 0.035 |
| 18 | 70 | 1500.00 | 760.46 | 0.30 | 0.00 | 0.035 |
| 19 | 70 | 400.00 | 760.46 | 6.00 | -5.70 | 0.015 |
| 23 | 70 | 214.62 | 555.64 | 2.26 | -1.96 | 0.015 |
| 24 | 70 | 175.41 | 556.18 | 12.50 | -12.20 | 0.015 |
| 43 | 70 | 180.28 | 756.90 | 4.00 | -3.70 | 0.015 |
| 44 | 70 | 435.98 | 756.90 | 0.30 | 0.00 | 0.035 |
| 4 | 71 | 99.25 | 472.07 | 4.00 | -3.70 | 0.015 |
| 5 | 71 | 700.00 | 472.07 | 0.30 | 0.00 | 0.035 |
| 7 | 71 | 111.80 | 750.00 | 4.00 | -3.70 | 0.015 |
| 8 | 71 | 660.00 | 750.00 | 0.30 | 0.00 | 0.035 |
| 17 | 71 | 626.62 | 789.96 | 0.30 | 0.00 | 0.035 |
| 18 | 71 | 1500.00 | 789.96 | 0.30 | 0.00 | 0.035 |


| 19 | 71 | 400.00 | 789.96 | 6.00 | -5.70 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23 | 71 | 223.32 | 539.75 | 2.97 | -2.67 | 0.015 |
| 24 | 71 | 169.02 | 503.35 | 12.50 | -12.20 | 0.015 |
| 43 | 71 | 167.71 | 706.77 | 4.00 | -3.70 | 0.015 |
| 44 | 71 | 700.36 | 706.77 | 0.30 | 0.00 | 0.035 |
| 4 | 72 | 129.03 | 637.38 | 4.00 | -3.70 | 0.015 |
| 5 | 72 | 800.00 | 637.38 | 0.30 | 0.00 | 0.035 |
| 7 | 72 | 103.08 | 608.79 | 4.00 | -3.70 | 0.015 |
| 18 | 72 | 1400.00 | 768.70 | 0.30 | 0.00 | 0.035 |
| 19 | 72 | 400.00 | 768.70 | 6.00 | -5.70 | 0.015 |
| 23 | 72 | 212.06 | 465.57 | 5.72 | -5.42 | 0.015 |
| 24 | 72 | 188.46 | 450.09 | 12.50 | -12.20 | 0.015 |
| 43 | 72 | 158.11 | 583.10 | 4.00 | -3.70 | 0.015 |
| 44 | 72 | 848.92 | 583.10 | 0.30 | 0.00 | 0.035 |
| 4 | 73 | 138.29 | 593.49 | 4.00 | -3.70 | 0.015 |
| 5 | 73 | 750.00 | 593.49 | 0.30 | 0.00 | 0.035 |
| 18 | 73 | 1500.00 | 920.98 | 0.30 | 0.00 | 0.035 |
| 19 | 73 | 350.00 | 939.24 | 6.00 | -5.70 | 0.015 |
| 23 | 73 | 192.25 | 446.72 | 6.42 | -6.12 | 0.015 |
| 24 | 73 | 183.46 | 456.84 | 12.50 | -12.20 | 0.015 |
| 43 | 73 | 142.21 | 730.56 | 4.00 | -3.70 | 0.015 |
| 44 | 73 | 677.56 | 730.56 | 0.30 | 0.00 | 0.035 |
| 4 | 74 | 86.02 | 539.07 | 4.00 | -3.70 | 0.015 |
| 5 | 74 | 750.00 | 539.07 | 0.30 | 0.00 | 0.035 |
| 19 | 74 | 150.00 | 550.00 | 6.00 | -5.70 | 0.015 |
| 23 | 74 | 162.53 | 456.30 | 5.62 | -5.32 | 0.015 |
| 24 | 74 | 158.40 | 459.66 | 12.50 | -12.20 | 0.015 |
| 43 | 74 | 142.21 | 733.30 | 4.00 | -3.70 | 0.015 |
| 44 | 74 | 675.04 | 733.30 | 0.30 | 0.00 | 0.035 |
| 4 | 75 | 129.03 | 675.02 | 4.00 | -3.70 | 0.015 |
| 5 | 75 | 900.00 | 675.02 | 0.30 | 0.00 | 0.035 |
| 18 | 75 | 1200.00 | 659.62 | 0.30 | 0.00 | 0.035 |
| 19 | 75 | 150.00 | 659.62 | 5.50 | -5.20 | 0.015 |
| 23 | 75 | 132.19 | 420.70 | 3.03 | -2.73 | 0.015 |
| 24 | 75 | 137.18 | 418.01 | 12.50 | -12.20 | 0.015 |
| 43 | 75 | 190.39 | 617.45 | 4.00 | -3.70 | 0.015 |
| 44 | 75 | 801.68 | 617.45 | 0.30 | 0.00 | 0.035 |
| 4 | 76 | 180.28 | 570.09 | 4.00 | -3.70 | 0.015 |
| 5 | 76 | 800.00 | 570.09 | 0.30 | 0.00 | 0.035 |
| 18 | 76 | 898.26 | 551.06 | 0.30 | 0.00 | 0.035 |
| 19 | 76 | 150.00 | 551.06 | 5.50 | -5.20 | 0.015 |
| 23 | 76 | 262.26 | 513.46 | 0.30 | 0.00 | 0.035 |
| 24 | 76 | 153.09 | 506.97 | 11.50 | -11.20 | 0.015 |
| 43 | 76 | 201.56 | 665.68 | 4.00 | -3.70 | 0.015 |
| 44 | 76 | 743.60 | 665.68 | 0.30 | 0.00 | 0.035 |
| 4 | 77 | 145.77 | 664.27 | 4.00 | -3.70 | 0.015 |
| 5 | 77 | 800.00 | 664.27 | 0.30 | 0.00 | 0.035 |
| 18 | 77 | 704.64 | 702.49 | 0.30 | 0.00 | 0.035 |
| 19 | 77 | 150.00 | 702.49 | 5.00 | -4.70 | 0.015 |
| 23 | 77 | 282.80 | 489.27 | 0.30 | 0.00 | 0.035 |
| 24 | 77 | 164.28 | 527.01 | 10.50 | -10.20 | 0.015 |
| 43 | 77 | 176.78 | 649.04 | 4.00 | -3.70 | 0.015 |
| 44 | 77 | 762.66 | 649.04 | 0.30 | 0.00 | 0.035 |
| 4 | 78 | 117.05 | 530.75 | 4.00 | -3.70 | 0.015 |
| 5 | 78 | 700.00 | 530.75 | 0.30 | 0.00 | 0.035 |


| 19 | 78 | 150.00 | 649.44 | 4.50 | -4.20 | 0.015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 24 | 78 | 139.11 | 489.98 | 10.50 | -10.20 | 0.015 |
| 43 | 78 | 180.28 | 721.11 | 4.00 | -3.70 | 0.015 |
| 44 | 78 | 686.44 | 721.11 | 0.30 | 0.00 | 0.035 |
| 4 | 79 | 120.42 | 701.78 | 4.00 | -3.70 | 0.015 |
| 5 | 79 | 900.00 | 701.78 | 0.30 | 0.00 | 0.035 |
| 22 | 79 | 113.15 | 522.62 | 6.00 | -5.70 | 0.015 |
| 23 | 79 | 151.91 | 529.39 | 8.00 | -7.70 | 0.015 |
| 24 | 79 | 140.83 | 508.50 | 8.00 | -7.70 | 0.015 |
| 43 | 79 | 223.61 | 610.33 | 4.00 | -3.70 | 0.015 |
| 44 | 79 | 847.22 | 610.33 | 0.30 | 0.00 | 0.035 |
| 4 | 80 | 107.35 | 634.05 | 4.00 | -3.70 | 0.015 |
| 5 | 80 | 800.00 | 634.05 | 0.30 | 0.00 | 0.035 |
| 21 | 80 | 914.48 | 541.29 | 0.30 | 0.00 | 0.035 |
| 22 | 80 | 150.00 | 541.29 | 6.50 | -6.20 | 0.015 |
| 24 | 80 | 140.83 | 587.83 | 10.50 | -10.20 | 0.015 |
| 43 | 80 | 201.56 | 638.85 | 4.00 | -3.70 | 0.015 |
| 44 | 80 | 501.56 | 638.85 | 4.00 | -3.70 | 0.015 |
| 4 | 81 | 120.21 | 607.82 | 4.00 | -3.70 | 0.015 |
| 5 | 81 | 800.00 | 607.82 | 0.30 | 0.00 | 0.035 |
| 21 | 81 | 1000.00 | 589.82 | 0.30 | 0.00 | 0.035 |
| 22 | 81 | 150.00 | 589.82 | 7.00 | -6.70 | 0.015 |
| 24 | 81 | 157.98 | 583.93 | 10.50 | -10.20 | 0.015 |
| 43 | 81 | 201.56 | 1038.85 | 4.00 | -3.70 | 0.015 |
| 44 | 81 | 1003.12 | 1038.85 | 0.30 | 0.00 | 0.035 |
| 4 | 82 | 105.95 | 685.00 | 4.00 | -3.70 | 0.015 |
| 5 | 82 | 675.00 | 685.00 | 0.30 | 0.00 | 0.035 |
| 43 | 82 | 201.56 | 2300.00 | 4.00 | -3.70 | 0.015 |
| 44 | 82 | 1003.12 | 2300.00 | 0.30 | 0.00 | 0.035 |
| 4 | 83 | 105.95 | 603.92 | 4.00 | -3.70 | 0.015 |
| 5 | 83 | 600.00 | 603.92 | 0.30 | 0.00 | 0.035 |
| 43 | 83 | 201.56 | 700.00 | 4.00 | -3.70 | 0.015 |
| 44 | 83 | 1003.12 | 700.00 | 0.30 | 0.00 | 0.035 |
| 4 | 84 | 103.08 | 589.94 | 4.00 | -3.70 | 0.015 |
| 5 | 84 | 600.00 | 589.94 | 0.30 | 0.00 | 0.035 |
| 43 | 84 | 201.56 | 1650.00 | 4.00 | -3.70 | 0.015 |
| 44 | 84 | 1003.12 | 1650.00 | 0.30 | 0.00 | 0.035 |
| 4 | 85 | 134.54 | 628.01 | 4.00 | -3.70 | 0.015 |
| 5 | 85 | 525.00 | 628.01 | 0.30 | 0.00 | 0.035 |
| 43 | 85 | 201.56 | 2140.00 | 4.00 | -3.70 | 0.015 |
| 44 | 85 | 800.00 | 2140.00 | 0.30 | 0.00 | 0.035 |
| 43 | 86 | 201.56 | 1820.00 | 4.00 | -3.70 | 0.015 |
| 44 | 86 | 1000.00 | 1820.00 | 0.30 | 0.00 | 0.035 |
| 43 | 87 | 201.56 | 2570.00 | 4.00 | -3.70 | 0.015 |
| 44 | 87 | 1400.00 | 2570.00 | 0.30 | 0.00 | 0.035 |
| 43 | 88 | 201.56 | 2160.00 | 4.00 | -3.70 | 0.015 |
| 44 | 88 | 1600.00 | 2160.00 | 0.30 | 0.00 | 0.035 |
| 43 | 89 | 101.56 | 2500.00 | 4.00 | -3.70 | 0.015 |
| 44 | 89 | 903.12 | 2500.00 | 0.30 | 0.00 | 0.035 |
| 43 | 90 | 101.56 | 2100.00 | 4.00 | -3.70 | 0.015 |
| 44 | 90 | 603.12 | 2100.00 | 0.30 | 0.00 | 0.035 |
| 43 | 91 | 101.56 | 1900.00 | 4.00 | -3.70 | 0.015 |
| 44 | 91 | 903.12 | 1900.00 | 0.30 | 0.00 | 0.035 |
| 43 | 92 | 101.56 | 2360.00 | 4.00 | -3.70 | 0.015 |
| 44 | 92 | 1403.12 | 2360.00 | 0.30 | 0.00 | 0.035 |


| 43 | 93 | 101.56 | 2000.00 | 4.00 | -3.70 | 0.015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 44 | 93 | 803.12 | 2000.00 | 0.30 | 0.00 | 0.035 |
| 43 | 94 | 71.56 | 1350.00 | 4.00 | -3.70 | 0.015 |
| 44 | 94 | 603.12 | 1350.00 | 0.30 | 0.00 | 0.035 |
| 43 | 95 | 71.56 | 1220.00 | 4.00 | -3.70 | 0.015 |
| 44 | 95 | 603.12 | 1220.00 | 0.30 | 0.00 | 0.035 |
| 43 | 96 | 71.56 | 1550.00 | 4.00 | -3.70 | 0.015 |
| 44 | 96 | 603.12 | 1550.00 | 0.30 | 0.00 | 0.035 |
| 43 | 97 | 71.56 | 710.00 | 4.00 | -3.70 | 0.015 |
| 44 | 97 | 603.12 | 710.00 | 0.30 | 0.00 | 0.035 |


| C I |  | J XLNUTME | YLTUTMN C |  | CCUE | CCVE | CCUN | CCVN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| C |  |  |  |  |  |  |  |  |
| 30 | 5 | 28.6640 | 5.1879 | 0.6001682 | 0.7998738 | -0.7998738 | 0.6001682 | $-1.0000000$ |
| 31 | 5 | 28.9620 | 4.7698 | 0.6388829 | 0.7693040 | -0.7693040 | 0.6388829 | -1.0000000 |
| 32 | 5 | 29.2820 | 4.3761 | 0.6796822 | 0.7335067 | -0.7335067 | 0.6796822 | -1.0000000 |
| 33 | 5 | 29.6730 | 3.9819 | 0.7768199 | 0.6297228 | -0.6297228 | 0.7768199 | -1.0000000 |
| 30 | 6 | 29.0820 | 5.5352 | 0.6028011 | 0.7978915 | -0.7978915 | 0.6028011 | -1.0000000 |
| 31 | 6 | 29.3690 | 5.1438 | 0.6482844 | 0.7613983 | -0.7613983 | 0.6482844 | -1.0000000 |
| 32 | 6 | 29.6770 | 4.7749 | 0.6946676 | 0.7193308 | -0.7193308 | 0.6946676 | -1.0000000 |
| 33 | 6 | 30.0280 | 4.4300 | 0.7529863 | 0.6580362 | -0.6580362 | 0.7529863 | -1.0000000 |
| 4 | 7 | 33.5230 | 45.5200 | 0.7695577 | -0.6385772 | 0.6385772 | 0.7695577 | -1.0000000 |
| 5 | 7 | 33.5230 | 45.5200 | 0.7695577 | -0.6385772 | 0.6385772 | 0.7695577 | -0.0100000 |
| 24 | 7 | 28.0190 | 7.6812 | 0.6344351 | 0.7729761 | -0.7729761 | 0.6344351 | -1.0000000 |
| 25 | 7 | 28.2230 | 7.4320 | 0.6772386 | 0.7357635 | -0.7357635 | 0.6772386 | -1.0000000 |
| 26 | 7 | 28.4460 | 7.1765 | 0.6609870 | 0.7503973 | -0.7503973 | 0.6609870 | -1.0000000 |
| 30 | 7 | 29.4490 | 5.8471 | 0.6120957 | 0.7907837 | -0.7907837 | 0.6120957 | -1.0000000 |
| 31 | 7 | 29.7220 | 5.4720 | 0.6479908 | 0.7616482 | -0.7616482 | 0.6479908 | -1.0000000 |
| 32 | 7 | 30.0100 | 5.1238 | 0.6942729 | 0.7197118 | -0.7197118 | 0.6942729 | -1.0000000 |
| 33 | 7 | 30.3260 | 4.8009 | 0.7750288 | 0.6319259 | -0.6319259 | 0.7750288 | -1.0000000 |
| 4 | 8 | 33.1580 | 46.0970 | 0.9412231 | -0.3377853 | 0.3377853 | 0.9412231 | -1.0000000 |
| 5 | 8 | 33.1580 | 46.0970 | 0.9412231 | -0.3377853 | 0.3377853 | 0.9412231 | -0.0100000 |
| 24 | 8 | 28.2870 | 7.9212 | 0.6309558 | 0.7758188 | -0.7758188 | 0.6309558 | -1.0000000 |
| 25 | 8 | 28.4840 | 7.6746 | 0.6651604 | 0.7467005 | -0.7467005 | 0.6651604 | -1.0000000 |
| 26 | 8 | 28.7070 | 7.4198 | 0.6756155 | 0.7372542 | -0.7372542 | 0.6756155 | -1.0000000 |
| 27 | 8 | 28.9600 | 7.1332 | 0.6596907 | 0.7515372 | -0.7515372 | 0.6596907 | -1.0000000 |
| 30 | 8 | 29.7570 | 6.1308 | 0.6470400 | 0.7624561 | -0.7624561 | 0.6470400 | -1.0000000 |
| 31 | 8 | 30.0340 | 5.7690 | 0.6630404 | 0.7485836 | -0.7485836 | 0.6630404 | -1.0000000 |
| 32 | 8 | 30.3100 | 5.4391 | 0.6918504 | 0.7220409 | -0.7220409 | 0.6918504 | -1.0000000 |
| 33 | 8 | 30.5830 | 5.1411 | 0.7389004 | 0.6738147 | -0.6738147 | 0.7389004 | -1.0000000 |
| 4 | 9 | 32.9380 | 46.7320 | 0.9547390 | -0.2974447 | 0.2974447 | 0.9547390 | -1.0000000 |
| 5 | 9 | 32.9380 | 46.7320 | 0.9547390 | -0.2974447 | 0.2974447 | 0.9547390 | -0.0100000 |
| 24 | 9 | 28.5550 | 8.1530 | 0.6153132 | 0.7882828 | -0.7882828 | 0.6153132 | -1.0000000 |
| 25 | 9 | 28.7400 | 7.9174 | 0.6749297 | 0.7378821 | -0.7378821 | 0.6749297 | -1.0000000 |
| 26 | 9 | 28.9640 | 7.6705 | 0.6988578 | 0.7152606 | -0.7152606 | 0.6988578 | -1.0000000 |
| 27 | 9 | 29.2290 | 7.3811 | 0.6789733 | 0.7341629 | -0.7341629 | 0.6789733 | -1.0000000 |
| 28 | 9 | 29.4970 | 7.0564 | 0.6529557 | 0.7573961 | -0.7573961 | 0.6529557 | -1.0000000 |
| 29 | 9 | 29.7580 | 6.7258 | 0.6482436 | 0.7614330 | -0.7614330 | 0.6482436 | -1.0000000 |
| 30 | 9 | 30.0350 | 6.3874 | 0.6454550 | 0.7637983 | -0.7637983 | 0.6454550 | -1.0000000 |
| 31 | 9 | 30.3190 | 6.0487 | 0.6757985 | 0.7370864 | -0.7370864 | 0.6757985 | -1.0000000 |
| 32 | 9 | 30.5940 | 5.7372 | 0.6998252 | 0.7143142 | -0.7143142 | 0.6998252 | -1.0000000 |
| 33 | 9 | 30.8620 | 5.4467 | 0.6859388 | 0.7276593 | -0.7276593 | 0.6859388 | -1.0000000 |
| 4 | 10 | 32.6210 | 47.1310 | 0.5318148 | -0.8468607 | 0.8468607 | 0.5318148 | -1.0000000 |
| 5 | 10 | 32.6210 | 47.1310 | 0.5318148 | -0.8468607 | 0.8468607 | 0.5318148 | -0.0100000 |
| 23 | 10 | 28.6800 | 8.6055 | 0.5486947 | 0.8360228 | -0.8360228 | 0.5486947 | -1.0000000 |
| 24 | 10 | 28.8490 | 8.3695 | 0.5735527 | 0.8191687 | -0.8191687 | 0.5735527 | -1.0000000 |
| 25 | 10 | 29.0040 | 8.1696 | 0.6566030 | 0.7542363 | -0.7542363 | 0.6566030 | -1.0000000 |
| 26 | 10 | 29.2030 | 7.9222 | 0.6866325 | 0.7270048 | -0.7270048 | 0.6866325 | -1.0000000 |
| 27 | 10 | 29.4740 | 7.6099 | 0.6745856 | 0.7381966 | -0.7381966 | 0.6745856 | -1.0000000 |
| 28 | 10 | 29.7480 | 7.2974 | 0.6649767 | 0.7468641 | -0.7468641 | 0.6649767 | -1.0000000 |
| 29 | 10 | 30.0210 | 6.9668 | 0.6622297 | 0.7493010 | -0.7493010 | 0.6622297 | -1.0000000 |
| 30 | 10 | 30.3110 | 6.6319 | 0.6732972 | 0.7393720 | -0.7393720 | 0.6732972 | -1.0000000 |
| 31 | 10 | 30.5880 | 6.3166 | 0.6797018 | 0.7334886 | -0.7334886 | 0.6797018 | -1.0000000 |
| 32 | 10 | 30.8630 | 6.0256 | 0.7320566 | 0.6812438 | -0.6812438 | 0.7320566 | -1.0000000 |
| 33 | 10 | 31.1380 | 5.7369 | 0.7078733 | 0.7063394 | -0.7063394 | 0.7078733 | -1.0000000 |
| 3 | 11 | 32.1750 | 47.2450 | 0.4929779 | -0.8700418 | 0.8700418 | 0.4929779 | -1.0000000 |
| 4 | 11 | 32.1750 | 47.3550 | 0.4929779 | -0.8700418 | 0.8700418 | 0.4929779 | -1.0000000 |
| 5 | 11 | 32.1750 | 47.3550 | 0.4929779 | -0.8700418 | 0.8700418 | 0.4929779 | -0.0100000 |
| 23 | 11 | 29.0800 | 8.8595 | 0.5535087 | 0.8328434 | -0.8328434 | 0.5535087 | -1.0000000 |
| 24 | 11 | 29.2360 | 8.6438 | 0.6071576 | 0.7945815 | -0.7945815 | 0.6071576 | -1.0000000 |
| 25 | 11 | 29.3690 | 8.4587 | 0.5625369 | 0.8267722 | -0.8267722 | 0.5625369 | -1.0000000 |
| 26 | 11 | 29.5440 | 8.2018 | 0.5647510 | 0.8252614 | -0.8252614 | 0.5647510 | -1.0000000 |
| 27 | 11 | 29.7880 | 7.8883 | 0.6587110 | 0.7523960 | -0.7523960 | 0.6587110 | -1.0000000 |
| 28 | 11 | 30.0535 | 7.5724 | 0.6678548 | 0.7441901 | -0.7441901 | 0.6678548 | -1.0000000 |
| 29 | 11 | 30.3190 | 7.2564 | 0.6769987 | 0.7359842 | -0.7359842 | 0.6769987 | -1.0000000 |
| 30 | 11 | 30.6110 | 6.9355 | 0.7159277 | 0.6981744 | -0.6981744 | 0.7159277 | -1.0000000 |
| 31 | 11 | 30.8900 | 6.6348 | 0.7010635 | 0.7130989 | -0.7130989 | 0.7010635 | -1.0000000 |
| 32 | 11 | 31.1660 | 6.3502 | 0.7262066 | 0.6874765 | -0.6874765 | 0.7262066 | -1.0000000 |
| 33 | 11 | 31.4390 | 6.0647 | 0.6865529 | 0.7270799 | -0.7270799 | 0.6865529 | -1.0000000 |
| 4 | 12 | 32.0000 | 47.7430 | 0.9989396 | 0.0460408 | -0.0460408 | 0.9989396 | -1.0000000 |


| 5 | 12 | 32.0000 | 47.7430 | 0.9989396 | 0.0460408 | -0.0460408 | 0.9989396 | -0.0100000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23 | 12 | 29.5680 | 9.1768 | 0.5495203 | 0.8354804 | -0.8354804 | 0.5495203 | -1.0000000 |
| 24 | 12 | 29.6990 | 8.9865 | 0.5917935 | 0.8060896 | -0.8060896 | 0.5917935 | $-1.0000000$ |
| 25 | 12 | 29.8380 | 8.7892 | 0.5730659 | 0.8195093 | -0.8195093 | 0.5730659 | -1.0000000 |
| 27 | 12 | 30.2180 | 8.2379 | 0.5987160 | 0.8009614 | -0.8009614 | 0.5987160 | $-1.0000000$ |
| 28 | 12 | 30.4615 | 7.9312 | 0.6319295 | 0.7738387 | -0.7738387 | 0.6319295 | $-1.0000000$ |
| 29 | 12 | 30.7050 | 7.6244 | 0.6651430 | 0.7467161 | -0.7467161 | 0.6651430 | $-1.0000000$ |
| 30 | 12 | 30.9760 | 7.3184 | 0.6952052 | 0.7188114 | -0.7188114 | 0.6952052 | -1.0000000 |
| 31 | 12 | 31.2530 | 7.0230 | 0.7065539 | 0.7076592 | -0.7076592 | 0.7065539 | $-1.0000000$ |
| 32 | 12 | 31.5290 | 6.7329 | 0.7077067 | 0.7065063 | -0.7065063 | 0.7077067 | $-1.0000000$ |
| 33 | 12 | 31.8060 | 6.4421 | 0.7027650 | 0.7114220 | -0.7114220 | 0.7027650 | $-1.0000000$ |
| 4 | 13 | 31.8250 | 48.2500 | 0.7254514 | -0.6882734 | 0.6882734 | 0.7254514 | -1.0000000 |
| 5 | 13 | 31.8250 | 48.2500 | 0.7254514 | -0.6882734 | 0.6882734 | 0.7254514 | -0.0100000 |
| 23 | 13 | 30.0680 | 9.5190 | 0.5866270 | 0.8098572 | -0.8098572 | 0.5866270 | -1.0000000 |
| 24 | 13 | 30.1990 | 9.3415 | 0.5822363 | 0.8130196 | -0.8130196 | 0.5822363 | -1.0000000 |
| 25 | 13 | 30.3380 | 9.1375 | 0.5522318 | 0.8336906 | -0.8336906 | 0.5522318 | -1.0000000 |
| 26 | 13 | 30.5175 | 8.8680 | 0.5866877 | 0.8098133 | -0.8098133 | 0.5866877 | -1.0000000 |
| 27 | 13 | 30.6970 | 8.5984 | 0.5866877 | 0.8098133 | -0.8098133 | 0.5866877 | $-1.0000000$ |
| 28 | 13 | 30.9070 | 8.3116 | 0.6182861 | 0.7859532 | -0.7859532 | 0.6182861 | -1.0000000 |
| 29 | 13 | 31.1560 | 8.0202 | 0.6450496 | 0.7624805 | -0.7624805 | 0.6450496 | -1.0000000 |
| 30 | 13 | 31.4050 | 7.7289 | 0.6718131 | 0.7390079 | -0.7390079 | 0.6718131 | -1.0000000 |
| 31 | 13 | 31.6540 | 7.4375 | 0.6985766 | 0.7155352 | -0.7155352 | 0.6985766 | $-1.0000000$ |
| 32 | 13 | 31.9260 | 7.1516 | 0.7109920 | 0.7032000 | -0.7032000 | 0.7109920 | -1.0000000 |
| 33 | 13 | 32.2040 | 6.8684 | 0.7247499 | 0.6890120 | -0.6890120 | 0.7247499 | -1.0000000 |
| 4 | 14 | 31.3000 | 48.4930 | -0.0124970 | -0.9999219 | 0.9999219 | -0.0124970 | $-1.0000000$ |
| 5 | 14 | 31.3000 | 48.4930 | -0.0124970 | -0.9999219 | 0.9999219 | -0.0124970 | -0.0100000 |
| 23 | 14 | 30.5680 | 9.8490 | 0.5580681 | 0.8297951 | -0.8297951 | 0.5580681 | -1.0000000 |
| 24 | 14 | 30.6990 | 9.6715 | 0.5630806 | 0.8264019 | -0.8264019 | 0.5630806 | -1.0000000 |
| 25 | 14 | 30.8320 | 9.4710 | 0.5399109 | 0.8417222 | -0.8417222 | 0.5399109 | $-1.0000000$ |
| 26 | 14 | 30.9940 | 9.2200 | 0.5727640 | 0.8197203 | -0.8197203 | 0.5727640 | $-1.0000000$ |
| 27 | 14 | 31.1740 | 8.9626 | 0.6011724 | 0.7991194 | -0.7991194 | 0.6011724 | -1.0000000 |
| 28 | 14 | 31.3730 | 8.6969 | 0.6280314 | 0.7781880 | -0.7781880 | 0.6280314 | -1.0000000 |
| 29 | 14 | 31.5920 | 8.4243 | 0.6537747 | 0.7566892 | -0.7566892 | 0.6537747 | $-1.0000000$ |
| 30 | 14 | 31.8290 | 8.1477 | 0.6768548 | 0.7361166 | -0.7361166 | 0.6768548 | $-1.0000000$ |
| 31 | 14 | 32.0810 | 7.8698 | 0.6953431 | 0.7186779 | -0.7186779 | 0.6953431 | -1.0000000 |
| 32 | 14 | 32.3450 | 7.5919 | 0.7085580 | 0.7056525 | -0.7056525 | 0.7085580 | $-1.0000000$ |
| 33 | 14 | 32.6170 | 7.3151 | 0.7161195 | 0.6979777 | -0.6979777 | 0.7161195 | $-1.0000000$ |
| 34 | 14 | 32.9000 | 7.0411 | 0.7244834 | 0.6892922 | -0.6892922 | 0.7244834 | $-1.0000000$ |
| 4 | 15 | 30.8750 | 48.7100 | 0.7983498 | -0.6021940 | 0.6021940 | 0.7983498 | -1.0000000 |
| 5 | 15 | 30.8750 | 48.7100 | 0.7983498 | -0.6021940 | 0.6021940 | 0.7983498 | -0.0100000 |
| 23 | 15 | 31.0180 | 10.1660 | 0.6410422 | 0.7675056 | -0.7675056 | 0.6410422 | $-1.0000000$ |
| 24 | 15 | 31.1430 | 9.9888 | 0.5650052 | 0.8250874 | -0.8250874 | 0.5650052 | -1.0000000 |
| 25 | 15 | 31.2750 | 9.7938 | 0.6155017 | 0.7881356 | -0.7881356 | 0.6155017 | -1.0000000 |
| 26 | 15 | 31.4440 | 9.5743 | 0.6325455 | 0.7745232 | -0.7745232 | 0.6325455 | $-1.0000000$ |
| 27 | 15 | 31.6290 | 9.3409 | 0.6436657 | 0.7653068 | -0.7653068 | 0.6436657 | $-1.0000000$ |
| 28 | 15 | 31.8290 | 9.0956 | 0.6558726 | 0.7548716 | -0.7548716 | 0.6558726 | -1.0000000 |
| 29 | 15 | 32.0440 | 8.8410 | 0.6704367 | 0.7419667 | -0.7419667 | 0.6704367 | -1.0000000 |
| 30 | 15 | 32.2740 | 8.5800 | 0.6862755 | 0.7273417 | -0.7273417 | 0.6862755 | -1.0000000 |
| 31 | 15 | 32.5180 | 8.3148 | 0.7016494 | 0.7125223 | -0.7125223 | 0.7016494 | $-1.0000000$ |
| 32 | 15 | 32.7740 | 8.0470 | 0.7152123 | 0.6989073 | -0.6989073 | 0.7152123 | -1.0000000 |
| 34 | 15 | 33.3150 | 7.5026 | 0.7367963 | 0.6761148 | -0.6761148 | 0.7367963 | -1.0000000 |
| 35 | 15 | 33.5830 | 7.2376 | 0.7430266 | 0.6692619 | -0.6692619 | 0.7430266 | $-1.0000000$ |
| 36 | 15 | 33.8390 | 6.9933 | 0.7355433 | 0.6774777 | -0.6774777 | 0.7355433 | $-1.0000000$ |
| 37 | 15 | 34.1060 | 6.7256 | 0.7077494 | 0.7064636 | -0.7064636 | 0.7077494 | -1.0000000 |
| 43 | 15 | 35.4910 | 46.6510 | 0.9433333 | 0.3318467 | -0.3318467 | 0.9433333 | $-1.0000000$ |
| 4 | 16 | 31.0250 | 49.0350 | 0.3951271 | 0.9186265 | -0.9186265 | 0.3951271 | $-1.0000000$ |
| 5 | 16 | 31.0250 | 49.0350 | 0.3951271 | 0.9186265 | -0.9186265 | 0.3951271 | -0.0100000 |
| 23 | 16 | 31.4440 | 10.5090 | 0.6623239 | 0.7492176 | -0.7492176 | 0.6623239 | -1.0000000 |
| 24 | 16 | 31.5690 | 10.3560 | 0.6115171 | 0.7912312 | -0.7912312 | 0.6115171 | $-1.0000000$ |
| 25 | 16 | 31.7060 | 10.1800 | 0.6786932 | 0.7344219 | -0.7344219 | 0.6786932 | $-1.0000000$ |
| 26 | 16 | 31.8850 | 9.9709 | 0.6667828 | 0.7452521 | -0.7452521 | 0.6667828 | $-1.0000000$ |
| 27 | 16 | 32.0750 | 9.7469 | 0.6655040 | 0.7463943 | -0.7463943 | 0.6655040 | $-1.0000000$ |
| 28 | 16 | 32.2760 | 9.5120 | 0.6715572 | 0.7409527 | -0.7409527 | 0.6715572 | $-1.0000000$ |
| 29 | 16 | 32.4545 | 9.2503 | 0.6715572 | 0.7409527 | -0.7409527 | 0.6715572 | -1.0000000 |
| 31 | 16 | 32.9530 | 8.7686 | 0.7104025 | 0.7037957 | -0.7037957 | 0.7104025 | $-1.0000000$ |
| 32 | 16 | 33.2040 | 8.5135 | 0.7240213 | 0.6897777 | -0.6897777 | 0.7240213 | -1.0000000 |
| 35 | 16 | 34.0120 | 7.7439 | 0.7465540 | 0.6653248 | -0.6653248 | 0.7465540 | $-1.0000000$ |
| 36 | 16 | 34.2830 | 7.4877 | 0.7335630 | 0.6796215 | -0.6796215 | 0.7335630 | $-1.0000000$ |
| 37 | 16 | 34.5560 | 7.2078 | 0.7065688 | 0.7076443 | -0.7076443 | 0.7065688 | -1.0000000 |
| 42 | 16 | 35.4080 | 47.2940 | 0.9998655 | -0.0164028 | 0.0164028 | 0.9998655 | $-1.0000000$ |
| 43 | 16 | 35.5840 | 47.2940 | 0.9999119 | 0.0132733 | -0.0132733 | 0.9999119 | $-1.0000000$ |


| 44 | 16 | 35.5840 | 47.2940 | 0.9999119 | 0.0132733 | -0.0132733 | 0.9999119 | -0.0100000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 17 | 31.5500 | 49.3180 | 0.6234583 | 0.7818566 | -0.7818566 | 0.6234583 | -1.0000000 |
| 5 | 17 | 31.5500 | 49.3180 | 0.6234583 | 0.7818566 | -0.7818566 | 0.6234583 | -0.0100000 |
| 22 | 17 | 31.7230 | 11.0800 | 0.6277183 | 0.7784406 | -0.7784406 | 0.6277183 | $-1.0000000$ |
| 23 | 17 | 31.8730 | 10.8890 | 0.6934327 | 0.7205215 | -0.7205215 | 0.6934327 | $-1.0000000$ |
| 24 | 17 | 31.9940 | 10.7620 | 0.7079788 | 0.7062337 | -0.7062337 | 0.7079788 | $-1.0000000$ |
| 25 | 17 | 32.1360 | 10.6040 | 0.6761965 | 0.7367213 | -0.7367213 | 0.6761965 | $-1.0000000$ |
| 26 | 17 | 32.3220 | 10.3860 | 0.6693958 | 0.7429060 | -0.7429060 | 0.6693958 | -1.0000000 |
| 27 | 17 | 32.5140 | 10.1620 | 0.6715772 | 0.7409346 | -0.7409346 | 0.6715772 | -1.0000000 |
| 28 | 17 | 32.7130 | 9.9327 | 0.6794583 | 0.7337142 | -0.7337142 | 0.6794583 | -1.0000000 |
| 29 | 17 | 32.9230 | 9.6993 | 0.6910695 | 0.7227883 | -0.7227883 | 0.6910695 | $-1.0000000$ |
| 31 | 17 | 33.3770 | 9.2219 | 0.7194868 | 0.6945061 | -0.6945061 | 0.7194868 | -1.0000000 |
| 32 | 17 | 33.6230 | 8.9790 | 0.7332160 | 0.6799958 | -0.6799958 | 0.7332160 | -1.0000000 |
| 33 | 17 | 33.8830 | 8.7333 | 0.7442197 | 0.6679350 | -0.6679350 | 0.7442197 | -1.0000000 |
| 34 | 17 | 34.1560 | 8.4847 | 0.7507761 | 0.6605567 | -0.6605567 | 0.7507761 | -1.0000000 |
| 35 | 17 | 34.4380 | 8.2353 | 0.7543160 | 0.6565115 | -0.6565115 | 0.7543160 | -1.0000000 |
| 36 | 17 | 34.7140 | 7.9904 | 0.7539246 | 0.6569610 | -0.6569610 | 0.7539246 | -1.0000000 |
| 42 | 17 | 35.4020 | 47.9410 | 0.9998367 | -0.0180733 | 0.0180733 | 0.9998367 | -1.0000000 |
| 43 | 17 | 35.5720 | 47.9470 | 0.9991729 | -0.0406649 | 0.0406649 | 0.9991729 | $-1.0000000$ |
| 44 | 17 | 35.5720 | 47.9470 | 0.9991729 | -0.0406649 | 0.0406649 | 0.9991729 | -0.0100000 |
| 4 | 18 | 31.5500 | 49.7370 | 0.8638121 | -0.5038141 | 0.5038141 | 0.8638121 | -1.0000000 |
| 5 | 18 | 31.5500 | 49.7370 | 0.8638121 | -0.5038141 | 0.5038141 | 0.8638121 | -0.0100000 |
| 22 | 18 | 32.1260 | 11.4760 | 0.6805899 | 0.7326646 | -0.7326646 | 0.6805899 | $-1.0000000$ |
| 23 | 18 | 32.2760 | 11.2980 | 0.6860554 | 0.7275493 | -0.7275493 | 0.6860554 | -1.0000000 |
| 24 | 18 | 32.3940 | 11.1680 | 0.7079788 | 0.7062337 | -0.7062337 | 0.7079788 | -1.0000000 |
| 25 | 18 | 32.5430 | 11.0090 | 0.6895712 | 0.7242179 | -0.7242179 | 0.6895712 | -1.0000000 |
| 26 | 18 | 32.7380 | 10.7900 | 0.6833805 | 0.7300624 | -0.7300624 | 0.6833805 | -1.0000000 |
| 27 | 18 | 32.9340 | 10.5700 | 0.6846561 | 0.7288663 | -0.7288663 | 0.6846561 | -1.0000000 |
| 28 | 18 | 33.1340 | 10.3490 | 0.6913524 | 0.7225177 | -0.7225177 | 0.6913524 | -1.0000000 |
| 29 | 18 | 33.3400 | 10.1250 | 0.7019037 | 0.7122719 | -0.7122719 | 0.7019037 | $-1.0000000$ |
| 30 | 18 | 33.5570 | 9.8983 | 0.7149797 | 0.6991452 | -0.6991452 | 0.7149797 | -1.0000000 |
| 31 | 18 | 33.7840 | 9.6696 | 0.7291431 | 0.6843613 | -0.6843613 | 0.7291431 | -1.0000000 |
| 32 | 18 | 34.0260 | 9.4373 | 0.7423468 | 0.6700159 | -0.6700159 | 0.7423468 | -1.0000000 |
| 34 | 18 | 34.5580 | 8.9483 | 0.7450547 | 0.6670034 | -0.6670034 | 0.7450547 | $-1.0000000$ |
| 35 | 18 | 34.8220 | 8.6810 | 0.7163877 | 0.6977025 | -0.6977025 | 0.7163877 | -1.0000000 |
| 42 | 18 | 35.3930 | 48.5290 | 0.9984258 | -0.0560881 | 0.0560881 | 0.9984258 | -1.0000000 |
| 43 | 18 | 35.5570 | 48.5420 | 0.9990983 | -0.0424564 | 0.0424564 | 0.9990983 | -1.0000000 |
| 44 | 18 | 35.5570 | 48.5420 | 0.9990983 | -0.0424564 | 0.0424564 | 0.9990983 | -0.0100000 |
| 4 | 19 | 31.0250 | 49.9170 | -0.2767535 | -0.9609410 | 0.9609410 | -0.2767535 | $-1.0000000$ |
| 5 | 19 | 31.0250 | 49.9170 | -0.2767535 | -0.9609410 | 0.9609410 | -0.2767535 | -0.0100000 |
| 20 | 19 | 32.1960 | 12.2820 | 0.6305340 | 0.7761616 | -0.7761616 | 0.6305340 | -1.0000000 |
| 21 | 19 | 32.3630 | 12.0770 | 0.6679710 | 0.7441873 | -0.7441873 | 0.6679710 | -1.0000000 |
| 22 | 19 | 32.5300 | 11.8840 | 0.6722420 | 0.7403315 | -0.7403315 | 0.6722420 | -1.0000000 |
| 23 | 19 | 32.6760 | 11.7080 | 0.6752896 | 0.7375527 | -0.7375527 | 0.6752896 | $-1.0000000$ |
| 24 | 19 | 32.8060 | 11.5750 | 0.7358146 | 0.6771830 | -0.6771830 | 0.7358146 | -1.0000000 |
| 25 | 19 | 32.9630 | 11.4130 | 0.6635685 | 0.7481155 | -0.7481155 | 0.6635685 | -1.0000000 |
| 26 | 19 | 33.1530 | 11.1940 | 0.6770118 | 0.7359721 | -0.7359721 | 0.6770118 | -1.0000000 |
| 27 | 19 | 33.3450 | 10.9780 | 0.6874166 | 0.7262633 | -0.7262633 | 0.6874166 | -1.0000000 |
| 28 | 19 | 33.5420 | 10.7630 | 0.6979731 | 0.7161240 | -0.7161240 | 0.6979731 | -1.0000000 |
| 29 | 19 | 33.7440 | 10.5480 | 0.7100492 | 0.7041521 | -0.7041521 | 0.7100492 | $-1.0000000$ |
| 30 | 19 | 33.9550 | 10.3320 | 0.7239887 | 0.6898119 | -0.6898119 | 0.7239887 | -1.0000000 |
| 31 | 19 | 34.1760 | 10.1150 | 0.7396690 | 0.6729708 | -0.6729708 | 0.7396690 | -1.0000000 |
| 32 | 19 | 34.4120 | 9.8950 | 0.7566243 | 0.6538498 | -0.6538498 | 0.7566243 | -1.0000000 |
| 33 | 19 | 34.6690 | 9.6687 | 0.7727824 | 0.6346710 | -0.6346710 | 0.7727824 | $-1.0000000$ |
| 34 | 19 | 34.9370 | 9.4099 | 0.7383686 | 0.6743974 | -0.6743974 | 0.7383686 | -1.0000000 |
| 42 | 19 | 35.3220 | 49.1390 | 0.9754158 | -0.2203725 | 0.2203725 | 0.9754158 | -1.0000000 |
| 43 | 19 | 35.4920 | 49.1770 | 0.9835003 | -0.1809066 | 0.1809066 | 0.9835003 | -1.0000000 |
| 44 | 19 | 35.4920 | 49.1770 | 0.9835003 | -0.1809066 | 0.1809066 | 0.9835003 | -0.0100000 |
| 4 | 20 | 30.5500 | 49.7130 | -0.3124077 | -0.9499481 | 0.9499481 | -0.3124077 | -1.0000000 |
| 5 | 20 | 30.5500 | 49.7130 | -0.3124077 | -0.9499481 | 0.9499481 | -0.3124077 | -0.0100000 |
| 19 | 20 | 32.4750 | 12.8680 | 0.6385090 | 0.7696144 | -0.7696144 | 0.6385090 | -1.0000000 |
| 20 | 20 | 32.6380 | 12.6650 | 0.6371729 | 0.7707209 | -0.7707209 | 0.6371729 | $-1.0000000$ |
| 21 | 20 | 32.7980 | 12.4680 | 0.6510249 | 0.7590564 | -0.7590564 | 0.6510249 | -1.0000000 |
| 22 | 20 | 32.9520 | 12.2850 | 0.6624232 | 0.7491298 | -0.7491298 | 0.6624232 | $-1.0000000$ |
| 23 | 20 | 33.0980 | 12.1140 | 0.6617146 | 0.7497558 | -0.7497558 | 0.6617146 | -1.0000000 |
| 24 | 20 | 33.2310 | 11.9680 | 0.6851623 | 0.7283905 | -0.7283905 | 0.6851623 | -1.0000000 |
| 25 | 20 | 33.3830 | 11.8020 | 0.6672414 | 0.7448416 | -0.7448416 | 0.6672414 | -1.0000000 |
| 26 | 20 | 33.5650 | 11.5940 | 0.6818820 | 0.7314622 | -0.7314622 | 0.6818820 | $-1.0000000$ |
| 27 | 20 | 33.7520 | 11.3860 | 0.6925446 | 0.7213750 | -0.7213750 | 0.6925446 | -1.0000000 |
| 28 | 20 | 33.9430 | 11.1790 | 0.7040743 | 0.7101264 | -0.7101264 | 0.7040743 | -1.0000000 |
| 29 | 20 | 34.1400 | 10.9740 | 0.7174609 | 0.6965988 | -0.6965988 | 0.7174609 | -1.0000000 |


| 30 | 20 | 34.3450 | 10.7690 | 0.7325891 | 0.6806712 | -0.6806712 | 0.7325891 | $-1.0000000$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 31 | 20 | 34.5580 | 10.5650 | 0.7494920 | 0.6620134 | -0.6620134 | 0.7494920 | $-1.0000000$ |
| 32 | 20 | 34.7830 | 10.3610 | 0.7699677 | 0.6380828 | -0.6380828 | 0.7699677 | -1.0000000 |
| 33 | 20 | 35.0220 | 10.1560 | 0.8023614 | 0.5968385 | -0.5968385 | 0.8023614 | -1.0000000 |
| 43 | 20 | 35.3170 | 49.7430 | 0.8284003 | -0.5601366 | 0.5601366 | 0.8284003 | -1.0000000 |
| 44 | 20 | 35.3170 | 49.7430 | 0.8284003 | -0.5601366 | 0.5601366 | 0.8284003 | -0.0100000 |
| 4 | 21 | 30.3750 | 49.8820 | 0.9863912 | -0.1644152 | 0.1644152 | 0.9863912 | -1.0000000 |
| 5 | 21 | 30.3750 | 49.8820 | 0.9863912 | -0.1644152 | 0.1644152 | 0.9863912 | -0.0100000 |
| 18 | 21 | 32.7090 | 13.4270 | 0.6519520 | 0.7582602 | -0.7582602 | 0.6519520 | -1.0000000 |
| 19 | 21 | 32.9520 | 13.1570 | 0.6593421 | 0.7518431 | -0.7518431 | 0.6593421 | -1.0000000 |
| 21 | 21 | 33.2170 | 12.8520 | 0.6780766 | 0.7349913 | -0.7349913 | 0.6780766 | -1.0000000 |
| 22 | 21 | 33.3730 | 12.6790 | 0.6838822 | 0.7295924 | -0.7295924 | 0.6838822 | -1.0000000 |
| 23 | 21 | 33.5210 | 12.5160 | 0.6903108 | 0.7235130 | -0.7235130 | 0.6903108 | -1.0000000 |
| 24 | 21 | 33.6440 | 12.3740 | 0.6787784 | 0.7343432 | -0.7343432 | 0.6787784 | $-1.0000000$ |
| 25 | 21 | 33.7870 | 12.2060 | 0.6995152 | 0.7146177 | -0.7146177 | 0.6995152 | -1.0000000 |
| 26 | 21 | 33.9720 | 11.9990 | 0.6889566 | 0.7248026 | -0.7248026 | 0.6889566 | -1.0000000 |
| 27 | 21 | 34.1560 | 11.7970 | 0.6961728 | 0.7178743 | -0.7178743 | 0.6961728 | -1.0000000 |
| 28 | 21 | 34.3420 | 11.5990 | 0.7108852 | 0.7033080 | -0.7033080 | 0.7108852 | -1.0000000 |
| 29 | 21 | 34.5320 | 11.4060 | 0.7291807 | 0.6843212 | -0.6843212 | 0.7291807 | -1.0000000 |
| 30 | 21 | 34.7290 | 11.2170 | 0.7495099 | 0.6619932 | -0.6619932 | 0.7495099 | -1.0000000 |
| 31 | 21 | 34.9350 | 11.0310 | 0.7704629 | 0.6374848 | -0.6374848 | 0.7704629 | $-1.0000000$ |
| 32 | 21 | 35.1500 | 10.8490 | 0.7892618 | 0.6140569 | -0.6140569 | 0.7892618 | -1.0000000 |
| 33 | 21 | 35.3770 | 10.6590 | 0.7836223 | 0.6212376 | -0.6212376 | 0.7836223 | -1.0000000 |
| 43 | 21 | 34.8950 | 50.1180 | 0.5809758 | -0.8139208 | 0.8139208 | 0.5809758 | -1.0000000 |
| 44 | 21 | 34.8950 | 50.1180 | 0.5809758 | -0.8139208 | 0.8139208 | 0.5809758 | -0.0100000 |
| 4 | 22 | 30.2000 | 50.3880 | 0.7447912 | -0.6672976 | 0.6672976 | 0.7447912 | -1.0000000 |
| 5 | 22 | 30.2000 | 50.3880 | 0.7447912 | -0.6672976 | 0.6672976 | 0.7447912 | -0.0100000 |
| 17 | 22 | 32.8900 | 13.8500 | 0.6364682 | 0.7713029 | -0.7713029 | 0.6364682 | -1.0000000 |
| 18 | 22 | 33.1200 | 13.7650 | 0.6364682 | 0.7713029 | -0.7713029 | 0.6364682 | -1.0000000 |
| 20 | 22 | 33.4420 | 13.4010 | 0.6886922 | 0.7250539 | -0.7250539 | 0.6886922 | -1.0000000 |
| 21 | 22 | 33.6060 | 13.2280 | 0.6969191 | 0.7171497 | -0.7171497 | 0.6969191 | -1.0000000 |
| 22 | 22 | 33.7660 | 13.0610 | 0.6951142 | 0.7188994 | -0.7188994 | 0.6951142 | -1.0000000 |
| 23 | 22 | 33.9190 | 12.8980 | 0.6794225 | 0.7337473 | -0.7337473 | 0.6794225 | -1.0000000 |
| 24 | 22 | 34.0440 | 12.7670 | 0.6844263 | 0.7290820 | -0.7290820 | 0.6844263 | -1.0000000 |
| 25 | 22 | 34.1880 | 12.6080 | 0.6661321 | 0.7458338 | -0.7458338 | 0.6661321 | -1.0000000 |
| 26 | 22 | 34.3720 | 12.4000 | 0.6896967 | 0.7240984 | -0.7240984 | 0.6896967 | -1.0000000 |
| 27 | 22 | 34.5530 | 12.2070 | 0.7088799 | 0.7053292 | -0.7053292 | 0.7088799 | -1.0000000 |
| 28 | 22 | 34.7340 | 12.0240 | 0.7266873 | 0.6869684 | -0.6869684 | 0.7266873 | -1.0000000 |
| 29 | 22 | 34.9180 | 11.8470 | 0.7453201 | 0.6667067 | -0.6667067 | 0.7453201 | -1.0000000 |
| 30 | 22 | 35.1070 | 11.6770 | 0.7664203 | 0.6423393 | -0.6423393 | 0.7664203 | -1.0000000 |
| 31 | 22 | 35.3040 | 11.5130 | 0.7911461 | 0.6116272 | -0.6116272 | 0.7911461 | -1.0000000 |
| 32 | 22 | 35.5120 | 11.3520 | 0.8066753 | 0.5909949 | -0.5909949 | 0.8066753 | -1.0000000 |
| 43 | 22 | 34.6950 | 50.4870 | 0.8634763 | 0.5043895 | -0.5043895 | 0.8634763 | -1.0000000 |
| 44 | 22 | 34.6950 | 50.4870 | 0.8634763 | 0.5043895 | -0.5043895 | 0.8634763 | -0.0100000 |
| 4 | 23 | 29.6880 | 50.6380 | 0.3606009 | -0.9327202 | 0.9327202 | 0.3606009 | -1.0000000 |
| 5 | 23 | 29.6880 | 50.6380 | 0.3606009 | -0.9327202 | 0.9327202 | 0.3606009 | -0.0100000 |
| 17 | 23 | 33.2247 | 14.1757 | 0.7236159 | 0.6902029 | -0.6902029 | 0.7236159 | -1.0000000 |
| 21 | 23 | 33.9500 | 13.5840 | 0.7322029 | 0.6810866 | -0.6810866 | 0.7322029 | -1.0000000 |
| 22 | 23 | 34.1220 | 13.4230 | 0.7334517 | 0.6797416 | -0.6797416 | 0.7334517 | -1.0000000 |
| 23 | 23 | 34.3010 | 13.2580 | 0.7152382 | 0.6988808 | -0.6988808 | 0.7152382 | -1.0000000 |
| 24 | 23 | 34.4430 | 13.1230 | 0.6844263 | 0.7290820 | -0.7290820 | 0.6844263 | -1.0000000 |
| 25 | 23 | 34.5820 | 12.9800 | 0.6942698 | 0.7197148 | -0.7197148 | 0.6942698 | -1.0000000 |
| 26 | 23 | 34.7590 | 12.8030 | 0.7317309 | 0.6815936 | -0.6815936 | 0.7317309 | -1.0000000 |
| 27 | 23 | 34.9410 | 12.6350 | 0.7507888 | 0.6605423 | -0.6605423 | 0.7507888 | -1.0000000 |
| 28 | 23 | 35.1240 | 12.4730 | 0.7616604 | 0.6479764 | -0.6479764 | 0.7616604 | -1.0000000 |
| 29 | 23 | 35.3060 | 12.3160 | 0.7709812 | 0.6368579 | -0.6368579 | 0.7709812 | -1.0000000 |
| 30 | 23 | 35.4810 | 12.1610 | 0.7696230 | 0.6384986 | -0.6384986 | 0.7696230 | -1.0000000 |
| 31 | 23 | 35.6650 | 12.0210 | 0.8373985 | 0.5465928 | -0.5465928 | 0.8373985 | -1.0000000 |
| 32 | 23 | 35.8800 | 11.8680 | 0.8015182 | 0.5979704 | -0.5979704 | 0.8015182 | -1.0000000 |
| 43 | 23 | 35.0130 | 50.9500 | 0.7714882 | 0.6362436 | -0.6362436 | 0.7714882 | -1.0000000 |
| 44 | 23 | 35.0130 | 50.9500 | 0.7714882 | 0.6362436 | -0.6362436 | 0.7714882 | -0.0100000 |
| 4 | 24 | 29.4130 | 51.0250 | 0.9983510 | 0.0574046 | -0.0574046 | 0.9983510 | -1.0000000 |
| 5 | 24 | 29.4130 | 51.0250 | 0.9983510 | 0.0574046 | -0.0574046 | 0.9983510 | -0.0100000 |
| 17 | 24 | 33.4497 | 14.4707 | 0.8075448 | 0.5898062 | -0.5898062 | 0.8075448 | -1.0000000 |
| 21 | 24 | 34.2400 | 13.9280 | 0.7868346 | 0.6171640 | -0.6171640 | 0.7868346 | -1.0000000 |
| 23 | 24 | 34.6170 | 13.6320 | 0.8127608 | 0.5825975 | -0.5825975 | 0.8127608 | -1.0000000 |
| 24 | 24 | 34.7800 | 13.5030 | 0.8036001 | 0.5951697 | -0.5951697 | 0.8036001 | -1.0000000 |
| 25 | 24 | 34.9290 | 13.3880 | 0.8087633 | 0.5881342 | -0.5881342 | 0.8087633 | -1.0000000 |
| 26 | 24 | 35.1090 | 13.2560 | 0.8265285 | 0.5628948 | -0.5628948 | 0.8265285 | -1.0000000 |
| 27 | 24 | 35.3000 | 13.1210 | 0.8316516 | 0.5552979 | -0.5552979 | 0.8316516 | -1.0000000 |
| 28 | 24 | 35.4970 | 12.9830 | 0.8277945 | 0.5610314 | -0.5610314 | 0.8277945 | -1.0000000 |


| 29 | 24 | 35.6950 | 12.8400 | 0.8147732 | 0.5797799 | -0.5797799 | 0.8147732 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 43 | 24 | 35.0620 | 51.4750 | 0.8646075 | -0.5024480 | 0.5024480 | 0.8646075 | -1.0000000 |
| 44 | 24 | 35.0620 | 51.4750 | 0.8646075 | -0.5024480 | 0.5024480 | 0.8646075 | -0.0100000 |
| 4 | 25 | 29.5250 | 51.6750 | 0.9723064 | 0.2337099 | -0.2337099 | 0.9723064 | -1.0000000 |
| 5 | 25 | 29.5250 | 51.6750 | 0.9723064 | 0.2337099 | -0.2337099 | 0.9723064 | -0.0100000 |
| 17 | 25 | 33.6377 | 14.7597 | 0.8662348 | 0.4996371 | -0.4996371 | 0.8662348 | -1.0000000 |
| 21 | 25 | 34.4940 | 14.3080 | 0.8562559 | 0.5165517 | -0.5165517 | 0.8562559 | -1.0000000 |
| 23 | 25 | 34.8740 | 14.0700 | 0.8670021 | 0.4983045 | -0.4983045 | 0.8670021 | $-1.0000000$ |
| 24 | 25 | 35.0300 | 13.9730 | 0.8902680 | 0.4554370 | -0.4554370 | 0.8902680 | -1.0000000 |
| 25 | 25 | 35.1910 | 13.8870 | 0.9153073 | 0.4027561 | -0.4027561 | 0.9153073 | -1.0000000 |
| 26 | 25 | 35.3920 | 13.7880 | 0.9077492 | 0.4195133 | -0.4195133 | 0.9077492 | $-1.0000000$ |
| 27 | 25 | 35.6030 | 13.6830 | 0.9032380 | 0.4291400 | -0.4291400 | 0.9032380 | $-1.0000000$ |
| 28 | 25 | 35.8240 | 13.5710 | 0.8965612 | 0.4429198 | -0.4429198 | 0.8965612 | $-1.0000000$ |
| 29 | 25 | 36.0530 | 13.4450 | 0.8772830 | 0.4799734 | -0.4799734 | 0.8772830 | -1.0000000 |
| 30 | 25 | 36.2680 | 13.3140 | 0.8692204 | 0.4944249 | -0.4944249 | 0.8692204 | $-1.0000000$ |
| 31 | 25 | 36.4930 | 13.1870 | 0.8786447 | 0.4774762 | -0.4774762 | 0.8786447 | $-1.0000000$ |
| 43 | 25 | 34.7870 | 52.0370 | 0.9402708 | -0.3404275 | 0.3404275 | 0.9402708 | -1.0000000 |
| 44 | 25 | 34.7870 | 52.0370 | 0.9402708 | -0.3404275 | 0.3404275 | 0.9402708 | -0.0100000 |
| 4 | 26 | 29.5750 | 52.3120 | 0.9997522 | 0.0222634 | -0.0222634 | 0.9997522 | $-1.0000000$ |
| 5 | 26 | 29.5750 | 52.3120 | 0.9997522 | 0.0222634 | -0.0222634 | 0.9997522 | -0.0100000 |
| 17 | 26 | 33.8107 | 15.1217 | 0.9381382 | 0.3462611 | -0.3462611 | 0.9381382 | $-1.0000000$ |
| 21 | 26 | 34.7160 | 14.7630 | 0.9232709 | 0.3841495 | -0.3841495 | 0.9232709 | -1.0000000 |
| 23 | 26 | 35.1210 | 14.5860 | 0.9174718 | 0.3978008 | -0.3978008 | 0.9174718 | $-1.0000000$ |
| 24 | 26 | 35.2680 | 14.5190 | 0.9019989 | 0.4317383 | -0.4317383 | 0.9019989 | -1.0000000 |
| 25 | 26 | 35.4180 | 14.4530 | 0.9265593 | 0.3761488 | -0.3761488 | 0.9265593 | -1.0000000 |
| 26 | 26 | 35.6240 | 14.3720 | 0.9356194 | 0.3530105 | -0.3530105 | 0.9356194 | $-1.0000000$ |
| 27 | 26 | 35.8470 | 14.2840 | 0.9350733 | 0.3544543 | -0.3544543 | 0.9350733 | $-1.0000000$ |
| 28 | 26 | 36.0900 | 14.1820 | 0.9233244 | 0.3840209 | -0.3840209 | 0.9233244 | -1.0000000 |
| 43 | 26 | 34.7830 | 52.6120 | 0.9404899 | 0.3398216 | -0.3398216 | 0.9404899 | -1.0000000 |
| 44 | 26 | 34.7830 | 52.6120 | 0.9404899 | 0.3398216 | -0.3398216 | 0.9404899 | -0.0100000 |
| 4 | 27 | 29.6750 | 52.9150 | 0.8787842 | 0.4772194 | -0.4772194 | 0.8787842 | $-1.0000000$ |
| 5 | 27 | 29.6750 | 52.9150 | 0.8787842 | 0.4772194 | -0.4772194 | 0.8787842 | -0.0100000 |
| 18 | 27 | 34.0660 | 15.4620 | 0.9706644 | 0.2404383 | -0.2404383 | 0.9706644 | -1.0000000 |
| 19 | 27 | 34.4100 | 15.3990 | 0.9702239 | 0.2422097 | -0.2422097 | 0.9702239 | -1.0000000 |
| 20 | 27 | 34.6530 | 15.3360 | 0.9679702 | 0.2510650 | -0.2510650 | 0.9679702 | $-1.0000000$ |
| 21 | 27 | 34.8910 | 15.2720 | 0.9626060 | 0.2709052 | -0.2709052 | 0.9626060 | -1.0000000 |
| 22 | 27 | 35.1140 | 15.2040 | 0.9514159 | 0.3079087 | -0.3079087 | 0.9514159 | -1.0000000 |
| 23 | 27 | 35.3180 | 15.1350 | 0.9560024 | 0.2933590 | -0.2933590 | 0.9560024 | $-1.0000000$ |
| 24 | 27 | 35.4800 | 15.0890 | 0.9700117 | 0.2430585 | -0.2430585 | 0.9700117 | $-1.0000000$ |
| 25 | 27 | 35.6360 | 15.0470 | 0.9527801 | 0.3036612 | -0.3036612 | 0.9527801 | -1.0000000 |
| 26 | 27 | 35.8380 | 14.9830 | 0.9488931 | 0.3155977 | -0.3155977 | 0.9488931 | -1.0000000 |
| 27 | 27 | 36.0690 | 14.9040 | 0.9422109 | 0.3350203 | -0.3350203 | 0.9422109 | $-1.0000000$ |
| 28 | 27 | 36.3320 | 14.8060 | 0.9331436 | 0.3595040 | -0.3595040 | 0.9331436 | $-1.0000000$ |
| 29 | 27 | 36.6450 | 14.6860 | 0.9374290 | 0.3481765 | -0.3481765 | 0.9374290 | -1.0000000 |
| 43 | 27 | 35.0450 | 53.2100 | 0.9357268 | 0.3527257 | -0.3527257 | 0.9357268 | -1.0000000 |
| 44 | 27 | 35.0450 | 53.2100 | 0.9357268 | 0.3527257 | -0.3527257 | 0.9357268 | -0.0100000 |
| 4 | 28 | 30.0500 | 53.4900 | 0.7936341 | 0.6083954 | -0.6083954 | 0.7936341 | -1.0000000 |
| 5 | 28 | 30.0500 | 53.4900 | 0.7936341 | 0.6083954 | -0.6083954 | 0.7936341 | -0.0100000 |
| 22 | 28 | 35.2250 | 15.7510 | 0.9874628 | 0.1578516 | -0.1578516 | 0.9874628 | -1.0000000 |
| 23 | 28 | 35.4310 | 15.7070 | 0.9862506 | 0.1652567 | -0.1652567 | 0.9862506 | $-1.0000000$ |
| 24 | 28 | 35.5980 | 15.6730 | 0.9897601 | 0.1427409 | -0.1427409 | 0.9897601 | $-1.0000000$ |
| 25 | 28 | 35.7840 | 15.6390 | 0.9853291 | 0.1706653 | -0.1706653 | 0.9853291 | -1.0000000 |
| 26 | 28 | 36.0250 | 15.5860 | 0.9666363 | 0.2561528 | -0.2561528 | 0.9666363 | $-1.0000000$ |
| 27 | 28 | 36.2880 | 15.5140 | 0.9506565 | 0.3102454 | -0.3102454 | 0.9506565 | $-1.0000000$ |
| 28 | 28 | 36.5820 | 15.4250 | 0.9371049 | 0.3490480 | -0.3490480 | 0.9371049 | $-1.0000000$ |
| 29 | 28 | 36.9210 | 15.3210 | 0.9268980 | 0.3753133 | -0.3753133 | 0.9268980 | -1.0000000 |
| 30 | 28 | 37.3610 | 15.1850 | 0.9321541 | 0.3620618 | -0.3620618 | 0.9321541 | $-1.0000000$ |
| 43 | 28 | 34.9950 | 53.7980 | 0.8825015 | -0.4703096 | 0.4703096 | 0.8825015 | -1.0000000 |
| 44 | 28 | 34.9950 | 53.7980 | 0.8825015 | -0.4703096 | 0.4703096 | 0.8825015 | -0.0100000 |
| 4 | 29 | 30.0500 | 54.0220 | 0.7538046 | -0.6570987 | 0.6570987 | 0.7538046 | -1.0000000 |
| 5 | 29 | 30.0500 | 54.0220 | 0.7538046 | -0.6570987 | 0.6570987 | 0.7538046 | -0.0100000 |
| 21 | 29 | 35.0460 | 16.3610 | 0.9899104 | 0.1416951 | -0.1416951 | 0.9899104 | -1.0000000 |
| 22 | 29 | 35.2720 | 16.3200 | 0.9935336 | 0.1135388 | -0.1135388 | 0.9935336 | -1.0000000 |
| 23 | 29 | 35.4870 | 16.2850 | 0.9935947 | 0.1130023 | -0.1130023 | 0.9935947 | $-1.0000000$ |
| 24 | 29 | 35.6540 | 16.2440 | 0.9725293 | 0.2327804 | -0.2327804 | 0.9725293 | $-1.0000000$ |
| 25 | 29 | 35.8660 | 16.1990 | 0.9910555 | 0.1334501 | -0.1334501 | 0.9910555 | $-1.0000000$ |
| 26 | 29 | 36.1630 | 16.1490 | 0.9801720 | 0.1981484 | -0.1981484 | 0.9801720 | $-1.0000000$ |
| 27 | 29 | 36.4670 | 16.0830 | 0.9695137 | 0.2450373 | -0.2450373 | 0.9695137 | $-1.0000000$ |
| 28 | 29 | 36.7940 | 15.9980 | 0.9611259 | 0.2761104 | -0.2761104 | 0.9611259 | $-1.0000000$ |
| 29 | 29 | 37.1640 | 15.8960 | 0.9567010 | 0.2910723 | -0.2910723 | 0.9567010 | $-1.0000000$ |
| 30 | 29 | 37.6020 | 15.7760 | 0.9561771 | 0.2927890 | -0.2927890 | 0.9561771 | -1.0000000 |


| 31 | 29 | 38.0160 | 15.6620 | 0.9516390 | 0.3072185 | -0.3072185 | 0.9516390 | $-1.0000000$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 43 | 29 | 34.4200 | 54.2130 | 0.4679558 | -0.8837519 | 0.8837519 | 0.4679558 | $-1.0000000$ |
| 44 | 29 | 34.4200 | 54.2130 | 0.4679558 | -0.8837519 | 0.8837519 | 0.4679558 | -0.0100000 |
| 4 | 30 | 29.5250 | 54.1780 | -0.2899929 | -0.9570288 | 0.9570288 | -0.2899929 | -1.0000000 |
| 5 | 30 | 29.5250 | 54.1780 | -0.2899929 | -0.9570288 | 0.9570288 | -0.2899929 | -0.0100000 |
| 20 | 30 | 34.8800 | 16.9390 | 0.9999987 | 0.0016130 | -0.0016130 | 0.9999987 | -1.0000000 |
| 21 | 30 | 35.1040 | 16.9260 | 0.9865689 | 0.1633455 | -0.1633455 | 0.9865689 | -1.0000000 |
| 22 | 30 | 35.3220 | 16.8770 | 0.9859585 | 0.1669906 | -0.1669906 | 0.9859585 | -1.0000000 |
| 23 | 30 | 35.5360 | 16.8280 | 0.9851987 | 0.1714165 | -0.1714165 | 0.9851987 | -1.0000000 |
| 24 | 30 | 35.7160 | 16.7890 | 0.9885377 | 0.1509742 | -0.1509742 | 0.9885377 | -1.0000000 |
| 25 | 30 | 35.9510 | 16.7560 | 0.9887211 | 0.1497684 | -0.1497684 | 0.9887211 | -1.0000000 |
| 26 | 30 | 36.2720 | 16.7060 | 0.9837881 | 0.1793347 | -0.1793347 | 0.9837881 | -1.0000000 |
| 27 | 30 | 36.5980 | 16.6400 | 0.9774381 | 0.2112221 | -0.2112221 | 0.9774381 | -1.0000000 |
| 28 | 30 | 36.9410 | 16.5570 | 0.9724447 | 0.2331337 | -0.2331337 | 0.9724447 | -1.0000000 |
| 29 | 30 | 37.3180 | 16.4550 | 0.9701402 | 0.2425450 | -0.2425450 | 0.9701402 | -1.0000000 |
| 30 | 30 | 37.7540 | 16.3220 | 0.9670946 | 0.2544171 | -0.2544171 | 0.9670946 | -1.0000000 |
| 43 | 30 | 33.8120 | 54.5130 | 0.4910846 | -0.8711119 | 0.8711119 | 0.4910846 | -1.0000000 |
| 44 | 30 | 33.8120 | 54.5130 | 0.4910846 | -0.8711119 | 0.8711119 | 0.4910846 | -0.0100000 |
| 4 | 31 | 28.9870 | 53.8800 | -0.5232630 | -0.8521712 | 0.8521712 | -0.5232630 | -1.0000000 |
| 5 | 31 | 28.9870 | 53.8800 | -0.5232630 | -0.8521712 | 0.8521712 | -0.5232630 | -0.0100000 |
| 20 | 31 | 34.9820 | 17.5810 | 0.9830278 | 0.1834570 | -0.1834570 | 0.9830278 | -1.0000000 |
| 21 | 31 | 35.2120 | 17.5380 | 0.9775684 | 0.2106183 | -0.2106183 | 0.9775684 | -1.0000000 |
| 22 | 31 | 35.4350 | 17.4890 | 0.9737473 | 0.2276317 | -0.2276317 | 0.9737473 | -1.0000000 |
| 23 | 31 | 35.6490 | 17.4380 | 0.9705318 | 0.2409729 | -0.2409729 | 0.9705318 | -1.0000000 |
| 24 | 31 | 35.8350 | 17.4110 | 0.9936789 | 0.1122602 | -0.1122602 | 0.9936789 | -1.0000000 |
| 25 | 31 | 36.0740 | 17.3720 | 0.9738554 | 0.2271687 | -0.2271687 | 0.9738554 | -1.0000000 |
| 26 | 31 | 36.3940 | 17.2920 | 0.9718764 | 0.2354914 | -0.2354914 | 0.9718764 | -1.0000000 |
| 27 | 31 | 36.7325 | 17.2055 | 0.9718764 | 0.2354914 | -0.2354914 | 0.9718764 | -1.0000000 |
| 28 | 31 | 37.0710 | 17.1190 | 0.9698592 | 0.2436661 | -0.2436661 | 0.9698592 | -1.0000000 |
| 29 | 31 | 37.4400 | 17.0190 | 0.9711381 | 0.2385179 | -0.2385179 | 0.9711381 | -1.0000000 |
| 30 | 31 | 37.8740 | 16.8850 | 0.9633242 | 0.2683400 | -0.2683400 | 0.9633242 | -1.0000000 |
| 43 | 31 | 33.2500 | 54.8450 | 0.4646240 | -0.8855081 | 0.8855081 | 0.4646240 | -1.0000000 |
| 44 | 31 | 33.2500 | 54.8450 | 0.4646240 | -0.8855081 | 0.8855081 | 0.4646240 | -0.0100000 |
| 4 | 32 | 28.5380 | 53.8750 | 0.7834641 | -0.6214371 | 0.6214371 | 0.7834641 | -1.0000000 |
| 5 | 32 | 28.5380 | 53.8750 | 0.7834641 | -0.6214371 | 0.6214371 | 0.7834641 | -0.0100000 |
| 20 | 32 | 35.1530 | 18.2500 | 0.9514315 | 0.3078605 | -0.3078605 | 0.9514315 | -1.0000000 |
| 21 | 32 | 35.3750 | 18.1810 | 0.9604740 | 0.2783697 | -0.2783697 | 0.9604740 | -1.0000000 |
| 22 | 32 | 35.5930 | 18.1220 | 0.9672033 | 0.2540035 | -0.2540035 | 0.9672033 | -1.0000000 |
| 23 | 32 | 35.8090 | 18.0700 | 0.9713605 | 0.2376105 | -0.2376105 | 0.9713605 | -1.0000000 |
| 24 | 32 | 35.9910 | 18.0330 | 0.9756268 | 0.2194364 | -0.2194364 | 0.9756268 | -1.0000000 |
| 25 | 32 | 36.2220 | 17.9650 | 0.9522054 | 0.3054585 | -0.3054585 | 0.9522054 | -1.0000000 |
| 26 | 32 | 36.5390 | 17.8620 | 0.9614001 | 0.2751543 | -0.2751543 | 0.9614001 | -1.0000000 |
| 28 | 32 | 37.2300 | 17.7025 | 0.9695262 | 0.2449876 | -0.2449876 | 0.9695262 | -1.0000000 |
| 29 | 32 | 37.5940 | 17.6030 | 0.9695262 | 0.2449876 | -0.2449876 | 0.9695262 | -1.0000000 |
| 30 | 32 | 38.0000 | 17.4900 | 0.9695262 | 0.2449876 | -0.2449876 | 0.9695262 | -1.0000000 |
| 43 | 32 | 32.8530 | 55.3200 | 0.9633588 | -0.2682160 | 0.2682160 | 0.9633588 | -1.0000000 |
| 44 | 32 | 32.8530 | 55.3200 | 0.9633588 | -0.2682160 | 0.2682160 | 0.9633588 | -0.0100000 |
| 4 | 33 | 28.4130 | 54.3880 | 0.9856443 | 0.1688350 | -0.1688350 | 0.9856443 | -1.0000000 |
| 5 | 33 | 28.4130 | 54.3880 | 0.9856443 | 0.1688350 | -0.1688350 | 0.9856443 | -0.0100000 |
| 19 | 33 | 35.0860 | 18.8500 | 0.9633446 | 0.2682669 | -0.2682669 | 0.9633446 | -1.0000000 |
| 20 | 33 | 35.3140 | 18.7800 | 0.9545438 | 0.2980707 | -0.2980707 | 0.9545438 | -1.0000000 |
| 21 | 33 | 35.5260 | 18.7090 | 0.9571682 | 0.2895324 | -0.2895324 | 0.9571682 | -1.0000000 |
| 22 | 33 | 35.7350 | 18.6440 | 0.9609579 | 0.2766946 | -0.2766946 | 0.9609579 | -1.0000000 |
| 23 | 33 | 35.9460 | 18.5830 | 0.9640154 | 0.2658465 | -0.2658465 | 0.9640154 | -1.0000000 |
| 24 | 33 | 36.1160 | 18.5280 | 0.9557030 | 0.2943330 | -0.2943330 | 0.9557030 | -1.0000000 |
| 25 | 33 | 36.3360 | 18.4590 | 0.9766065 | 0.2150341 | -0.2150341 | 0.9766065 | -1.0000000 |
| 26 | 33 | 36.6510 | 18.3800 | 0.9823801 | 0.1868940 | -0.1868940 | 0.9823801 | -1.0000000 |
| 28 | 33 | 37.3100 | 18.2750 | 0.9908106 | 0.1352568 | -0.1352568 | 0.9908106 | -1.0000000 |
| 29 | 33 | 37.7090 | 18.1700 | 0.9908106 | 0.1352568 | -0.1352568 | 0.9908106 | -1.0000000 |
| 30 | 33 | 38.0600 | 18.0650 | 0.9583178 | 0.2857045 | -0.2857045 | 0.9583178 | -1.0000000 |
| 43 | 33 | 32.7320 | 55.8750 | 0.9634417 | -0.2679183 | 0.2679183 | 0.9634417 | -1.0000000 |
| 44 | 33 | 32.7320 | 55.8750 | 0.9634417 | -0.2679183 | 0.2679183 | 0.9634417 | -0.0100000 |
| 4 | 34 | 28.2370 | 54.9470 | 0.7143571 | -0.6997814 | 0.6997814 | 0.7143571 | -1.0000000 |
| 5 | 34 | 28.2370 | 54.9470 | 0.7143571 | -0.6997814 | 0.6997814 | 0.7143571 | -0.0100000 |
| 19 | 34 | 35.2020 | 19.3090 | 0.9746724 | 0.2236374 | -0.2236374 | 0.9746724 | -1.0000000 |
| 20 | 34 | 35.4380 | 19.2470 | 0.9665051 | 0.2566476 | -0.2566476 | 0.9665051 | -1.0000000 |
| 21 | 34 | 35.6550 | 19.1850 | 0.9637472 | 0.2668169 | -0.2668169 | 0.9637472 | -1.0000000 |
| 22 | 34 | 35.8620 | 19.1240 | 0.9642811 | 0.2648811 | -0.2648811 | 0.9642811 | -1.0000000 |
| 23 | 34 | 36.0640 | 19.0650 | 0.9684821 | 0.2490834 | -0.2490834 | 0.9684821 | -1.0000000 |
| 24 | 34 | 36.2210 | 19.0100 | 0.9492237 | 0.3146019 | -0.3146019 | 0.9492237 | -1.0000000 |
| 25 | 34 | 36.4150 | 18.9590 | 0.9859970 | 0.1667632 | -0.1667632 | 0.9859970 | -1.0000000 |


| 26 | 34 | 36.7000 | 18.9170 | 0.9974847 | 0.0708827 | -0.0708827 | 0.9974847 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 28 | 34 | 37.3405 | 18.8320 | 0.9842117 | 0.1769952 | -0.1769952 | 0.9842117 | $-1.0000000$ |
| 29 | 34 | 37.7710 | 18.7470 | 0.9842117 | 0.1769952 | -0.1769952 | 0.9842117 | -1.0000000 |
| 30 | 34 | 38.1590 | 18.6400 | 0.9710875 | 0.2387238 | -0.2387238 | 0.9710875 | -1.0000000 |
| 43 | 34 | 32.3220 | 56.0250 | -0.1465520 | -0.9892030 | 0.9892030 | -0.1465520 | $-1.0000000$ |
| 44 | 34 | 32.3220 | 56.0250 | -0.1465520 | -0.9892030 | 0.9892030 | -0.1465520 | -0.0100000 |
| 4 | 35 | 27.6750 | 55.3350 | 0.3562477 | -0.9343916 | 0.9343916 | 0.3562477 | -1.0000000 |
| 5 | 35 | 27.6750 | 55.3350 | 0.3562477 | -0.9343916 | 0.9343916 | 0.3562477 | -0.0100000 |
| 19 | 35 | 35.2790 | 19.7730 | 0.9900692 | 0.1405811 | -0.1405811 | 0.9900692 | $-1.0000000$ |
| 20 | 35 | 35.5350 | 19.7310 | 0.9873909 | 0.1583007 | -0.1583007 | 0.9873909 | $-1.0000000$ |
| 21 | 35 | 35.7580 | 19.6940 | 0.9878280 | 0.1555502 | -0.1555502 | 0.9878280 | -1.0000000 |
| 22 | 35 | 35.9620 | 19.6610 | 0.9898052 | 0.1424280 | -0.1424280 | 0.9898052 | -1.0000000 |
| 23 | 35 | 36.1500 | 19.6320 | 0.9928368 | 0.1194780 | -0.1194780 | 0.9928368 | $-1.0000000$ |
| 24 | 35 | 36.3020 | 19.5940 | 0.9732487 | 0.2297542 | -0.2297542 | 0.9732487 | $-1.0000000$ |
| 25 | 35 | 36.4570 | 19.5540 | 0.9968849 | 0.0788707 | -0.0788707 | 0.9968849 | -1.0000000 |
| 26 | 35 | 36.6710 | 19.5300 | 0.9998183 | -0.0190645 | 0.0190645 | 0.9998183 | -1.0000000 |
| 27 | 35 | 36.9940 | 19.5020 | 0.9999999 | 0.0004214 | -0.0004214 | 0.9999999 | -1.0000000 |
| 28 | 35 | 37.4240 | 19.4570 | 0.9961948 | 0.0871549 | -0.0871549 | 0.9961948 | -1.0000000 |
| 29 | 35 | 37.8660 | 19.3930 | 0.9851460 | 0.1717187 | -0.1717187 | 0.9851460 | -1.0000000 |
| 30 | 35 | 38.2780 | 19.3100 | 0.9802942 | 0.1975432 | -0.1975432 | 0.9802942 | -1.0000000 |
| 43 | 35 | 31.6420 | 55.9750 | 0.1915226 | -0.9814882 | 0.9814882 | 0.1915226 | -1.0000000 |
| 44 | 35 | 31.6420 | 55.9750 | 0.1915226 | -0.9814882 | 0.9814882 | 0.1915226 | -0.0100000 |
| 4 | 36 | 27.0750 | 55.5880 | 0.3271701 | -0.9449654 | 0.9449654 | 0.3271701 | -1.0000000 |
| 19 | 36 | 35.3340 | 20.2400 | 0.9949861 | 0.1000137 | -0.1000137 | 0.9949861 | -1.0000000 |
| 20 | 36 | 35.5860 | 20.2200 | 0.9981602 | 0.0606322 | -0.0606322 | 0.9981602 | $-1.0000000$ |
| 21 | 36 | 35.8080 | 20.2080 | 0.9990488 | 0.0436072 | -0.0436072 | 0.9990488 | -1.0000000 |
| 22 | 36 | 36.0140 | 20.1990 | 0.9989846 | 0.0450535 | -0.0450535 | 0.9989846 | -1.0000000 |
| 23 | 36 | 36.2080 | 20.1930 | 0.9972126 | 0.0746120 | -0.0746120 | 0.9972126 | -1.0000000 |
| 24 | 36 | 36.3670 | 20.1770 | 0.9842031 | 0.1770432 | -0.1770432 | 0.9842031 | $-1.0000000$ |
| 25 | 36 | 36.5080 | 20.1570 | 0.9914514 | 0.1304767 | -0.1304767 | 0.9914514 | -1.0000000 |
| 26 | 36 | 36.6870 | 20.1450 | 0.9923452 | 0.1234950 | -0.1234950 | 0.9923452 | -1.0000000 |
| 27 | 36 | 37.0000 | 20.1205 | 0.9961115 | 0.0881021 | -0.0881021 | 0.9961115 | -1.0000000 |
| 28 | 36 | 37.4850 | 20.0960 | 0.9961115 | 0.0881021 | -0.0881021 | 0.9961115 | $-1.0000000$ |
| 29 | 36 | 37.9660 | 20.0630 | 0.9942500 | 0.1070836 | -0.1070836 | 0.9942500 | -1.0000000 |
| 30 | 36 | 38.4020 | 20.0280 | 0.9928061 | 0.1197332 | -0.1197332 | 0.9928061 | -1.0000000 |
| 43 | 36 | 31.4300 | 56.3200 | 0.9572459 | 0.2892753 | -0.2892753 | 0.9572459 | -1.0000000 |
| 44 | 36 | 31.4300 | 56.3200 | 0.9572459 | 0.2892753 | -0.2892753 | 0.9572459 | -0.0100000 |
| 4 | 37 | 26.5500 | 55.7000 | 0.0000000 | -1.0000000 | 1.0000000 | 0.0000000 | -1.0000000 |
| 5 | 37 | 26.4870 | 56.1500 | 0.0166598 | -0.9998612 | 0.9998612 | 0.0166598 | $-1.0000000$ |
| 6 | 37 | 26.5250 | 56.8380 | 0.1019175 | -0.9947929 | 0.9947929 | 0.1019175 | $-1.0000000$ |
| 7 | 37 | 26.5500 | 57.3500 | 0.0000000 | -1.0000000 | 1.0000000 | 0.0000000 | -1.0000000 |
| 20 | 37 | 35.5990 | 20.6940 | 0.9999930 | 0.0037393 | -0.0037393 | 0.9999930 | -1.0000000 |
| 21 | 37 | 35.8190 | 20.6940 | 0.9999934 | -0.0036145 | 0.0036145 | 0.9999934 | -1.0000000 |
| 22 | 37 | 36.0300 | 20.6930 | 0.9999002 | 0.0141315 | -0.0141315 | 0.9999002 | $-1.0000000$ |
| 23 | 37 | 36.2390 | 20.6870 | 0.9996251 | 0.0273795 | -0.0273795 | 0.9996251 | -1.0000000 |
| 24 | 37 | 36.4060 | 20.6820 | 1.0000000 | 0.0000000 | 0.0000000 | 1.0000000 | -1.0000000 |
| 25 | 37 | 36.5660 | 20.6830 | 0.9998916 | 0.0147209 | -0.0147209 | 0.9998916 | -1.0000000 |
| 26 | 37 | 36.7520 | 20.6830 | 0.9996842 | 0.0251325 | -0.0251325 | 0.9996842 | -1.0000000 |
| 27 | 37 | 36.9220 | 20.6830 | 0.9996842 | 0.0251325 | -0.0251325 | 0.9996842 | -1.0000000 |
| 28 | 37 | 37.1320 | 20.6830 | 0.9996842 | 0.0251325 | -0.0251325 | 0.9996842 | -1.0000000 |
| 30 | 37 | 37.6580 | 20.6780 | 0.9998260 | 0.0186575 | -0.0186575 | 0.9998260 | -1.0000000 |
| 31 | 37 | 38.0310 | 20.6760 | 0.9984464 | 0.0557206 | -0.0557206 | 0.9984464 | -1.0000000 |
| 43 | 37 | 31.7180 | 56.8820 | 0.8378644 | 0.5458784 | -0.5458784 | 0.8378644 | -1.0000000 |
| 44 | 37 | 31.7180 | 56.8820 | 0.8378644 | 0.5458784 | -0.5458784 | 0.8378644 | -0.0100000 |
| 4 | 38 | 25.9620 | 55.7500 | 0.1554819 | -0.9878387 | 0.9878387 | 0.1554819 | -1.0000000 |
| 7 | 38 | 26.2500 | 57.4750 | 0.3690618 | -0.9294049 | 0.9294049 | 0.3690618 | $-1.0000000$ |
| 8 | 38 | 26.2500 | 57.4750 | 0.3690618 | -0.9294049 | 0.9294049 | 0.3690618 | -0.0100000 |
| 21 | 38 | 35.8290 | 21.2390 | 0.9999892 | 0.0046539 | -0.0046539 | 0.9999892 | -1.0000000 |
| 22 | 38 | 36.0370 | 21.2370 | 0.9996812 | 0.0252500 | -0.0252500 | 0.9996812 | $-1.0000000$ |
| 23 | 38 | 36.2410 | 21.2290 | 0.9997902 | 0.0204827 | -0.0204827 | 0.9997902 | $-1.0000000$ |
| 24 | 38 | 36.4050 | 21.2250 | 0.9999996 | -0.0008606 | 0.0008606 | 0.9999996 | -1.0000000 |
| 25 | 38 | 36.5680 | 21.2260 | 0.9997911 | -0.0204365 | 0.0204365 | 0.9997911 | -1.0000000 |
| 26 | 38 | 36.7570 | 21.2290 | 0.9998496 | -0.0173443 | 0.0173443 | 0.9998496 | $-1.0000000$ |
| 27 | 38 | 36.9220 | 21.2290 | 0.9989399 | 0.0460324 | -0.0460324 | 0.9989399 | $-1.0000000$ |
| 28 | 38 | 37.1320 | 21.2270 | 0.9980754 | 0.0620122 | -0.0620122 | 0.9980754 | -1.0000000 |
| 29 | 38 | 37.3980 | 21.2250 | 0.9993327 | 0.0365269 | -0.0365269 | 0.9993327 | -1.0000000 |
| 30 | 38 | 37.6960 | 21.2270 | 0.9994706 | 0.0325357 | -0.0325357 | 0.9994706 | -1.0000000 |
| 31 | 38 | 38.0620 | 21.2330 | 0.9999999 | 0.0002705 | -0.0002705 | 0.9999999 | $-1.0000000$ |
| 32 | 38 | 38.4240 | 21.2390 | 0.9997694 | -0.0214729 | 0.0214729 | 0.9997694 | -1.0000000 |
| 43 | 38 | 32.1250 | 57.4870 | 0.8012773 | 0.5982931 | -0.5982931 | 0.8012773 | -1.0000000 |
| 44 | 38 | 32.1250 | 57.4870 | 0.8012773 | 0.5982931 | -0.5982931 | 0.8012773 | -0.0100000 |


| 4 | 39 | 25.2700 | 55.8500 | 0.1554819 | -0.9878387 | 0.9878387 | 0.1554819 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 39 | 25.2700 | 55.8500 | 0.1554819 | -0.9878387 | 0.9878387 | 0.1554819 | -1.0000000 |
| 7 | 39 | 25.8400 | 57.8250 | 0.4584127 | -0.8887394 | 0.8887394 | 0.4584127 | -1.0000000 |
| 8 | 39 | 25.8400 | 57.8250 | 0.4584127 | -0.8887394 | 0.8887394 | 0.4584127 | -0.0100000 |
| 22 | 39 | 36.0400 | 21.8140 | 0.9998346 | 0.0181869 | -0.0181869 | 0.9998346 | -1.0000000 |
| 23 | 39 | 36.2410 | 21.8070 | 0.9999337 | 0.0115199 | -0.0115199 | 0.9999337 | -1.0000000 |
| 24 | 39 | 36.4050 | 21.8050 | 1.0000000 | 0.0000000 | 0.0000000 | 1.0000000 | -1.0000000 |
| 25 | 39 | 36.5600 | 21.8070 | 0.9999191 | -0.0127214 | 0.0127214 | 0.9999191 | -1.0000000 |
| 26 | 39 | 36.7490 | 21.8110 | 0.9999030 | -0.0139281 | 0.0139281 | 0.9999030 | -1.0000000 |
| 27 | 39 | 36.9470 | 21.8130 | 0.9999957 | 0.0029346 | -0.0029346 | 0.9999957 | -1.0000000 |
| 28 | 39 | 37.1680 | 21.8120 | 0.9999439 | 0.0105946 | -0.0105946 | 0.9999439 | -1.0000000 |
| 29 | 39 | 37.4190 | 21.8100 | 0.9999998 | -0.0006578 | 0.0006578 | 0.9999998 | -1.0000000 |
| 30 | 39 | 37.7120 | 21.8120 | 0.9998754 | -0.0157842 | 0.0157842 | 0.9998754 | -1.0000000 |
| 31 | 39 | 38.0650 | 21.8160 | 0.9998900 | -0.0148300 | 0.0148300 | 0.9998900 | -1.0000000 |
| 32 | 39 | 38.4240 | 21.8160 | 0.9998900 | -0.0148300 | 0.0148300 | 0.9998900 | -1.0000000 |
| 43 | 39 | 32.6380 | 57.9380 | 0.6455021 | 0.7637585 | -0.7637585 | 0.6455021 | -1.0000000 |
| 44 | 39 | 32.6380 | 57.9380 | 0.6455021 | 0.7637585 | -0.7637585 | 0.6455021 | -0.0100000 |
| 4 | 40 | 24.6080 | 55.9900 | 0.1554819 | -0.9878387 | 0.9878387 | 0.1554819 | -1.0000000 |
| 5 | 40 | 24.6080 | 55.9900 | 0.1554819 | -0.9878387 | 0.9878387 | 0.1554819 | -1.0000000 |
| 7 | 40 | 25.3150 | 58.0000 | -0.1525558 | -0.9882948 | 0.9882948 | -0.1525558 | -1.0000000 |
| 8 | 40 | 25.3150 | 58.0000 | -0.1525558 | -0.9882948 | 0.9882948 | -0.1525558 | -0.0100000 |
| 23 | 40 | 36.2410 | 22.3970 | 0.9999999 | -0.0004132 | 0.0004132 | 0.9999999 | -1.0000000 |
| 24 | 40 | 36.4050 | 22.3970 | 0.9999999 | -0.0004132 | 0.0004132 | 0.9999999 | -1.0000000 |
| 25 | 40 | 36.5570 | 22.4010 | 0.9997835 | -0.0208103 | 0.0208103 | 0.9997835 | -1.0000000 |
| 26 | 40 | 36.7440 | 22.4060 | 0.9998882 | -0.0149534 | 0.0149534 | 0.9998882 | -1.0000000 |
| 27 | 40 | 36.9470 | 22.4080 | 0.9999945 | -0.0033126 | 0.0033126 | 0.9999945 | -1.0000000 |
| 28 | 40 | 37.1710 | 22.4060 | 0.9999669 | 0.0081378 | -0.0081378 | 0.9999669 | -1.0000000 |
| 29 | 40 | 37.4200 | 22.4010 | 0.9998735 | 0.0159050 | -0.0159050 | 0.9998735 | -1.0000000 |
| 30 | 40 | 37.7080 | 22.3920 | 0.9997147 | 0.0238873 | -0.0238873 | 0.9997147 | -1.0000000 |
| 31 | 40 | 38.0610 | 22.3670 | 0.9988315 | 0.0483286 | -0.0483286 | 0.9988315 | -1.0000000 |
| 43 | 40 | 32.7320 | 58.2750 | 0.7821845 | -0.6230469 | 0.6230469 | 0.7821845 | -1.0000000 |
| 44 | 40 | 32.7320 | 58.2750 | 0.7821845 | -0.6230469 | 0.6230469 | 0.7821845 | -0.0100000 |
| 4 | 41 | 23.9750 | 55.9700 | 0.0000000 | -1.0000000 | 1.0000000 | 0.0000000 | -1.0000000 |
| 5 | 41 | 23.9750 | 55.9700 | 0.0000000 | -1.0000000 | 1.0000000 | 0.0000000 | -0.0100000 |
| 7 | 41 | 24.7880 | 58.0250 | 0.4718579 | -0.8816746 | 0.8816746 | 0.4718579 | -1.0000000 |
| 8 | 41 | 24.7880 | 58.0250 | 0.4718579 | -0.8816746 | 0.8816746 | 0.4718579 | -0.0100000 |
| 23 | 41 | 36.2400 | 23.0530 | 0.9999999 | -0.0003536 | 0.0003536 | 0.9999999 | -1.0000000 |
| 24 | 41 | 36.3980 | 23.0530 | 0.9999577 | -0.0091924 | 0.0091924 | 0.9999577 | -1.0000000 |
| 25 | 41 | 36.5510 | 23.0520 | 0.9999986 | -0.0016686 | 0.0016686 | 0.9999986 | -1.0000000 |
| 26 | 41 | 36.7500 | 23.0470 | 0.9994124 | 0.0342764 | -0.0342764 | 0.9994124 | -1.0000000 |
| 28 | 41 | 37.1930 | 23.0190 | 0.9971045 | 0.0760432 | -0.0760432 | 0.9971045 | -1.0000000 |
| 29 | 41 | 37.4440 | 22.9980 | 0.9967363 | 0.0807263 | -0.0807263 | 0.9967363 | -1.0000000 |
| 30 | 41 | 37.7310 | 22.9740 | 0.9971915 | 0.0748933 | -0.0748933 | 0.9971915 | -1.0000000 |
| 31 | 41 | 38.0970 | 22.9200 | 0.9875750 | 0.1571484 | -0.1571484 | 0.9875750 | -1.0000000 |
| 43 | 41 | 32.2320 | 58.7000 | 0.6852555 | -0.7283028 | 0.7283028 | 0.6852555 | -1.0000000 |
| 44 | 41 | 32.2320 | 58.7000 | 0.6852555 | -0.7283028 | 0.7283028 | 0.6852555 | -0.0100000 |
| 4 | 42 | 23.7630 | 56.2630 | 0.9549969 | 0.2966157 | -0.2966157 | 0.9549969 | -1.0000000 |
| 5 | 42 | 23.7630 | 56.2630 | 0.9549969 | 0.2966157 | -0.2966157 | 0.9549969 | -0.0100000 |
| 7 | 42 | 24.8120 | 58.2250 | 0.7428958 | 0.6694070 | -0.6694070 | 0.7428958 | -1.0000000 |
| 8 | 42 | 24.8120 | 58.2250 | 0.7428958 | 0.6694070 | -0.6694070 | 0.7428958 | -0.0100000 |
| 22 | 42 | 36.0260 | 23.7720 | 0.9954726 | 0.0950491 | -0.0950491 | 0.9954726 | -1.0000000 |
| 23 | 42 | 36.2590 | 23.7600 | 0.9996480 | 0.0265299 | -0.0265299 | 0.9996480 | -1.0000000 |
| 24 | 42 | 36.4100 | 23.7470 | 0.9907788 | 0.1354895 | -0.1354895 | 0.9907788 | -1.0000000 |
| 25 | 42 | 36.5780 | 23.7220 | 0.9942727 | 0.1068730 | -0.1068730 | 0.9942727 | -1.0000000 |
| 26 | 42 | 36.8030 | 23.6930 | 0.9900991 | 0.1403699 | -0.1403699 | 0.9900991 | -1.0000000 |
| 28 | 42 | 37.2720 | 23.6180 | 0.9834660 | 0.1810931 | -0.1810931 | 0.9834660 | -1.0000000 |
| 29 | 42 | 37.5260 | 23.5750 | 0.9819075 | 0.1893611 | -0.1893611 | 0.9819075 | -1.0000000 |
| 30 | 42 | 37.8060 | 23.5300 | 0.9843979 | 0.1759569 | -0.1759569 | 0.9843979 | -1.0000000 |
| 43 | 42 | 32.0130 | 59.2750 | 0.9538864 | 0.3001678 | -0.3001678 | 0.9538864 | -1.0000000 |
| 44 | 42 | 32.0130 | 59.2750 | 0.9538864 | 0.3001678 | -0.3001678 | 0.9538864 | -0.0100000 |
| 4 | 43 | 23.8000 | 56.8750 | 0.9944351 | -0.1053515 | 0.1053515 | 0.9944351 | -1.0000000 |
| 5 | 43 | 23.8000 | 56.8750 | 0.9944351 | -0.1053515 | 0.1053515 | 0.9944351 | -0.0100000 |
| 7 | 43 | 25.1120 | 58.5500 | 0.9885548 | 0.1508618 | -0.1508618 | 0.9885548 | -1.0000000 |
| 8 | 43 | 25.1120 | 58.5500 | 0.9885548 | 0.1508618 | -0.1508618 | 0.9885548 | -0.0100000 |
| 23 | 43 | 36.3340 | 24.4160 | 0.9958323 | 0.0912032 | -0.0912032 | 0.9958323 | -1.0000000 |
| 24 | 43 | 36.4860 | 24.3910 | 0.9547542 | 0.2973960 | -0.2973960 | 0.9547542 | -1.0000000 |
| 25 | 43 | 36.6650 | 24.3430 | 0.9835840 | 0.1804507 | -0.1804507 | 0.9835840 | -1.0000000 |
| 26 | 43 | 36.9070 | 24.2960 | 0.9797590 | 0.2001808 | -0.2001808 | 0.9797590 | -1.0000000 |
| 27 | 43 | 37.1480 | 24.2450 | 0.9768392 | 0.2139748 | -0.2139748 | 0.9768392 | -1.0000000 |
| 28 | 43 | 37.3940 | 24.1910 | 0.9750853 | 0.2218301 | -0.2218301 | 0.9750853 | -1.0000000 |
| 29 | 43 | 37.6480 | 24.1320 | 0.9750735 | 0.2218819 | -0.2218819 | 0.9750735 | -1.0000000 |


| 30 | 43 | 37.9170 | 24.0700 | 0.9777054 | 0.2099816 | -0.2099816 | 0.9777054 | -1.0000000 |
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| 31 | 43 | 38.1940 | 23.9880 | 0.9594585 | 0.2818498 | -0.2818498 | 0.9594585 | -1.0000000 |
| 43 | 43 | 32.4700 | 59.6750 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -1.0000000 |
| 44 | 43 | 32.4700 | 59.6750 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -0.0100000 |
| 4 | 44 | 23.4130 | 57.3120 | 0.2847662 | -0.9585970 | 0.9585970 | 0.2847662 | -1.0000000 |
| 5 | 44 | 23.4130 | 57.3120 | 0.2847662 | -0.9585970 | 0.9585970 | 0.2847662 | -0.0100000 |
| 7 | 44 | 25.0380 | 58.8500 | 0.7447901 | -0.6672989 | 0.6672989 | 0.7447901 | -1.0000000 |
| 8 | 44 | 25.0380 | 58.8500 | 0.7447901 | -0.6672989 | 0.6672989 | 0.7447901 | -0.0100000 |
| 23 | 44 | 36.4540 | 24.9960 | 0.9691049 | 0.2466490 | -0.2466490 | 0.9691049 | -1.0000000 |
| 24 | 44 | 36.6050 | 24.9650 | 0.9867981 | 0.1619553 | -0.1619553 | 0.9867981 | -1.0000000 |
| 25 | 44 | 36.7910 | 24.9330 | 0.9751245 | 0.2216581 | -0.2216581 | 0.9751245 | -1.0000000 |
| 26 | 44 | 37.0370 | 24.8810 | 0.9746800 | 0.2236042 | -0.2236042 | 0.9746800 | -1.0000000 |
| 27 | 44 | 37.2830 | 24.8250 | 0.9731268 | 0.2302702 | -0.2302702 | 0.9731268 | -1.0000000 |
| 28 | 44 | 37.5330 | 24.7670 | 0.9715608 | 0.2367900 | -0.2367900 | 0.9715608 | -1.0000000 |
| 29 | 44 | 37.7900 | 24.7070 | 0.9699782 | 0.2431921 | -0.2431921 | 0.9699782 | -1.0000000 |
| 30 | 44 | 38.0520 | 24.6410 | 0.9651112 | 0.2618403 | -0.2618403 | 0.9651112 | -1.0000000 |
| 31 | 44 | 38.3340 | 24.5910 | 0.9651112 | 0.2618403 | -0.2618403 | 0.9651112 | -1.0000000 |
| 33 | 44 | 0.0000 | 0.0000 | 1.0000000 | 0.0000000 | 0.0000000 | 1.0000000 | -1.0000000 |
| 34 | 44 | 0.0000 | 0.0000 | 1.0000000 | 0.0000000 | 0.0000000 | 1.0000000 | -1.0000000 |
| 35 | 44 | 0.0000 | 0.0000 | 1.0000000 | 0.0000000 | 0.0000000 | 1.0000000 | -1.0000000 |
| 43 | 44 | 33.1400 | 59.8380 | 0.2684468 | 0.9632945 | -0.9632945 | 0.2684468 | -1.0000000 |
| 44 | 44 | 33.1400 | 59.8380 | 0.2684468 | 0.9632945 | -0.9632945 | 0.2684468 | -0.0100000 |
| 4 | 45 | 22.7750 | 57.4500 | 0.1380637 | -0.9904233 | 0.9904233 | 0.1380637 | -1.0000000 |
| 5 | 45 | 22.7750 | 57.4500 | 0.1380637 | -0.9904233 | 0.9904233 | 0.1380637 | -0.0100000 |
| 7 | 45 | 24.7500 | 59.0370 | 0.5658044 | -0.8245395 | 0.8245395 | 0.5658044 | -1.0000000 |
| 8 | 45 | 24.7500 | 59.0370 | 0.5658044 | -0.8245395 | 0.8245395 | 0.5658044 | -0.0100000 |
| 22 | 45 | 36.3250 | 25.6660 | 0.9862451 | 0.1652894 | -0.1652894 | 0.9862451 | -1.0000000 |
| 23 | 45 | 36.5770 | 25.6100 | 0.9654331 | 0.2606511 | -0.2606511 | 0.9654331 | -1.0000000 |
| 24 | 45 | 36.7300 | 25.5700 | 0.9898514 | 0.1421067 | -0.1421067 | 0.9898514 | -1.0000000 |
| 25 | 45 | 36.9220 | 25.5380 | 0.9816907 | 0.1904819 | -0.1904819 | 0.9816907 | -1.0000000 |
| 26 | 45 | 37.1750 | 25.4860 | 0.9779080 | 0.2090356 | -0.2090356 | 0.9779080 | -1.0000000 |
| 27 | 45 | 37.4230 | 25.4330 | 0.9776682 | 0.2101546 | -0.2101546 | 0.9776682 | -1.0000000 |
| 28 | 45 | 37.6700 | 25.3810 | 0.9800057 | 0.1989692 | -0.1989692 | 0.9800057 | -1.0000000 |
| 29 | 45 | 37.9240 | 25.3270 | 0.9826248 | 0.1856028 | -0.1856028 | 0.9826248 | -1.0000000 |
| 30 | 45 | 38.1720 | 25.2770 | 0.9826248 | 0.1856028 | -0.1856028 | 0.9826248 | -1.0000000 |
| 43 | 45 | 33.6700 | 60.2130 | 0.6968868 | 0.7171811 | -0.7171811 | 0.6968868 | -1.0000000 |
| 44 | 45 | 33.6700 | 60.2130 | 0.6968868 | 0.7171811 | -0.7171811 | 0.6968868 | -0.0100000 |
| 4 | 46 | 22.2000 | 57.6950 | 0.3963514 | -0.9180989 | 0.9180989 | 0.3963514 | -1.0000000 |
| 5 | 46 | 22.2000 | 57.6950 | 0.3963514 | -0.9180989 | 0.9180989 | 0.3963514 | -0.0100000 |
| 7 | 46 | 24.4130 | 59.0880 | 0.1706729 | -0.9853277 | 0.9853277 | 0.1706729 | -1.0000000 |
| 8 | 46 | 24.4130 | 59.0880 | 0.1706729 | -0.9853277 | 0.9853277 | 0.1706729 | -0.0100000 |
| 22 | 46 | 36.4310 | 26.2920 | 0.9851137 | 0.1719037 | -0.1719037 | 0.9851137 | -1.0000000 |
| 23 | 46 | 36.6960 | 26.2360 | 0.9772605 | 0.2120425 | -0.2120425 | 0.9772605 | -1.0000000 |
| 24 | 46 | 36.8690 | 26.2010 | 0.9822542 | 0.1875545 | -0.1875545 | 0.9822542 | -1.0000000 |
| 25 | 46 | 37.0520 | 26.1730 | 0.9820345 | 0.1887013 | -0.1887013 | 0.9820345 | -1.0000000 |
| 26 | 46 | 37.3100 | 26.1330 | 0.9848358 | 0.1734891 | -0.1734891 | 0.9848358 | -1.0000000 |
| 27 | 46 | 37.5590 | 26.0970 | 0.9860556 | 0.1664166 | -0.1664166 | 0.9860556 | -1.0000000 |
| 28 | 46 | 37.7930 | 26.0550 | 0.9820862 | 0.1884321 | -0.1884321 | 0.9820862 | -1.0000000 |
| 43 | 46 | 34.1750 | 60.6380 | 0.5301558 | 0.8479002 | -0.8479002 | 0.5301558 | -1.0000000 |
| 44 | 46 | 34.1750 | 60.6380 | 0.5301558 | 0.8479002 | -0.8479002 | 0.5301558 | -0.0100000 |
| 4 | 47 | 21.6120 | 57.9200 | 0.1464948 | -0.9892114 | 0.9892114 | 0.1464948 | -1.0000000 |
| 5 | 47 | 21.6120 | 57.9200 | 0.1464948 | -0.9892114 | 0.9892114 | 0.1464948 | -0.0100000 |
| 7 | 47 | 24.2630 | 59.3750 | 0.9562553 | -0.2925336 | 0.2925336 | 0.9562553 | -1.0000000 |
| 8 | 47 | 24.2630 | 59.3750 | 0.9562553 | -0.2925336 | 0.2925336 | 0.9562553 | -0.0100000 |
| 22 | 47 | 36.5060 | 26.8930 | 0.9929683 | 0.1183803 | -0.1183803 | 0.9929683 | -1.0000000 |
| 23 | 47 | 36.7830 | 26.8490 | 0.9917006 | 0.1285683 | -0.1285683 | 0.9917006 | -1.0000000 |
| 24 | 47 | 36.9760 | 26.8380 | 0.9980386 | -0.0626007 | 0.0626007 | 0.9980386 | -1.0000000 |
| 25 | 47 | 37.1510 | 26.8430 | 0.9975356 | 0.0701625 | -0.0701625 | 0.9975356 | -1.0000000 |
| 26 | 47 | 37.3890 | 26.8300 | 0.9992295 | 0.0392482 | -0.0392482 | 0.9992295 | -1.0000000 |
| 27 | 47 | 37.6300 | 26.8190 | 0.9997009 | 0.0244566 | -0.0244566 | 0.9997009 | -1.0000000 |
| 43 | 47 | 34.5500 | 60.5750 | -0.9122413 | 0.4096533 | -0.4096533 | -0.9122413 | -1.0000000 |
| 44 | 47 | 34.5500 | 60.5750 | -0.9122413 | 0.4096533 | -0.4096533 | -0.9122413 | -0.0100000 |
| 4 | 48 | 20.9380 | 58.0750 | 0.4587257 | -0.8885779 | 0.8885779 | 0.4587257 | -1.0000000 |
| 5 | 48 | 20.9380 | 58.0750 | 0.4587257 | -0.8885779 | 0.8885779 | 0.4587257 | -0.0100000 |
| 7 | 48 | 24.2120 | 59.9500 | 0.8217237 | -0.5698861 | 0.5698861 | 0.8217237 | -1.0000000 |
| 8 | 48 | 24.2120 | 59.9500 | 0.8217237 | -0.5698861 | 0.5698861 | 0.8217237 | -0.0100000 |
| 21 | 48 | 36.2130 | 27.4910 | 0.9973906 | -0.0721946 | 0.0721946 | 0.9973906 | -1.0000000 |
| 22 | 48 | 36.5080 | 27.5060 | 0.9982136 | -0.0597467 | 0.0597467 | 0.9982136 | -1.0000000 |
| 23 | 48 | 36.7850 | 27.5160 | 0.9985209 | -0.0543693 | 0.0543693 | 0.9985209 | -1.0000000 |
| 24 | 48 | 36.9750 | 27.5260 | 0.9949672 | -0.1002008 | 0.1002008 | 0.9949672 | -1.0000000 |
| 25 | 48 | 37.1510 | 27.5360 | 0.9986563 | -0.0518219 | 0.0518219 | 0.9986563 | -1.0000000 |

Appendix B

| 26 | 48 | 37.3900 | 27.5410 | 0.9997076 | -0.0241826 | 0.0241826 | 0.9997076 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27 | 48 | 37.6310 | 27.5470 | 0.9998791 | -0.0155474 | 0.0155474 | 0.9998791 | -1.0000000 |
| 28 | 48 | 37.8770 | 27.5540 | 0.9998484 | -0.0174091 | 0.0174091 | 0.9998484 | -1.0000000 |
| 43 | 48 | 34.5880 | 59.9500 | -0.9965926 | -0.0824812 | 0.0824812 | -0.9965926 | -1.0000000 |
| 44 | 48 | 34.5880 | 59.9500 | -0.9965926 | -0.0824812 | 0.0824812 | -0.9965926 | -0.0100000 |
| 4 | 49 | 20.3370 | 58.4250 | 0.7024830 | -0.7117006 | 0.7117006 | 0.7024830 | -1.0000000 |
| 5 | 49 | 20.3370 | 58.4250 | 0.7024830 | -0.7117006 | 0.7117006 | 0.7024830 | -0.0100000 |
| 7 | 49 | 23.8880 | 60.1750 | -0.1740423 | -0.9847382 | 0.9847382 | -0.1740423 | -1.0000000 |
| 8 | 49 | 23.8880 | 60.1750 | -0.1740423 | -0.9847382 | 0.9847382 | -0.1740423 | -0.0100000 |
| 21 | 49 | 36.1360 | 28.1210 | 0.9899129 | -0.1416776 | 0.1416776 | 0.9899129 | -1.0000000 |
| 22 | 49 | 36.4310 | 28.1580 | 0.9909456 | -0.1342640 | 0.1342640 | 0.9909456 | -1.0000000 |
| 23 | 49 | 36.7150 | 28.1890 | 0.9932561 | -0.1159412 | 0.1159412 | 0.9932561 | -1.0000000 |
| 24 | 49 | 36.9120 | 28.2040 | 0.9987404 | -0.0501760 | 0.0501760 | 0.9987404 | -1.0000000 |
| 25 | 49 | 37.1010 | 28.2130 | 0.9976012 | -0.0692236 | 0.0692236 | 0.9976012 | -1.0000000 |
| 26 | 49 | 37.3590 | 28.2280 | 0.9984771 | -0.0551680 | 0.0551680 | 0.9984771 | -1.0000000 |
| 27 | 49 | 37.6050 | 28.2400 | 0.9982271 | -0.0595206 | 0.0595206 | 0.9982271 | -1.0000000 |
| 28 | 49 | 37.8380 | 28.2630 | 0.9915454 | -0.1297605 | 0.1297605 | 0.9915454 | -1.0000000 |
| 43 | 49 | 34.7870 | 59.3500 | -0.7189829 | 0.6950278 | -0.6950278 | -0.7189829 | -1.0000000 |
| 44 | 49 | 34.7870 | 59.3500 | -0.7189829 | 0.6950278 | -0.6950278 | -0.7189829 | -0.0100000 |
| 4 | 50 | 19.8880 | 58.9320 | 0.7616763 | -0.6479577 | 0.6479577 | 0.7616763 | -1.0000000 |
| 5 | 50 | 19.8880 | 58.9320 | 0.7616763 | -0.6479577 | 0.6479577 | 0.7616763 | -0.0100000 |
| 7 | 50 | 23.6250 | 59.9380 | -0.8744090 | -0.4851896 | 0.4851896 | -0.8744090 | -1.0000000 |
| 8 | 50 | 23.6250 | 59.9380 | -0.8744090 | -0.4851896 | 0.4851896 | -0.8744090 | -0.0100000 |
| 21 | 50 | 36.0180 | 28.8060 | 0.9882227 | -0.1530227 | 0.1530227 | 0.9882227 | -1.0000000 |
| 22 | 50 | 36.3280 | 28.8420 | 0.9919519 | -0.1266151 | 0.1266151 | 0.9919519 | -1.0000000 |
| 23 | 50 | 36.6230 | 28.8740 | 0.9926128 | -0.1213249 | 0.1213249 | 0.9926128 | -1.0000000 |
| 24 | 50 | 36.8290 | 28.8890 | 0.9973666 | -0.0725246 | 0.0725246 | 0.9973666 | -1.0000000 |
| 25 | 50 | 37.0350 | 28.9000 | 0.9950978 | -0.0988962 | 0.0988962 | 0.9950978 | -1.0000000 |
| 26 | 50 | 37.3020 | 28.9190 | 0.9965158 | -0.0834041 | 0.0834041 | 0.9965158 | -1.0000000 |
| 27 | 50 | 37.5370 | 28.9340 | 0.9957656 | -0.0919290 | 0.0919290 | 0.9957656 | -1.0000000 |
| 28 | 50 | 37.7570 | 28.9340 | 0.9957656 | -0.0919290 | 0.0919290 | 0.9957656 | -1.0000000 |
| 43 | 50 | 35.2870 | 59.0880 | 0.0203985 | 0.9997919 | -0.9997919 | 0.0203985 | -1.0000000 |
| 44 | 50 | 35.2870 | 59.0880 | 0.0203985 | 0.9997919 | -0.9997919 | 0.0203985 | -0.0100000 |
| 4 | 51 | 19.3750 | 59.1600 | 0.0675665 | -0.9977148 | 0.9977148 | 0.0675665 | -1.0000000 |
| 5 | 51 | 19.3750 | 59.1600 | 0.0675665 | -0.9977148 | 0.9977148 | 0.0675665 | -1.0000000 |
| 7 | 51 | 23.4380 | 59.8380 | 0.3907743 | -0.9204865 | 0.9204865 | 0.3907743 | -1.0000000 |
| 8 | 51 | 23.4380 | 59.8380 | 0.3907743 | -0.9204865 | 0.9204865 | 0.3907743 | -0.0100000 |
| 21 | 51 | 35.9270 | 29.5420 | 0.9985710 | -0.0534405 | 0.0534405 | 0.9985710 | -1.0000000 |
| 22 | 51 | 36.2490 | 29.5550 | 0.9984819 | -0.0550818 | 0.0550818 | 0.9984819 | -1.0000000 |
| 23 | 51 | 36.5400 | 29.5670 | 0.9973355 | -0.0729517 | 0.0729517 | 0.9973355 | -1.0000000 |
| 24 | 51 | 36.7470 | 29.5740 | 0.9988502 | -0.0479403 | 0.0479403 | 0.9988502 | -1.0000000 |
| 25 | 51 | 36.9660 | 29.5820 | 0.9980040 | -0.0631503 | 0.0631503 | 0.9980040 | -1.0000000 |
| 27 | 51 | 37.5020 | 29.5990 | 0.9999764 | 0.0068682 | -0.0068682 | 0.9999764 | -1.0000000 |
| 28 | 51 | 37.7220 | 29.5990 | 0.9999764 | 0.0068682 | -0.0068682 | 0.9999764 | -1.0000000 |
| 43 | 51 | 35.5750 | 59.3000 | 0.8816737 | 0.4718597 | -0.4718597 | 0.8816737 | -1.0000000 |
| 44 | 51 | 35.5750 | 59.3000 | 0.8816737 | 0.4718597 | -0.4718597 | 0.8816737 | -0.0100000 |
| 4 | 52 | 18.7750 | 59.1030 | -0.0045454 | -0.9999897 | 0.9999897 | -0.0045454 | -1.0000000 |
| 5 | 52 | 18.7750 | 59.1030 | -0.0045454 | -0.9999897 | 0.9999897 | -0.0045454 | -1.0000000 |
| 7 | 52 | 23.1250 | 60.2370 | 0.8905388 | -0.4549073 | 0.4549073 | 0.8905388 | -1.0000000 |
| 8 | 52 | 23.1250 | 60.2370 | 0.8905388 | -0.4549073 | 0.4549073 | 0.8905388 | -0.0100000 |
| 21 | 52 | 35.9040 | 30.2900 | 0.9997537 | 0.0221955 | -0.0221955 | 0.9997537 | -1.0000000 |
| 22 | 52 | 36.2260 | 30.2780 | 0.9997140 | 0.0239155 | -0.0239155 | 0.9997140 | -1.0000000 |
| 23 | 52 | 36.5080 | 30.2660 | 0.9996815 | 0.0252368 | -0.0252368 | 0.9996815 | -1.0000000 |
| 24 | 52 | 36.7150 | 30.2600 | 0.9999998 | -0.0007032 | 0.0007032 | 0.9999998 | -1.0000000 |
| 25 | 52 | 36.9470 | 30.2540 | 0.9996598 | 0.0260798 | -0.0260798 | 0.9996598 | -1.0000000 |
| 27 | 52 | 37.5160 | 30.2210 | 0.9986607 | 0.0517387 | -0.0517387 | 0.9986607 | -1.0000000 |
| 28 | 52 | 37.7360 | 30.2210 | 0.9986607 | 0.0517387 | -0.0517387 | 0.9986607 | $-1.0000000$ |
| 43 | 52 | 35.9250 | 59.5330 | -0.0291295 | 0.9995757 | -0.9995757 | -0.0291295 | -1.0000000 |
| 44 | 52 | 35.9250 | 59.5330 | -0.0291295 | 0.9995757 | -0.9995757 | -0.0291295 | -0.0100000 |
| 4 | 53 | 18.1880 | 58.9950 | -0.2255335 | -0.9742354 | 0.9742354 | -0.2255335 | -1.0000000 |
| 5 | 53 | 18.1880 | 58.9950 | -0.2255335 | -0.9742354 | 0.9742354 | -0.2255335 | -1.0000000 |
| 7 | 53 | 22.8370 | 60.8620 | 0.8944252 | -0.4472176 | 0.4472176 | 0.8944252 | -1.0000000 |
| 8 | 53 | 22.8370 | 60.8620 | 0.8944252 | -0.4472176 | 0.4472176 | 0.8944252 | -0.0100000 |
| 20 | 53 | 35.5620 | 31.0000 | 0.9977726 | 0.0667063 | -0.0667063 | 0.9977726 | -1.0000000 |
| 21 | 53 | 35.9150 | 30.9760 | 0.9997212 | 0.0236134 | -0.0236134 | 0.9997212 | -1.0000000 |
| 22 | 53 | 36.2370 | 30.9730 | 0.9999347 | 0.0114232 | -0.0114232 | 0.9999347 | -1.0000000 |
| 23 | 53 | 36.5270 | 30.9720 | 0.9995725 | 0.0292360 | -0.0292360 | 0.9995725 | -1.0000000 |
| 24 | 53 | 36.7390 | 30.9640 | 0.9969051 | 0.0786149 | -0.0786149 | 0.9969051 | -1.0000000 |
| 25 | 53 | 36.9770 | 30.9450 | 0.9972379 | 0.0742743 | -0.0742743 | 0.9972379 | -1.0000000 |
| 26 | 53 | 37.2790 | 30.9220 | 0.9977372 | 0.0672345 | -0.0672345 | 0.9977372 | -1.0000000 |
| 27 | 53 | 37.5460 | 30.9030 | 0.9978903 | 0.0649225 | -0.0649225 | 0.9978903 | -1.0000000 |


| 28 | 53 | 37.7660 | 30.9030 | 0.9978903 | 0.0649225 | -0.0649225 | 0.9978903 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 43 | 53 | 36.2870 | 59.2320 | -0.8934140 | 0.4492344 | -0.4492344 | -0.8934140 | -1.0000000 |
| 44 | 53 | 36.2870 | 59.2320 | -0.8934140 | 0.4492344 | -0.4492344 | -0.8934140 | -0.0100000 |
| 4 | 54 | 17.6250 | 58.9950 | 0.1975061 | -0.9803017 | 0.9803017 | 0.1975061 | -1.0000000 |
| 5 | 54 | 17.6250 | 58.9950 | 0.1975061 | -0.9803017 | 0.9803017 | 0.1975061 | -1.0000000 |
| 7 | 54 | 22.4750 | 61.3880 | 0.6324716 | -0.7745836 | 0.7745836 | 0.6324716 | -1.0000000 |
| 8 | 54 | 22.4750 | 61.3880 | 0.6324716 | -0.7745836 | 0.7745836 | 0.6324716 | -0.0100000 |
| 20 | 54 | 35.5520 | 31.5130 | 0.9975114 | -0.0705053 | 0.0705053 | 0.9975114 | -1.0000000 |
| 21 | 54 | 36.2370 | 31.5130 | 0.9975114 | -0.0705053 | 0.0705053 | 0.9975114 | -1.0000000 |
| 24 | 54 | 36.7770 | 31.6050 | 0.9997712 | 0.0213893 | -0.0213893 | 0.9997712 | -1.0000000 |
| 25 | 54 | 36.9980 | 31.6090 | 0.9998575 | -0.0168831 | 0.0168831 | 0.9998575 | -1.0000000 |
| 26 | 54 | 37.3010 | 31.6150 | 0.9999893 | -0.0046180 | 0.0046180 | 0.9999893 | -1.0000000 |
| 27 | 54 | 37.5840 | 31.6110 | 0.9992100 | 0.0397417 | -0.0397417 | 0.9992100 | -1.0000000 |
| 43 | 54 | 36.4250 | 58.6750 | -0.9341418 | 0.3569020 | -0.3569020 | -0.9341418 | -1.0000000 |
| 44 | 54 | 36.4250 | 58.6750 | -0.9341418 | 0.3569020 | -0.3569020 | -0.9341418 | -0.0100000 |
| 4 | 55 | 17.0750 | 59.0750 | -0.0415583 | -0.9991361 | 0.9991361 | -0.0415583 | -1.0000000 |
| 5 | 55 | 17.0750 | 59.0750 | -0.0415583 | -0.9991361 | 0.9991361 | -0.0415583 | -1.0000000 |
| 7 | 55 | 22.1380 | 61.9630 | 0.8607903 | -0.5089598 | 0.5089598 | 0.8607903 | -1.0000000 |
| 8 | 55 | 22.1380 | 61.9630 | 0.8607903 | -0.5089598 | 0.5089598 | 0.8607903 | -0.0100000 |
| 20 | 55 | 35.4850 | 31.9310 | 0.9681812 | -0.2502502 | 0.2502502 | 0.9681812 | -1.0000000 |
| 21 | 55 | 36.2370 | 31.9310 | 0.9681812 | -0.2502502 | 0.2502502 | 0.9681812 | -1.0000000 |
| 24 | 55 | 36.7140 | 32.2590 | 0.9593446 | -0.2822373 | 0.2822373 | 0.9593446 | -1.0000000 |
| 25 | 55 | 36.9090 | 32.3200 | 0.9685498 | -0.2488196 | 0.2488196 | 0.9685498 | -1.0000000 |
| 26 | 55 | 37.1960 | 32.4030 | 0.9680122 | -0.2509030 | 0.2509030 | 0.9680122 | -1.0000000 |
| 27 | 55 | 37.4790 | 32.4810 | 0.9709209 | -0.2394007 | 0.2394007 | 0.9709209 | -1.0000000 |
| 43 | 55 | 36.7000 | 58.2750 | -0.2756825 | 0.9612488 | -0.9612488 | -0.2756825 | -1.0000000 |
| 44 | 55 | 36.7000 | 58.2750 | -0.2756825 | 0.9612488 | -0.9612488 | -0.2756825 | -0.0100000 |
| 4 | 56 | 16.5870 | 58.7370 | -0.7071068 | -0.7071068 | 0.7071068 | -0.7071068 | -1.0000000 |
| 5 | 56 | 16.5870 | 58.7370 | -0.7071068 | -0.7071068 | 0.7071068 | -0.7071068 | -1.0000000 |
| 7 | 56 | 22.0500 | 62.6620 | 0.9816423 | -0.1907311 | 0.1907311 | 0.9816423 | -1.0000000 |
| 8 | 56 | 22.0500 | 62.6620 | 0.9816423 | -0.1907311 | 0.1907311 | 0.9816423 | -0.0100000 |
| 20 | 56 | 35.3380 | 32.3790 | 0.9203100 | -0.3911899 | 0.3911899 | 0.9203100 | -1.0000000 |
| 21 | 56 | 35.6330 | 32.5100 | 0.9101977 | -0.4141740 | 0.4141740 | 0.9101977 | -1.0000000 |
| 22 | 56 | 35.9310 | 32.6480 | 0.9006951 | -0.4344518 | 0.4344518 | 0.9006951 | -1.0000000 |
| 23 | 56 | 36.2470 | 32.8030 | 0.8816995 | -0.4718114 | 0.4718114 | 0.8816995 | -1.0000000 |
| 24 | 56 | 36.4610 | 32.9180 | 0.8514692 | -0.5244046 | 0.5244046 | 0.8514692 | -1.0000000 |
| 25 | 56 | 36.6480 | 33.0280 | 0.8733473 | -0.4870980 | 0.4870980 | 0.8733473 | -1.0000000 |
| 26 | 56 | 36.9210 | 33.1880 | 0.8755207 | -0.4831806 | 0.4831806 | 0.8755207 | -1.0000000 |
| 27 | 56 | 37.1750 | 33.3510 | 0.8494966 | -0.5275940 | 0.5275940 | 0.8494966 | -1.0000000 |
| 43 | 56 | 36.8750 | 58.4500 | 0.9819567 | 0.1891057 | -0.1891057 | 0.9819567 | -1.0000000 |
| 44 | 56 | 36.8750 | 58.4500 | 0.9819567 | 0.1891057 | -0.1891057 | 0.9819567 | -0.0100000 |
| 4 | 57 | 16.2350 | 58.2370 | -0.5268322 | -0.8499693 | 0.8499693 | -0.5268322 | -1.0000000 |
| 5 | 57 | 16.2350 | 58.2370 | -0.5268322 | -0.8499693 | 0.8499693 | -0.5268322 | -1.0000000 |
| 7 | 57 | 21.9620 | 63.3500 | 0.9440873 | -0.3296956 | 0.3296956 | 0.9440873 | -1.0000000 |
| 8 | 57 | 21.9620 | 63.3500 | 0.9440873 | -0.3296956 | 0.3296956 | 0.9440873 | -0.0100000 |
| 20 | 57 | 35.0890 | 32.8700 | 0.8733296 | -0.4871298 | 0.4871298 | 0.8733296 | -1.0000000 |
| 21 | 57 | 35.3580 | 33.0230 | 0.8524539 | -0.5228024 | 0.5228024 | 0.8524539 | -1.0000000 |
| 22 | 57 | 35.6130 | 33.1860 | 0.8306066 | -0.5568597 | 0.5568597 | 0.8306066 | -1.0000000 |
| 23 | 57 | 35.8610 | 33.3560 | 0.7877274 | -0.6160241 | 0.6160241 | 0.7877274 | -1.0000000 |
| 24 | 57 | 36.0460 | 33.4890 | 0.7843815 | -0.6202787 | 0.6202787 | 0.7843815 | -1.0000000 |
| 25 | 57 | 36.2370 | 33.6240 | 0.7910602 | -0.6117383 | 0.6117383 | 0.7910602 | -1.0000000 |
| 26 | 57 | 36.4760 | 33.7950 | 0.7568150 | -0.6536292 | 0.6536292 | 0.7568150 | -1.0000000 |
| 43 | 57 | 37.0000 | 59.0250 | 0.8006779 | 0.5990951 | -0.5990951 | 0.8006779 | -1.0000000 |
| 44 | 57 | 37.0000 | 59.0250 | 0.8006779 | 0.5990951 | -0.5990951 | 0.8006779 | -0.0100000 |
| 4 | 58 | 15.7350 | 58.0400 | -0.0149208 | -0.9998887 | 0.9998887 | -0.0149208 | -1.0000000 |
| 5 | 58 | 15.7350 | 58.0400 | -0.0149208 | -0.9998887 | 0.9998887 | -0.0149208 | -1.0000000 |
| 7 | 58 | 21.9380 | 64.0370 | 0.9910745 | -0.1333093 | 0.1333093 | 0.9910745 | -1.0000000 |
| 8 | 58 | 21.9380 | 64.0370 | 0.9910745 | -0.1333093 | 0.1333093 | 0.9910745 | -0.0100000 |
| 20 | 58 | 34.7000 | 33.4320 | 0.8567287 | -0.5157673 | 0.5157673 | 0.8567287 | -1.0000000 |
| 21 | 58 | 35.0120 | 33.6020 | 0.8567287 | -0.5157673 | 0.5157673 | 0.8567287 | -1.0000000 |
| 22 | 58 | 35.2340 | 33.7440 | 0.8318910 | -0.5549390 | 0.5549390 | 0.8318910 | -1.0000000 |
| 23 | 58 | 35.4490 | 33.8910 | 0.8186513 | -0.5742909 | 0.5742909 | 0.8186513 | -1.0000000 |
| 24 | 58 | 35.6230 | 34.0160 | 0.8090318 | -0.5877649 | 0.5877649 | 0.8090318 | -1.0000000 |
| 25 | 58 | 35.8050 | 34.1440 | 0.8011975 | -0.5984000 | 0.5984000 | 0.8011975 | -1.0000000 |
| 26 | 58 | 36.0190 | 34.2900 | 0.8027360 | -0.5963345 | 0.5963345 | 0.8027360 | -1.0000000 |
| 43 | 58 | 37.3250 | 58.9630 | -0.5194607 | 0.8544943 | -0.8544943 | -0.5194607 | -1.0000000 |
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| 4 | 59 | 15.1000 | 57.9850 | -0.0743757 | -0.9972303 | 0.9972303 | -0.0743757 | -1.0000000 |
| 5 | 59 | 15.1000 | 57.9850 | -0.0743757 | -0.9972303 | 0.9972303 | -0.0743757 | -1.0000000 |
| 7 | 59 | 21.9130 | 64.6130 | 0.8903607 | -0.4552558 | 0.4552558 | 0.8903607 | -1.0000000 |
| 8 | 59 | 21.9130 | 64.6130 | 0.8903607 | -0.4552558 | 0.4552558 | 0.8903607 | -0.0100000 |


| 20 | 59 | 34.4180 | 34.1500 | 0.8854092 | -0.4648124 | 0.4648124 | 0.8854092 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 21 | 59 | 34.6680 | 34.1940 | 0.8854092 | -0.4648124 | 0.4648124 | 0.8854092 | -1.0000000 |
| 22 | 59 | 34.8880 | 34.3060 | 0.8782620 | -0.4781798 | 0.4781798 | 0.8782620 | $-1.0000000$ |
| 23 | 59 | 35.1630 | 34.4620 | 0.8703567 | -0.4924219 | 0.4924219 | 0.8703567 | $-1.0000000$ |
| 24 | 59 | 35.2960 | 34.5360 | 0.8771753 | -0.4801702 | 0.4801702 | 0.8771753 | $-1.0000000$ |
| 25 | 59 | 35.4790 | 34.6370 | 0.8807117 | -0.4736528 | 0.4736528 | 0.8807117 | $-1.0000000$ |
| 26 | 59 | 35.7190 | 34.7630 | 0.8987764 | -0.4384074 | 0.4384074 | 0.8987764 | $-1.0000000$ |
| 27 | 59 | 35.9950 | 34.9160 | 0.8831446 | -0.4691010 | 0.4691010 | 0.8831446 | -1.0000000 |
| 43 | 59 | 37.6880 | 58.8250 | 0.6672976 | 0.7447912 | -0.7447912 | 0.6672976 | -1.0000000 |
| 44 | 59 | 37.6880 | 58.8250 | 0.6672976 | 0.7447912 | -0.7447912 | 0.6672976 | -0.0100000 |
| 4 | 60 | 14.5000 | 57.8450 | -0.1527375 | -0.9882668 | 0.9882668 | -0.1527375 | -1.0000000 |
| 5 | 60 | 14.5000 | 57.8450 | -0.1527375 | -0.9882668 | 0.9882668 | -0.1527375 | -1.0000000 |
| 7 | 60 | 21.6380 | 65.1370 | 0.7839124 | -0.6208714 | 0.6208714 | 0.7839124 | -1.0000000 |
| 8 | 60 | 21.6380 | 65.1370 | 0.7839124 | -0.6208714 | 0.6208714 | 0.7839124 | -0.0100000 |
| 21 | 60 | 34.3250 | 34.8170 | 0.8981718 | -0.4396447 | 0.4396447 | 0.8981718 | -1.0000000 |
| 22 | 60 | 34.6250 | 34.8170 | 0.8981718 | -0.4396447 | 0.4396447 | 0.8981718 | -1.0000000 |
| 23 | 60 | 34.9010 | 34.9680 | 0.8881603 | -0.4595336 | 0.4595336 | 0.8881603 | -1.0000000 |
| 24 | 60 | 35.0370 | 35.0430 | 0.8816512 | -0.4719016 | 0.4719016 | 0.8816512 | -1.0000000 |
| 25 | 60 | 35.2260 | 35.1390 | 0.9049813 | -0.4254515 | 0.4254515 | 0.9049813 | -1.0000000 |
| 26 | 60 | 35.4850 | 35.2570 | 0.9074348 | -0.4201930 | 0.4201930 | 0.9074348 | -1.0000000 |
| 27 | 60 | 35.7610 | 35.3860 | 0.8922241 | -0.4515930 | 0.4515930 | 0.8922241 | -1.0000000 |
| 28 | 60 | 36.0310 | 35.5220 | 0.8705022 | -0.4921645 | 0.4921645 | 0.8705022 | -1.0000000 |
| 43 | 60 | 38.0750 | 59.3380 | 0.8695643 | 0.4938197 | -0.4938197 | 0.8695643 | -1.0000000 |
| 44 | 60 | 38.0750 | 59.3380 | 0.8695643 | 0.4938197 | -0.4938197 | 0.8695643 | -0.0100000 |
| 4 | 61 | 13.9000 | 57.6750 | -0.1221857 | -0.9925073 | 0.9925073 | -0.1221857 | -1.0000000 |
| 5 | 61 | 13.9000 | 57.6750 | -0.1221857 | -0.9925073 | 0.9925073 | -0.1221857 | -1.0000000 |
| 7 | 61 | 21.1380 | 65.3500 | 0.0489322 | -0.9988021 | 0.9988021 | 0.0489322 | -1.0000000 |
| 8 | 61 | 21.1380 | 65.3500 | 0.0489322 | -0.9988021 | 0.9988021 | 0.0489322 | -0.0100000 |
| 21 | 61 | 34.1650 | 35.3040 | 0.9083694 | -0.4181688 | 0.4181688 | 0.9083694 | -1.0000000 |
| 22 | 61 | 34.3950 | 35.3040 | 0.9083694 | -0.4181688 | 0.4181688 | 0.9083694 | $-1.0000000$ |
| 23 | 61 | 34.6210 | 35.4050 | 0.9023107 | -0.4310863 | 0.4310863 | 0.9023107 | $-1.0000000$ |
| 24 | 61 | 34.8100 | 35.5070 | 0.8565553 | -0.5160553 | 0.5160553 | 0.8565553 | -1.0000000 |
| 25 | 61 | 35.0050 | 35.6050 | 0.9167880 | -0.3993741 | 0.3993741 | 0.9167880 | -1.0000000 |
| 27 | 61 | 35.5390 | 35.8280 | 0.9083964 | -0.4181101 | 0.4181101 | 0.9083964 | $-1.0000000$ |
| 28 | 61 | 35.8020 | 35.9550 | 0.8976976 | -0.4406120 | 0.4406120 | 0.8976976 | -1.0000000 |
| 29 | 61 | 36.0810 | 36.0500 | 0.8976976 | -0.4406120 | 0.4406120 | 0.8976976 | -1.0000000 |
| 43 | 61 | 38.4870 | 59.8880 | 0.6370447 | 0.7708268 | -0.7708268 | 0.6370447 | -1.0000000 |
| 44 | 61 | 38.4870 | 59.8880 | 0.6370447 | 0.7708268 | -0.7708268 | 0.6370447 | -0.0100000 |
| 4 | 62 | 13.2500 | 57.5750 | -0.0356462 | -0.9993645 | 0.9993645 | -0.0356462 | -1.0000000 |
| 5 | 62 | 13.2500 | 57.5750 | -0.0356462 | -0.9993645 | 0.9993645 | -0.0356462 | $-1.0000000$ |
| 7 | 62 | 20.5380 | 65.4250 | 0.0995512 | -0.9950324 | 0.9950324 | 0.0995512 | -1.0000000 |
| 8 | 62 | 20.5380 | 65.4250 | 0.0995512 | -0.9950324 | 0.9950324 | 0.0995512 | -0.0100000 |
| 21 | 62 | 33.9500 | 35.8500 | 0.9598657 | -0.2804601 | 0.2804601 | 0.9598657 | -1.0000000 |
| 22 | 62 | 34.1840 | 35.8500 | 0.9598657 | -0.2804601 | 0.2804601 | 0.9598657 | $-1.0000000$ |
| 23 | 62 | 34.4110 | 35.9140 | 0.9555624 | -0.2947889 | 0.2947889 | 0.9555624 | -1.0000000 |
| 24 | 62 | 34.5830 | 35.9700 | 0.9160641 | -0.4010320 | 0.4010320 | 0.9160641 | -1.0000000 |
| 25 | 62 | 34.8030 | 36.0450 | 0.9329047 | -0.3601233 | 0.3601233 | 0.9329047 | -1.0000000 |
| 27 | 62 | 35.3740 | 36.2270 | 0.9485602 | -0.3165967 | 0.3165967 | 0.9485602 | -1.0000000 |
| 28 | 62 | 35.6510 | 36.3240 | 0.9442009 | -0.3293700 | 0.3293700 | 0.9442009 | -1.0000000 |
| 29 | 62 | 35.9310 | 36.3940 | 0.9442009 | -0.3293700 | 0.3293700 | 0.9442009 | $-1.0000000$ |
| 43 | 62 | 39.0250 | 60.0880 | -0.0574046 | 0.9983510 | -0.9983510 | -0.0574046 | -1.0000000 |
| 44 | 62 | 39.0250 | 60.0880 | -0.0574046 | 0.9983510 | -0.9983510 | -0.0574046 | -0.0100000 |
| 4 | 63 | 12.5250 | 57.5920 | 0.1711236 | -0.9852496 | 0.9852496 | 0.1711236 | -1.0000000 |
| 5 | 63 | 12.5250 | 57.5920 | 0.1711236 | -0.9852496 | 0.9852496 | 0.1711236 | $-1.0000000$ |
| 7 | 63 | 19.9380 | 65.5000 | -0.1947787 | -0.9808472 | 0.9808472 | -0.1947787 | -1.0000000 |
| 8 | 63 | 19.9380 | 65.5000 | -0.1947787 | -0.9808472 | 0.9808472 | -0.1947787 | -0.0100000 |
| 21 | 63 | 33.8800 | 36.5530 | 0.9998210 | 0.0189193 | -0.0189193 | 0.9998210 | -1.0000000 |
| 22 | 63 | 34.1050 | 36.5530 | 0.9998210 | 0.0189193 | -0.0189193 | 0.9998210 | $-1.0000000$ |
| 23 | 63 | 34.3260 | 36.5530 | 0.9999964 | 0.0026806 | -0.0026806 | 0.9999964 | -1.0000000 |
| 24 | 63 | 34.4890 | 36.5470 | 0.9965425 | 0.0830845 | -0.0830845 | 0.9965425 | $-1.0000000$ |
| 25 | 63 | 34.7040 | 36.5580 | 0.9988645 | -0.0476422 | 0.0476422 | 0.9988645 | $-1.0000000$ |
| 26 | 63 | 35.0060 | 36.5930 | 0.9943762 | -0.1059052 | 0.1059052 | 0.9943762 | -1.0000000 |
| 27 | 63 | 35.2910 | 36.6250 | 0.9950503 | -0.0993730 | 0.0993730 | 0.9950503 | -1.0000000 |
| 28 | 63 | 35.5870 | 36.6470 | 0.9975940 | -0.0693269 | 0.0693269 | 0.9975940 | -1.0000000 |
| 43 | 63 | 39.6250 | 60.2000 | 0.2470855 | 0.9689937 | -0.9689937 | 0.2470855 | -1.0000000 |
| 44 | 63 | 39.6250 | 60.2000 | 0.2470855 | 0.9689937 | -0.9689937 | 0.2470855 | -0.0100000 |
| 4 | 64 | 11.9050 | 57.8050 | 0.6183338 | -0.7859156 | 0.7859156 | 0.6183338 | -1.0000000 |
| 5 | 64 | 11.9050 | 57.8050 | 0.6183338 | -0.7859156 | 0.7859156 | 0.6183338 | $-1.0000000$ |
| 7 | 64 | 19.3000 | 65.2750 | -0.2609794 | -0.9653444 | 0.9653444 | -0.2609794 | -1.0000000 |
| 8 | 64 | 19.3000 | 65.2750 | -0.2609794 | -0.9653444 | 0.9653444 | -0.2609794 | -0.0100000 |
| 22 | 64 | 34.1140 | 37.2260 | 0.9865040 | 0.1637369 | -0.1637369 | 0.9865040 | -1.0000000 |

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| 23 | 64 | 34.3940 | 37.2260 | 0.9865040 | 0.1637369 | -0.1637369 | 0.9865040 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 24 | 64 | 34.5700 | 37.1880 | 0.9655078 | 0.2603740 | -0.2603740 | 0.9655078 | -1.0000000 |
| 25 | 64 | 34.7620 | 37.1530 | 0.9909816 | 0.1339979 | -0.1339979 | 0.9909816 | -1.0000000 |
| 26 | 64 | 35.0140 | 37.1350 | 0.9971902 | 0.0749107 | -0.0749107 | 0.9971902 | -1.0000000 |
| 27 | 64 | 35.2980 | 37.1150 | 0.9952444 | 0.0974089 | -0.0974089 | 0.9952444 | -1.0000000 |
| 28 | 64 | 35.6000 | 37.0810 | 0.9932615 | 0.1158948 | -0.1158948 | 0.9932615 | -1.0000000 |
| 43 | 64 | 40.0130 | 60.0370 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -1.0000000 |
| 44 | 64 | 40.0130 | 60.0370 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -0.0100000 |
| 4 | 65 | 11.5550 | 58.2870 | 0.9234211 | -0.3837884 | 0.3837884 | 0.9234211 | -1.0000000 |
| 5 | 65 | 11.5550 | 58.2870 | 0.9234211 | -0.3837884 | 0.3837884 | 0.9234211 | -1.0000000 |
| 7 | 65 | 18.6880 | 65.2250 | 0.3076707 | -0.9514929 | 0.9514929 | 0.3076707 | -1.0000000 |
| 8 | 65 | 18.6880 | 65.2250 | 0.3076707 | -0.9514929 | 0.9514929 | 0.3076707 | -0.0100000 |
| 22 | 65 | 34.3080 | 37.8420 | 0.9888400 | 0.1489810 | -0.1489810 | 0.9888400 | -1.0000000 |
| 23 | 65 | 34.4880 | 37.8420 | 0.9888400 | 0.1489810 | -0.1489810 | 0.9888400 | -1.0000000 |
| 24 | 65 | 34.6640 | 37.8110 | 0.9922076 | 0.1245959 | -0.1245959 | 0.9922076 | -1.0000000 |
| 25 | 65 | 34.8450 | 37.7860 | 0.9948414 | 0.1014425 | -0.1014425 | 0.9948414 | -1.0000000 |
| 26 | 65 | 35.0840 | 37.7610 | 0.9941316 | 0.1081773 | -0.1081773 | 0.9941316 | -1.0000000 |
| 27 | 65 | 35.3040 | 37.6610 | 0.9941316 | 0.1081773 | -0.1081773 | 0.9941316 | -1.0000000 |
| 43 | 65 | 40.1120 | 59.3880 | 0.9684899 | -0.2490529 | 0.2490529 | 0.9684899 | -1.0000000 |
| 44 | 65 | 40.1120 | 59.3880 | 0.9684899 | -0.2490529 | 0.2490529 | 0.9684899 | -0.0100000 |
| 4 | 66 | 11.5120 | 58.9380 | 0.9999809 | 0.0061868 | -0.0061868 | 0.9999809 | $-1.0000000$ |
| 5 | 66 | 11.5120 | 58.9380 | 0.9999809 | 0.0061868 | -0.0061868 | 0.9999809 | -1.0000000 |
| 7 | 66 | 18.0620 | 65.6000 | 0.6124235 | -0.7905298 | 0.7905298 | 0.6124235 | -1.0000000 |
| 8 | 66 | 18.0620 | 65.6000 | 0.6124235 | -0.7905298 | 0.7905298 | 0.6124235 | -0.0100000 |
| 19 | 66 | 33.5460 | 38.5080 | 0.9931667 | 0.1167046 | -0.1167046 | 0.9931667 | -1.0000000 |
| 20 | 66 | 33.7960 | 38.5080 | 0.9931667 | 0.1167046 | -0.1167046 | 0.9931667 | -1.0000000 |
| 21 | 66 | 34.0540 | 38.4830 | 0.9937400 | 0.1117173 | -0.1117173 | 0.9937400 | $-1.0000000$ |
| 22 | 66 | 34.3310 | 38.4640 | 0.9956812 | 0.0928388 | -0.0928388 | 0.9956812 | -1.0000000 |
| 23 | 66 | 34.5510 | 38.4320 | 0.9826409 | 0.1855181 | -0.1855181 | 0.9826409 | -1.0000000 |
| 24 | 66 | 34.7210 | 38.4010 | 0.9961902 | 0.0872063 | -0.0872063 | 0.9961902 | -1.0000000 |
| 25 | 66 | 34.8910 | 38.3880 | 0.9976149 | 0.0690250 | -0.0690250 | 0.9976149 | -1.0000000 |
| 43 | 66 | 40.3750 | 58.9000 | -0.4631031 | 0.8863044 | -0.8863044 | -0.4631031 | -1.0000000 |
| 44 | 66 | 40.3750 | 58.9000 | -0.4631031 | 0.8863044 | -0.8863044 | -0.4631031 | -0.0100000 |
| 4 | 67 | 11.2550 | 59.3880 | 0.6386882 | -0.7694656 | 0.7694656 | 0.6386882 | -1.0000000 |
| 5 | 67 | 11.2550 | 59.3880 | 0.6386882 | -0.7694656 | 0.7694656 | 0.6386882 | -1.0000000 |
| 7 | 67 | 17.6750 | 66.1250 | 0.9341418 | -0.3569020 | 0.3569020 | 0.9341418 | -1.0000000 |
| 8 | 67 | 17.6750 | 66.1250 | 0.9341418 | -0.3569020 | 0.3569020 | 0.9341418 | -0.0100000 |
| 18 | 67 | 33.3940 | 39.1580 | 0.9992611 | -0.0384342 | 0.0384342 | 0.9992611 | -1.0000000 |
| 19 | 67 | 33.5960 | 39.1650 | 0.9992186 | -0.0395252 | 0.0395252 | 0.9992186 | -1.0000000 |
| 20 | 67 | 33.8200 | 39.1620 | 0.9998428 | 0.0177342 | -0.0177342 | 0.9998428 | -1.0000000 |
| 23 | 67 | 34.5950 | 38.9590 | 0.9913591 | 0.1311760 | -0.1311760 | 0.9913591 | $-1.0000000$ |
| 24 | 67 | 34.7650 | 38.9340 | 0.9972720 | 0.0738143 | -0.0738143 | 0.9972720 | -1.0000000 |
| 25 | 67 | 34.9290 | 38.9280 | 0.9993833 | 0.0351133 | -0.0351133 | 0.9993833 | -1.0000000 |
| 43 | 67 | 40.9250 | 58.7700 | 0.0362918 | 0.9993412 | -0.9993412 | 0.0362918 | -1.0000000 |
| 44 | 67 | 40.9250 | 58.7700 | 0.0362918 | 0.9993412 | -0.9993412 | 0.0362918 | -0.0100000 |
| 4 | 68 | 10.6550 | 59.6300 | 0.4893424 | -0.8720917 | 0.8720917 | 0.4893424 | -1.0000000 |
| 5 | 68 | 10.6550 | 59.6300 | 0.4893424 | -0.8720917 | 0.8720917 | 0.4893424 | -1.0000000 |
| 7 | 68 | 17.5250 | 66.7500 | 0.9944351 | -0.1053515 | 0.1053515 | 0.9944351 | -1.0000000 |
| 8 | 68 | 17.5250 | 66.7500 | 0.9944351 | -0.1053515 | 0.1053515 | 0.9944351 | -0.0100000 |
| 18 | 68 | 33.3260 | 39.9250 | 0.9940071 | -0.1093160 | 0.1093160 | 0.9940071 | -1.0000000 |
| 19 | 68 | 33.5270 | 39.9440 | 0.9935773 | -0.1131556 | 0.1131556 | 0.9935773 | -1.0000000 |
| 24 | 68 | 34.8080 | 39.5750 | 0.9994620 | 0.0327986 | -0.0327986 | 0.9994620 | -1.0000000 |
| 43 | 68 | 41.4870 | 58.8570 | 0.1893369 | 0.9819122 | -0.9819122 | 0.1893369 | -1.0000000 |
| 44 | 68 | 41.4870 | 58.8570 | 0.1893369 | 0.9819122 | -0.9819122 | 0.1893369 | -0.0100000 |
| 4 | 69 | 10.1580 | 59.9550 | 0.6740256 | -0.7387080 | 0.7387080 | 0.6740256 | -1.0000000 |
| 5 | 69 | 10.1580 | 59.9550 | 0.6740256 | -0.7387080 | 0.7387080 | 0.6740256 | -1.0000000 |
| 7 | 69 | 17.4620 | 67.5000 | 0.9847905 | -0.1737459 | 0.1737459 | 0.9847905 | -1.0000000 |
| 8 | 69 | 17.4620 | 67.5000 | 0.9847905 | -0.1737459 | 0.1737459 | 0.9847905 | -0.0100000 |
| 17 | 69 | 33.1970 | 40.6920 | 0.9840309 | -0.1779976 | 0.1779976 | 0.9840309 | -0.0100000 |
| 18 | 69 | 33.1970 | 40.6920 | 0.9840309 | -0.1779976 | 0.1779976 | 0.9840309 | -1.0000000 |
| 19 | 69 | 33.3990 | 40.7230 | 0.9837608 | -0.1794844 | 0.1794844 | 0.9837608 | -1.0000000 |
| 23 | 69 | 34.6400 | 40.2230 | 0.9997077 | 0.0241757 | -0.0241757 | 0.9997077 | -1.0000000 |
| 24 | 69 | 34.8290 | 40.2100 | 0.9999974 | -0.0022844 | 0.0022844 | 0.9999974 | -1.0000000 |
| 43 | 69 | 41.9870 | 59.1250 | 0.5723556 | 0.8200055 | -0.8200055 | 0.5723556 | -1.0000000 |
| 44 | 69 | 41.9870 | 59.1250 | 0.5723556 | 0.8200055 | -0.8200055 | 0.5723556 | -0.0100000 |
| 4 | 70 | 9.7900 | 60.1620 | 0.5019256 | -0.8649108 | 0.8649108 | 0.5019256 | -1.0000000 |
| 5 | 70 | 9.7900 | 60.1620 | 0.5019256 | -0.8649108 | 0.8649108 | 0.5019256 | -1.0000000 |
| 7 | 70 | 17.3620 | 68.2250 | 0.8756611 | -0.4829261 | 0.4829261 | 0.8756611 | -1.0000000 |
| 8 | 70 | 17.3620 | 68.2250 | 0.8756611 | -0.4829261 | 0.4829261 | 0.8756611 | -0.0100000 |
| 17 | 70 | 33.0230 | 41.4300 | 0.9785854 | -0.2058412 | 0.2058412 | 0.9785854 | -0.0100000 |
| 18 | 70 | 33.0230 | 41.4300 | 0.9785854 | -0.2058412 | 0.2058412 | 0.9785854 | -1.0000000 |

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| 19 | 70 | 33.2310 | 41.4620 | 0.9806373 | -0.1958329 | 0.1958329 | 0.9806373 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 23 | 70 | 34.6130 | 40.7430 | 0.9997286 | -0.0232990 | 0.0232990 | 0.9997286 | -1.0000000 |
| 24 | 70 | 34.8080 | 40.7430 | 0.9997869 | -0.0206469 | 0.0206469 | 0.9997869 | -1.0000000 |
| 43 | 70 | 42.4750 | 59.5850 | 0.6230873 | 0.7821523 | -0.7821523 | 0.6230873 | -1.0000000 |
| 44 | 70 | 42.4750 | 59.5850 | 0.6230873 | 0.7821523 | -0.7821523 | 0.6230873 | -0.0100000 |
| 4 | 71 | 9.4325 | 60.0620 | 0.1084285 | -0.9941043 | 0.9941043 | 0.1084285 | -1.0000000 |
| 5 | 71 | 9.4325 | 60.0620 | 0.1084285 | -0.9941043 | 0.9941043 | 0.1084285 | -1.0000000 |
| 7 | 71 | 17.0250 | 68.8500 | 0.6407458 | -0.7677531 | 0.7677531 | 0.6407458 | -1.0000000 |
| 8 | 71 | 17.0250 | 68.8500 | 0.6407458 | -0.7677531 | 0.7677531 | 0.6407458 | -0.0100000 |
| 17 | 71 | 32.8710 | 42.1890 | 0.9917078 | -0.1285131 | 0.1285131 | 0.9917078 | -0.0100000 |
| 18 | 71 | 32.8710 | 42.1890 | 0.9917078 | -0.1285131 | 0.1285131 | 0.9917078 | -1.0000000 |
| 19 | 71 | 33.0910 | 42.2150 | 0.9929819 | -0.1182664 | 0.1182664 | 0.9929819 | -1.0000000 |
| 23 | 71 | 34.6290 | 41.2900 | 0.9954019 | 0.0957862 | -0.0957862 | 0.9954019 | -1.0000000 |
| 24 | 71 | 34.8240 | 41.2710 | 0.9938225 | 0.1109816 | -0.1109816 | 0.9938225 | -1.0000000 |
| 43 | 71 | 43.0620 | 60.0100 | 0.4570889 | 0.8894210 | -0.8894210 | 0.4570889 | -1.0000000 |
| 44 | 71 | 43.0620 | 60.0100 | 0.4570889 | 0.8894210 | -0.8894210 | 0.4570889 | -0.0100000 |
| 4 | 72 | 9.1625 | 60.2630 | 0.8397512 | -0.5429713 | 0.5429713 | 0.8397512 | -1.0000000 |
| 5 | 72 | 9.1625 | 60.2630 | 0.8397512 | -0.5429713 | 0.5429713 | 0.8397512 | -1.0000000 |
| 7 | 72 | 16.5130 | 69.2500 | 0.2858175 | -0.9582841 | 0.9582841 | 0.2858175 | -1.0000000 |
| 18 | 72 | 32.6880 | 42.9430 | 0.9531692 | -0.3024376 | 0.3024376 | 0.9531692 | -1.0000000 |
| 19 | 72 | 32.8960 | 43.0000 | 0.9531636 | -0.3024552 | 0.3024552 | 0.9531636 | -1.0000000 |
| 23 | 72 | 34.7090 | 41.7850 | 0.9730717 | 0.2305028 | -0.2305028 | 0.9730717 | -1.0000000 |
| 24 | 72 | 34.9040 | 41.7400 | 0.9762105 | 0.2168251 | -0.2168251 | 0.9762105 | -1.0000000 |
| 43 | 72 | 43.6250 | 60.3250 | 0.4178606 | 0.9085112 | -0.9085112 | 0.4178606 | -1.0000000 |
| 44 | 72 | 43.6250 | 60.3250 | 0.4178606 | 0.9085112 | -0.9085112 | 0.4178606 | -0.0100000 |
| 4 | 73 | 8.8700 | 60.7630 | 0.6413739 | -0.7672285 | 0.7672285 | 0.6413739 | -1.0000000 |
| 5 | 73 | 8.8700 | 60.7630 | 0.6413739 | -0.7672285 | 0.7672285 | 0.6413739 | -1.0000000 |
| 18 | 73 | 32.3440 | 43.7110 | 0.8784728 | -0.4777924 | 0.4777924 | 0.8784728 | -1.0000000 |
| 19 | 73 | 32.5130 | 43.8000 | 0.8829616 | -0.4694452 | 0.4694452 | 0.8829616 | -1.0000000 |
| 23 | 73 | 34.8130 | 42.2290 | 0.9726977 | 0.2320757 | -0.2320757 | 0.9726977 | -1.0000000 |
| 24 | 73 | 34.9950 | 42.1840 | 0.9786102 | 0.2057235 | -0.2057235 | 0.9786102 | -1.0000000 |
| 43 | 73 | 44.2370 | 60.5200 | 0.1495430 | 0.9887552 | -0.9887552 | 0.1495430 | -1.0000000 |
| 44 | 73 | 44.2370 | 60.5200 | 0.1495430 | 0.9887552 | -0.9887552 | 0.1495430 | -0.0100000 |
| 4 | 74 | 8.4350 | 61.1250 | 0.7369719 | -0.6759234 | 0.6759234 | 0.7369719 | -1.0000000 |
| 5 | 74 | 8.4350 | 61.1250 | 0.7369719 | -0.6759234 | 0.6759234 | 0.7369719 | -1.0000000 |
| 19 | 74 | 32.0570 | 44.4290 | 0.8754952 | -0.4832268 | 0.4832268 | 0.8754952 | -1.0000000 |
| 23 | 74 | 34.9240 | 42.6660 | 0.9686953 | 0.2482525 | -0.2482525 | 0.9686953 | -1.0000000 |
| 24 | 74 | 35.0820 | 42.6340 | 0.9829894 | 0.1836624 | -0.1836624 | 0.9829894 | -1.0000000 |
| 43 | 74 | 44.9630 | 60.6200 | 0.1629133 | 0.9866404 | -0.9866404 | 0.1629133 | -1.0000000 |
| 44 | 74 | 44.9630 | 60.6200 | 0.1629133 | 0.9866404 | -0.9866404 | 0.1629133 | -0.0100000 |
| 4 | 75 | 8.2275 | 61.6380 | 0.9511582 | -0.3087040 | 0.3087040 | 0.9511582 | -1.0000000 |
| 5 | 75 | 8.2275 | 61.6380 | 0.9511582 | -0.3087040 | 0.3087040 | 0.9511582 | -1.0000000 |
| 18 | 75 | 31.8430 | 44.9810 | 0.9966757 | -0.0814712 | 0.0814712 | 0.9966757 | -0.0100000 |
| 19 | 75 | 31.8430 | 44.9810 | 0.9966757 | -0.0814712 | 0.0814712 | 0.9966757 | -1.0000000 |
| 23 | 75 | 35.0350 | 43.0900 | 0.9773861 | 0.2114628 | -0.2114628 | 0.9773861 | -1.0000000 |
| 24 | 75 | 35.1680 | 43.0650 | 0.9813164 | 0.1924011 | -0.1924011 | 0.9813164 | -1.0000000 |
| 43 | 75 | 45.6120 | 60.7870 | 0.3792095 | 0.9253109 | -0.9253109 | 0.3792095 | -1.0000000 |
| 44 | 75 | 45.6120 | 60.7870 | 0.3792095 | 0.9253109 | -0.9253109 | 0.3792095 | -0.0100000 |
| 4 | 76 | 7.9500 | 61.9000 | 0.1601805 | -0.9870877 | 0.9870877 | 0.1601805 | -1.0000000 |
| 5 | 76 | 7.9500 | 61.9000 | 0.1601805 | -0.9870877 | 0.9870877 | 0.1601805 | -1.0000000 |
| 18 | 76 | 31.7810 | 45.5770 | 0.9644126 | -0.2644018 | 0.2644018 | 0.9644126 | -0.0100000 |
| 19 | 76 | 31.7810 | 45.5770 | 0.9644126 | -0.2644018 | 0.2644018 | 0.9644126 | -1.0000000 |
| 23 | 76 | 35.1350 | 43.5460 | 0.9788923 | 0.2043765 | -0.2043765 | 0.9788923 | -1.0000000 |
| 24 | 76 | 35.2740 | 43.5150 | 0.9694539 | 0.2452736 | -0.2452736 | 0.9694539 | -1.0000000 |
| 43 | 76 | 46.1750 | 61.0880 | 0.5301558 | 0.8479002 | -0.8479002 | 0.5301558 | -1.0000000 |
| 44 | 76 | 46.1750 | 61.0880 | 0.5301558 | 0.8479002 | -0.8479002 | 0.5301558 | -0.0100000 |
| 4 | 77 | 7.5625 | 62.1380 | 0.7725407 | -0.6349652 | 0.6349652 | 0.7725407 | -1.0000000 |
| 5 | 77 | 7.5625 | 62.1380 | 0.7725407 | -0.6349652 | 0.6349652 | 0.7725407 | -1.0000000 |
| 18 | 77 | 31.8620 | 46.1610 | 0.9407222 | 0.3391779 | -0.3391779 | 0.9407222 | -0.0100000 |
| 19 | 77 | 31.8620 | 46.1610 | 0.9407222 | 0.3391779 | -0.3391779 | 0.9407222 | -1.0000000 |
| 23 | 77 | 35.2170 | 44.0410 | 0.9972686 | 0.0738612 | -0.0738612 | 0.9972686 | -1.0000000 |
| 24 | 77 | 35.3680 | 44.0220 | 0.9868888 | 0.1614020 | -0.1614020 | 0.9868888 | -1.0000000 |
| 43 | 77 | 46.7630 | 61.3620 | 0.5048463 | 0.8632092 | -0.8632092 | 0.5048463 | -1.0000000 |
| 44 | 77 | 46.7630 | 61.3620 | 0.5048463 | 0.8632092 | -0.8632092 | 0.5048463 | -0.0100000 |
| 4 | 78 | 7.2700 | 62.6450 | 0.5539256 | -0.8325662 | 0.8325662 | 0.5539256 | -1.0000000 |
| 5 | 78 | 7.2700 | 62.6450 | 0.5539256 | -0.8325662 | 0.8325662 | 0.5539256 | -1.0000000 |
| 19 | 78 | 32.0820 | 46.7960 | 0.9803649 | 0.1971922 | -0.1971922 | 0.9803649 | -1.0000000 |
| 24 | 78 | 35.3280 | 44.5190 | 0.9819896 | -0.1889350 | 0.1889350 | 0.9819896 | -1.0000000 |
| 43 | 78 | 47.2750 | 61.7500 | 0.8320503 | 0.5547002 | -0.5547002 | 0.8320503 | -1.0000000 |
| 44 | 78 | 47.2750 | 61.7500 | 0.8320503 | 0.5547002 | -0.5547002 | 0.8320503 | -0.0100000 |
| 4 | 79 | 6.7950 | 62.6500 | -0.3218375 | -0.9467949 | 0.9467949 | -0.3218375 | -1.0000000 |

Appendix B

| 5 | 79 | 6.7950 | 62.6500 | -0.3218375 | -0.9467949 | 0.9467949 | -0.3218375 | -1.0000000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 22 | 79 | 34.9190 | 44.8930 | 0.9080355 | -0.4188932 | 0.4188932 | 0.9080355 | -1.0000000 |
| 23 | 79 | 35.0390 | 44.9500 | 0.9217742 | -0.3877271 | 0.3877271 | 0.9217742 | -1.0000000 |
| 24 | 79 | 35.1770 | 44.9940 | 0.9677365 | -0.2519643 | 0.2519643 | 0.9677365 | -1.0000000 |
| 43 | 79 | 47.6500 | 62.3000 | 0.8591530 | 0.5117188 | -0.5117188 | 0.8591530 | -1.0000000 |
| 44 | 79 | 47.6500 | 62.3000 | 0.8591530 | 0.5117188 | -0.5117188 | 0.8591530 | -1.0000000 |
| 4 | 80 | 6.2000 | 62.5620 | 0.3957258 | -0.9183687 | 0.9183687 | 0.3957258 | -1.0000000 |
| 5 | 80 | 6.2000 | 62.5620 | 0.3957258 | -0.9183687 | 0.9183687 | 0.3957258 | -1.0000000 |
| 21 | 80 | 34.5790 | 45.2610 | 0.3922861 | -0.9198433 | 0.9198433 | 0.3922861 | -0.0100000 |
| 22 | 80 | 34.5790 | 45.2610 | 0.3922861 | -0.9198433 | 0.9198433 | 0.3922861 | -1.0000000 |
| 24 | 80 | 35.1270 | 45.5270 | 0.9896071 | 0.1437977 | -0.1437977 | 0.9896071 | -1.0000000 |
| 43 | 80 | 47.9870 | 62.8250 | 0.8646075 | 0.5024480 | -0.5024480 | 0.8646075 | -1.0000000 |
| 44 | 80 | 47.9870 | 62.8250 | 0.8646075 | 0.5024480 | -0.5024480 | 0.8646075 | -1.0000000 |
| 4 | 81 | 5.8175 | 62.9580 | 0.8627293 | -0.5056661 | 0.5056661 | 0.8627293 | -1.0000000 |
| 5 | 81 | 5.8175 | 62.9580 | 0.8627293 | -0.5056661 | 0.5056661 | 0.8627293 | -1.0000000 |
| 21 | 81 | 34.0510 | 45.3300 | 0.0077300 | -0.9999701 | 0.9999701 | 0.0077300 | -1.0000000 |
| 22 | 81 | 34.0510 | 45.3300 | 0.0077300 | -0.9999701 | 0.9999701 | 0.0077300 | -1.0000000 |
| 24 | 81 | 35.2770 | 46.0860 | 0.8954697 | 0.4451226 | -0.4451226 | 0.8954697 | -1.0000000 |
| 43 | 81 | 48.1570 | 63.6050 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 44 | 81 | 48.1570 | 63.6050 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 4 | 82 | 5.3925 | 63.2500 | 0.1675430 | -0.9858648 | 0.9858648 | 0.1675430 | -1.0000000 |
| 5 | 82 | 5.3925 | 63.2500 | 0.1675430 | -0.9858648 | 0.9858648 | 0.1675430 | -1.0000000 |
| 43 | 82 | 47.5570 | 65.0350 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -1.0000000 |
| 44 | 82 | 47.5570 | 65.0350 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -1.0000000 |
| 4 | 83 | 4.7575 | 63.1750 | -0.2896261 | -0.9571398 | 0.9571398 | -0.2896261 | -1.0000000 |
| 5 | 83 | 4.7575 | 63.1750 | -0.2896261 | -0.9571398 | 0.9571398 | -0.2896261 | -1.0000000 |
| 43 | 83 | 46.9900 | 66.4000 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 44 | 83 | 46.9900 | 66.4000 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 4 | 84 | 4.1875 | 63.2000 | 0.2911504 | -0.9566773 | 0.9566773 | 0.2911504 | -1.0000000 |
| 5 | 84 | 4.1875 | 63.2000 | 0.2911504 | -0.9566773 | 0.9566773 | 0.2911504 | -1.0000000 |
| 43 | 84 | 46.4300 | 67.2800 | -0.7189829 | 0.6950278 | -0.6950278 | -0.7189829 | -1.0000000 |
| 44 | 84 | 46.4300 | 67.2800 | -0.7189829 | 0.6950278 | -0.6950278 | -0.7189829 | -1.0000000 |
| 4 | 85 | 3.7200 | 63.5500 | 0.7358159 | -0.6771816 | 0.6771816 | 0.7358159 | -1.0000000 |
| 5 | 85 | 3.7200 | 63.5500 | 0.7358159 | -0.6771816 | 0.6771816 | 0.7358159 | -1.0000000 |
| 43 | 85 | 45.7200 | 68.9000 | 0.9884899 | -0.1512869 | 0.1512869 | 0.9884899 | -1.0000000 |
| 44 | 85 | 45.7200 | 68.9000 | 0.9884899 | -0.1512869 | 0.1512869 | 0.9884899 | -1.0000000 |
| 43 | 86 | 45.5500 | 70.8500 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 44 | 86 | 45.4500 | 70.8500 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 43 | 87 | 45.3000 | 72.9700 | 0.9684899 | -0.2490529 | 0.2490529 | 0.9684899 | -1.0000000 |
| 44 | 87 | 45.4500 | 72.9700 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 43 | 88 | 45.6700 | 75.0500 | 0.7587842 | 0.6513421 | -0.6513421 | 0.7587842 | -1.0000000 |
| 44 | 88 | 45.6700 | 75.0500 | 0.7587842 | 0.6513421 | -0.6513421 | 0.7587842 | -1.0000000 |
| 43 | 89 | 46.4100 | 77.1600 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 44 | 89 | 46.4100 | 77.1600 | 1.0000000 | 0.0000000 | -0.0000000 | 1.0000000 | -1.0000000 |
| 43 | 90 | 47.2800 | 79.0800 | 0.6587842 | 0.7523320 | -0.7523320 | 0.6587842 | -1.0000000 |
| 44 | 90 | 47.2800 | 79.0800 | 0.6587842 | 0.7523320 | -0.7523320 | 0.6587842 | -1.0000000 |
| 43 | 91 | 47.5800 | 80.5800 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -1.0000000 |
| 44 | 91 | 47.5800 | 80.5800 | -0.8443334 | 0.5358183 | -0.5358183 | -0.8443334 | -1.0000000 |
| 43 | 92 | 47.5700 | 82.4800 | 0.9219567 | 0.3872930 | -0.3872930 | 0.9219567 | -1.0000000 |
| 44 | 92 | 47.5500 | 82.4800 | 0.9219567 | 0.3872930 | -0.3872930 | 0.9219567 | -1.0000000 |
| 43 | 93 | 46.9700 | 83.5300 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -1.0000000 |
| 44 | 93 | 46.9700 | 83.6300 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -1.0000000 |
| 43 | 94 | 45.5100 | 83.9000 | -0.6443334 | 0.7647447 | -0.7647447 | -0.6443334 | -1.0000000 |
| 44 | 94 | 45.5100 | 83.9000 | -0.6443334 | 0.7647447 | -0.7647447 | -0.6443334 | -1.0000000 |
| 43 | 95 | 44.4500 | 84.6000 | -0.4443334 | 0.8958615 | -0.8958615 | -0.4443334 | -1.0000000 |
| 44 | 95 | 44.4500 | 84.6000 | -0.4443334 | 0.8958615 | -0.8958615 | -0.4443334 | -1.0000000 |
| 43 | 96 | 43.1300 | 84.8000 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -1.0000000 |
| 44 | 96 | 43.1300 | 84.8000 | 0.0669665 | 0.9977552 | -0.9977552 | 0.0669665 | -1.0000000 |
| 43 | 97 | 42.1100 | 84.9700 | -0.6443334 | 0.7647447 | -0.7647447 | -0.6443334 | -1.0000000 |
| 44 | 97 | 42.1100 | 84.9700 | -0.6443334 | 0.7647447 | -0.7647447 | -0.6443334 | -1.0000000 |

```
C71B FOOD CHAIN MODEL OUTPUT CONTROL
C
    NFDCHZ: NUMBER OF SPATIAL ZONES
    HBFDCH: AVERAGING DEPTH FOR TOP PORTION OF BED (METERS)
    TFCAVG: TIME AVERAGING INTERVAL FOR FOOD CHAIN OUTPUT (SECONDS)
C
C71B ISFDCH NFDCHZ HBFDCH TFCAVG
    0 0.1524 86400.
--------------------------------------------------------------------------------------772
CONTROLS FOR HORIZONTAL SURFACE ELEVATION OR PRESSURE CONTOURING
C72 CONTROLS FOR HORIZONTAL SURFACE ELEVATION OR PRESSURE CONTOURING |
    C
    ISPPH: 1 TO WRITE FILE FOR SURF ELEVATION CONTOURING
|
                    2 ~ W R I T E ~ O N L Y ~ D U R I N G ~ L A S T ~ R E F E R E N C E ~ T I M E ~ P E R I O D ~
            NPPPH: NUMBER OF WRITES PER REFERENCE TIME PERIOD
            ISRPPH: 1 TO WRITE FILE FOR RESIDUAL SURFACE ELEVATION CONTOURNG IN
                HORIZONTAL PLANE
            IPPHXY: O DOES NOT WRITE I,J,X,Y IN surfplt.out and rsurfplt.out FILES
                    1 WRITES I,J ONLY IN surfplt.out and rsurfplt.out FILES
                    2 WRITES I,J,X,Y IN surfplt.out and rsurfplt.out FILES
                    3 WRITES EFDC EXPLORER BINARY FORMAT FILES
C
C72 ISPPH NPPPH ISRPPH IPPHXY ।
-------------------------------------------------------------------------------------------
C73 CONTROLS FOR HORIZONTAL PLANE VELOCITY VECTOR PLOTTING
C
            ISVPH: 1 TO WRITE FILE FOR VELOCITY PLOTTING IN HORIZONTAL PLANE
                    2 WRITE ONLY DURING LAST REFERENCE TIME PERIOD
            NPVPH: NUMBER OF WRITES PER REFERENCE TIME PERIOD
            ISRVPH: 1 TO WRITE FILE FOR RESIDUAL VELOCITY PLOTTIN IN
                HORIZONTAL PLANE
            IVPHXY: O DOES NOT WRITE I,J,X,Y IN velplth.out and rvelplth.out FILES
            1 WRITES I,J ONLY IN velplth.out and rvelplth.out FILES
                    2 \text { WRITES I,J,X,Y IN velplth.out and rvelplth.out FILES}
                    3 WRITES EFDC EXPLORER BINARY FORMAT FILES
C
C73 ISVPH NPVPH ISRVPH IVPHXY
```



```
C74 CONTROLS FOR VERTICAL PLANE SCALAR FIELD CONTOURING
C
            ISECSPV: N AN INTEGER NUMBER OF VERTICAL SECTIONS (N.LE.9) TO WRITE 
                N FILES FOR SCALAR FIELD CONTOURING
            NPSPV: NUMBER OF WRITES PER REFERENCE TIME PERIOD
            ISSPV: 1 TO ACTIVATE INSTANTANEOUS SCALAR FIELDS
            2 ~ T O ~ W R I T E ~ O N L Y ~ D U R I N G ~ L A S T ~ R E F E R E N C E ~ T I M E ~ P E R I O D ~
            ISRSPV: 1 TO ACTIVATE FOR RESIDUAL SCALAR FIELDS
            ISHPLTV: 1 FOR VERTICAL PLANE PLOTTING FOR MSL DATUMS, ZERO OTHERWISE
            DATA LINE REPEATS 7 TIMES FOR SAL,TEM,DYE,SFL,TOX,SED,SND
            ISECSPV IS DETERMINED FOR ALL 7 VARIABLES BY VALUE ON FIRST DATA LINE
C
\begin{tabular}{cccccc} 
C74 4 ISECSPV & NPSPV & ISSPV & ISRSPV & ISHPLTV & \\
1 & 6 & 0 & 0 & 1 & !SAL \\
0 & 6 & 0 & 0 & 1 & !TEM \\
0 & 6 & 0 & 0 & 1 & !DYE \\
0 & 6 & 0 & 0 & 1 & !SFL \\
0 & 6 & 0 & 0 & 1 & !TOX \\
0 & 6 & 0 & 0 & 1 & !SED
\end{tabular}
```



```
            2 TO WRITE TO 3D CHARACTER ARRAY FORMAT FILES (NOT ACTIVE) |
            3 TO WRITE TO 3D HDF IMAGE FORMAT FILES (NOT ACTIVE)
            4 ~ T O ~ W R I T E ~ T O ~ 3 D ~ H D F ~ F L O A T I N G ~ P O I N T ~ F O R M A T ~ F I L E S ~ ( N O T ~ A C T I V E ) ~
                ISR3DO: SAME AS IS3DO EXCEPT FOR RESIDUAL VARIABLES
                NP3DO: NUMBER OF WRITES PER LAST REF TIME PERIOD FOR INST VARIABLES
                KPC: NUMBER OF UNSTRETCHED PHYSICAL VERTICAL LAYERS
                NWGG: IF NWGG IS GREATER THAN ZERO, NWGG DEFINES THE NUMBER OF !2877।
                    WATER CELLS IN CARTESIAN 3D GRAPHICS GRID OVERLAY OF THE
                        CURVILINEAR GRID. FOR NWGG>0 AND EFDC RUNS ON A CURVILINEAR
                        GRID, I3DMI,I3DMA,J3DMI,J3DMA REFER TO CELL INDICES ON THE
                                    ON THE CARTESIAN GRAPHICS GRID OVERLAY DEFINED BY FILE
                                    gcell.inp. THE FILE gcell.inp IS NOT USED BY EFDC, BUT BY
                                    THE COMPANION GRID GENERATION CODE GEFDC.F. INFORMATION
                                    DEFINING THE OVERLAY IS READ BY EFDC.F FROM THE FILE
                                    gcellmp.inp. IF NWGG EQUALS 0, I3DMI,I3DMA,J3DMI,J3DMA REFER
                                    TO INDICES ON THE EFDC GRID DEFINED BY cell.inp.
                                    ACTIVATION OF THE REWRITE OPTION I3DRW=1 WRITES TO THE FULL
                                    GRID DEFINED BY cell.inp AS IF cell.inp DEFINES A CARTESIAN
                                    GRID. IF NWGG EQ O AND THE EFDC COMP GRID IS CO, THE REWRITE |
                    OPTION IS NOT RECOMMENDED AND A POST PROCESSOR SHOULD BE USED |
                        TO TRANSFER THE SHORT FORM, I3DRW=0, OUTPUT TO AN APPROPRIATE |
                    FORMAT FOR VISUALIZATION. CONTACT DEVELOPER FOR MORE DETAILS |
I3DMI: MINIMUM OR BEGINNING I INDEX FOR 3D ARRAY OUTPUT |
I3DMA: MAXIMUM OR ENDING I INDEX FOR 3D ARRAY OUTPUT
J3DMI: MINIMUM OR BEGINNING J INDEX FOR 3D ARRAY OUTPUT
J3DMA: MAXIMUM OR ENDING J INDEX FOR 3D ARRAY OUTPUT
I3DRW: O FILES WRITTEN FOR ACTIVE CO WATER CELLS ONLY
    1 REWRITE FILES TO CORRECT ORIENTATION DEFINED BY gcell.inp
        AND gcellmp.inp FOR CO WITH NWGG.GT.O OR BY cell.inp IF THE
        COMPUTATIONAL GRID IS CARTESIAN AND NWGG.EQ.O
        SELVMAX: MAXIMUM SURFACE ELEVATION FOR UNSTRETCHING (ABOVE MAX SELV )
        BELVMIN: MINIMUM BOTTOM ELEVATION FOR UNSTRETCHING (BELOW MIN BELV)
C
C80 IS3DO ISR3DO NP3DO KPC NWGG I3DMI I3DMA J3DMI J3DMA I3DRW SELVMAX BELVMIN |
```



```
-------------------------------------------------------------------------------------------------------
C81 OUTPUT ACTIVATION AND SCALES FOR 3D FIELD OUTPUT
C
    VARIABLE: DUMMY VARIBLE ID (DO NOT CHANGE ORDER) |
    IS3(VARID): 1 TO ACTIVATE THIS VARIBLES |
    JS3(VARID): 0 FOR NO SCALING OF THIS VARIABLE
    1 ~ F O R ~ A U T O ~ S C A L I N G ~ O F ~ T H I S ~ V A R I A B L E ~ O V E R ~ R A N G E ~ 0 < V A L < 2 5 5 ~
                        AUTO SCALES FOR EACH FRAME OUTPUT IN FILES out3d.dia AND |
                        rout3d.dia OUTPUT IN I4 FORMAT
    2 ~ F O R ~ S C A L I N G ~ S P E C I F I E D ~ I N ~ N E X T ~ T W O ~ C O L U M N S ~ W I T H ~ O U T P U T ~
    DEFINED OVER RANGE 0<VAL<255 AND WRITTEN IN I4 FORMAT
    3 ~ F O R ~ M U L T I P L I E R ~ S C A L I N G ~ B Y ~ M A X ~ S C A L E ~ V A L U E ~ W I T H ~ O U T P U T ~
        WRITTEN IN F7.1 FORMAT (IS3DO AND ISR3DO MUST BE 1)
C
\begin{tabular}{lllrr} 
'U VEL' & 1 & 3 & 100.0 & -1.0 \\
'V VEL' & 1 & 3 & 100.0 & -1.0 \\
'W VEL' & 0 & 0 & 1000.0 & \(-1.0 \mathrm{E}-3\) \\
'SALINITY' & 1 & 3 & 1.0 & 0.0 \\
'TEMP' & 1 & 3 & 1.0 & 10.0 \\
'DYE' & 0 & 0 & 1000.0 & 0.0 \\
'COH SED' & 1 & 3 & 1000.0 & 0.0 \\
'NCH SED' & 1 & 3 & 1000.0 & 0.0 \\
'TOX CON' & 1 & 3 & 1000.0 & 0.0
\end{tabular}
----------------------------------------------------------------------------------------
C82 INPLACE HARMONIC ANALYSIS PARAMETERS
C
ISLSHA: 1 FOR IN PLACE LEAST SQUARES HARMONIC ANALYSIS


```

C90 MMLVSFP TIMVSFP IVSFP JVSFP |
*******************************************************************************
*******************************************************************************
*******************************************************************************

```
```


# <-- set first character to "\#" to use extended annotation in this file

C------------------------------------------------------------------------------------
CO1 MAIN TITLE CARDS
C
C TITLE(M), M=1,3
C
C01 THREE TITLE CARDS FOLLOW:
': Cape Fear River Water Quality Model'
': Variable stream flow and variable NPDES discharges'
': using tetratech input files'
C------------------------------------------------------------------------------------
C02 I/O CONTROL VARIABLE CARD
C
CO2 ONE TITLE CARD FOLLOWS:

$$
C02 I/O control variables
$$
C02
C IWQLVL = kinetic complexity level
1 WASP5 LEVEL KINETICS (SINGLE ORGANIC CARBON, PHOSPHOROUS, AND
NITROGEN CLASSES + REACTIVE DOC/CBOD VARIABLE)
2 INTERMEDIATE LEVEL KINETICS (TOTAL REFRACTORY AND TOTAL LABILE
ORGANIC CARBON, PHOSPHOROUS, AND NITROGEN CLASSES
+ REACTIVE DOC/CBOD)
3 CE-QUAL-ICM (ORIGINAL CHES BAY VARIABLES)
4 EXTENDED CE-QUAL-ICM (4 ORGANIC CARBON, PHOSPHOROUS, AND
NITROGEN CLASSES + REACTIVE DOC/CBOD VARIABLE)
NWQV = number of water quality water column variables
NWQZ = max. number of spatial zones having varying water quality parameters
NWQPS = max. number of water quality point source locations
NWQTD = number of data points in the temperature lookup table
NWQTS = max. number of water quality time-series output locations
NTSWQV = max. number of water quality time-series output variables
NSMG = number of sediment model groups (= 3)
NSMZ = max. number of sediment model spatial variation zones
NTSSMV = max. number of sediment model time-series output variables
NSMTS = not used
NWQKDPT = number of kinetic updates per transport update
C
C02 IWQLVL NWQV NWQZ NWQPS NWQTD NWQTS NTSWQV NSMG NSMZ NTSSMV NSMTS NWQKDPT
3 21 1 1 3 3 % 551 22 m
CO2A TRANSPORT BYPASS FLAGS

| C | B | B | B | R | L | D | R | L | D | P | R | L | D | N | N | S | S | C | D | T | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| C | c | d | g | P | P | 0 | P | P | 0 | $\bigcirc$ | P | P | 0 | H | 0 | U | A | $\bigcirc$ | $\bigcirc$ | A | C |
| C |  |  |  | 0 | 0 | C | $\bigcirc$ | 0 | P | 4 | 0 | 0 | N | 4 | 3 |  |  | D |  | M |  |
| C |  |  |  | C | C |  | P | P |  | t | N | N |  |  |  |  |  |  |  |  |  |
|  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 |

C03
IWQDT = number of water quality time steps per hydrodynamic time step
=1 for 2 time level hydrodynamic and =2 for three time level
IWQM = full or reduced model switch (1=full model; 2=reduced model)
C IWQBEN = benthic flux model switch (0=specified flux; 1=predictive flux, 2
predictive flux w/ time and space variation)
C IWQSI = switch to activate silica state variables (0=off; 1=activated)
C IWQFCB = switch to activate fecal coliform bacteria (0=off; 1=activated)
C IWQSRP = switch for sediment sorption (1=TAM sorption; 2=sediment sorption)
C IWQSTOX = cyanobacteria salinity toxicity switch (0=no toxicity; 1=toxicity)
C IWQKA = reaeration option
C = 0, constant reaeration (WQKRO), no wind reaeration
C = 1, constant reaeration (WQKRO) plus wind reaeration
C = 2, use O'Connor-Dobbins (1958) formula
C = 3, use Owens \& Gibbs (1964) formula
C = 4, modified Owens \& Gibbs (1964) formula (for Christina River)

```









```

C38 SIX TITLE CARDS FOLLOW:

$$
C38 EAST OPEN BOUNDARY
$$

\$\$ TIME SERIES ID'S FOR EACH STATE VARIABLE

| \$ I | J | B | B | B | R | L | D | R | L | D | P | R | L | D | N | N | S | S | c | D | T |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \$ |  | c | d | 9 | P | P | 0 | P | P | 0 | 0 | P | P | 0 | H | 0 | U | A | 0 | 0 | A |
| \$ |  |  |  |  | $\bigcirc$ | 0 | C | $\bigcirc$ | 0 | P | 4 | 0 | $\bigcirc$ | N | 4 | 3 |  |  | D |  | M |
| \$ |  |  |  |  | c | c |  | P | P |  | t | N | N |  |  |  |  |  |  |  |  |
| 35 | 44 | 0 | O | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |

C-----------------------------------------------------------------------------------
C39 FIVE TITLE CARDS FOLLOW:

$$
C39 EAST OPEN BOUNDARY
$$

\$\$ CONSTANT BOTTOM CONCENTRATION BC'S

| $\$$ | I | J | BC | Bd | Bg | RPOC | LPOC | DOC |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\$$ |  |  | RPOP | LPOP | DOP | PO4t | RPON | LPON | DON |  |
| $\$$ |  | NH4 | NO3 | SU | SA | COD | DO | TAM | FCB |  |
| 35 | 44 | 0.0 | 0.0 | 0.3 | 0.1 | 0.0 | 0.1 |  |  |  |
|  |  | 0.05 | 0.0 | 0.05 | 0.1 | 0.2 | 0.0 | 0.2 |  |  |
|  |  |  | 0.04 | 0.08 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |

```

```

C41 SIX TITLE CARDS FOLLOW:

$$
C41 NORTH OPEN BOUNDARY
$$

\$\$ TIME SERIES ID'S FOR EACH STATE VARIABLE

| \$ I | J | B | B | B | R | L | D | R | L | D | P | R | L | D | N | N | S | S | C | D |  | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \$ |  | c | d | g | P | P | 0 | P | P | 0 | $\bigcirc$ | P | P | 0 | H | 0 | U | A | $\bigcirc$ | 0 |  | C |
| \$ |  |  |  |  | $\bigcirc$ | 0 | C | $\bigcirc$ | 0 | P | 4 | 0 | 0 | N | 4 | 3 |  |  | D |  |  |  |
| \$ |  |  |  |  | C | C |  | P | P |  | t | N | N |  |  |  |  |  |  |  |  |  |

C42 FIVE TITLE CARDS FOLLOW:

$$
C42 NORTH OPEN BOUNDARY
$$

$$
CONSTANT BOTTOM CONCENTRATION BC'S
```

```
C44 ONE TITLE CARD FOLLOWS:
$$ C44 constant ICs (g/m^3): TAM(mol/m^3), FCB(MPN/100mL) \$\$

C44
C Definitions:
C BC = cyanobacteria
C Bd = algae diatoms
C Bg = algae greens
C RPOC = refractory particulate carbon
C LPOC = labile particulate carbon
C DOC = dissolved organic carbon
C RPOP = refractory particulate organic phosphorus
C LPOP = labile particulate organic phosphorus

```

```

C FNO3 = benthic flux rate of nitrite+nitrite nitrogen
C FSAD = benthic flux rate of silica
C FCOD = benthic flux rate of chemical oxygen demand
C SOD = sediment oxygen demand rate
C
C47 FPO4 FNH4 FNO3 FSAD FCOD SOD
0.000 0.000 -0.05 0.000 0.000 -0.50

```
C48 ONE TITLE CARD FOLLOWS:
\$\$ C48 const PS (kg/d): PSQ(m^3/s),DO(g/m^3),TAM(kmol/d),FCB(MPN/100mL)
C48
C IWQPS = number of point sources
C NPSTMSR = number of point source time series
C
C48 const PS (kg/d): PSQ(m^3/s),DO(g/m^3),TAM(kmol/d),FCB(MPN/100mL)
C IWQPS NPSTMSR
C IWQPS NPSTMSR
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline & 8 & & & & & & & & & & & \\
\hline C & I & J & K & N & PSQ & BC & Bd & Bg & RPOC & LPOC & DOC & \\
\hline C & & & & S & RPOP & LPOP & DOP & PO4t & RPON & LPON & DON & \\
\hline C & & & & R & NH4 & NO3 & SU & SA & COD & DO & TAM & FCB \\
\hline & 4 & 32 & 0 & 1 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & DUPONT- \\
\hline \multicolumn{13}{|l|}{WILMINGTON/BRUNSWICK} \\
\hline & & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\
\hline & & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 \\
\hline & 4 & 32 & 0 & 2 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & DUPONT- \\
\hline
\end{tabular}
WILMINGTON/BRUNSWICK
\(0.0 \quad 0.0\)
    \(\begin{array}{lllll}4 & 32 & 0 & 3 & 0 .\end{array}\)
        0.0
    \(0.0 \quad 0.0\)
    \(\begin{array}{lllll}4 & 27 & 0 & 4 & 0.0\end{array}\)
SPECIALTIES-WILMINGTON
    \(0.0 \quad 0.0\)
    0.0
SPECIALTIES-WILMINGTON
    \(\begin{array}{llllllll}0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\ 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0\end{array}\)
\(\begin{array}{lllllllllll}4 & 63 & 0 & 6 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0\end{array}\)
PAPER BD CO-RIEGELWD
\begin{tabular}{llllllll}
0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\
0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0
\end{tabular}
\begin{tabular}{cccccccccc}
23 & 10 & 0 & 8 \\
TOWN-WWTP/SOUTHPORT
\end{tabular}

WWTP
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & . 0 \\
\hline 23 & 18 & 0 & 13 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & ARCHER \\
\hline \multicolumn{13}{|l|}{DANIELS MIDLAND CO.} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|r|}{\multirow[b]{2}{*}{0}} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline 4 & 31 & 0 & 14 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & TAKEDA \\
\hline \multicolumn{13}{|l|}{CHEMICAL PRODUCTS} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|r|}{\multirow[b]{2}{*}{0.0}} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline 4 & 31 & 0 & 14 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & TAKEDA \\
\hline \multicolumn{13}{|l|}{CHEMICAL PRODUCTS, number 2} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 4 & 15 & 0 & 15 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & Leland \\
\hline \multicolumn{13}{|l|}{INDUSTRIAL PARK WWTP} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & . 0 \\
\hline 18 & 72 & 0 & 16 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & BELVILLE, \\
\hline \multicolumn{13}{|l|}{TOWN - WWTP} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & . 0 \\
\hline 4 & 27 & 0 & 17 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & FORTRON \\
\hline \multicolumn{13}{|l|}{INDUSTRIES} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|r|}{\multirow[b]{2}{*}{0.0}} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline 4 & 85 & 0 & 18 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & cape fear \\
\hline \multicolumn{13}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & . 0 \\
\hline 7 & 72 & 0 & 19 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & black \\
\hline \multicolumn{13}{|l|}{river} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|r|}{\multirow[b]{2}{*}{0.0}} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline 43 & 97 & 0 & 20 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & ne cape \\
\hline \multicolumn{13}{|l|}{fear river} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 43 & 21 & 0 & 21 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & NE Cape \\
\hline \multicolumn{13}{|l|}{Fear, estuary 1} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|r|}{\multirow[b]{2}{*}{0}} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline 43 & 35 & 0 & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & NE Cape \\
\hline \multicolumn{13}{|l|}{Fear, estuary 2} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 43 & 93 & 0 & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & ! & NE Cape \\
\hline \multicolumn{13}{|l|}{Fear, estuary 3} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \multicolumn{2}{|l|}{} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 43 & 87 & 0 & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & NE Cape \\
\hline \multicolumn{13}{|l|}{Fear, estuary 4} \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 43 & 47 & 0 & 25 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & NE Cape \\
\hline \multicolumn{13}{|l|}{\multirow[t]{2}{*}{Fear, estuary 5 0 0}} \\
\hline & & & & & & & 0.0 & 0.0 & 0.0 & 0.0 & & \\
\hline & & & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & 0.0 \\
\hline 43 & 79 & 0 & & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & & NE Cape \\
\hline \multicolumn{13}{|l|}{\multirow[t]{3}{*}{\(\begin{array}{llllllll}\text { Fear, estuary 6 } & & & \\ & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 \\ & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0\end{array}\)}} \\
\hline & & & & & & & & & & & & \\
\hline & & & & & & & & & & & & \\
\hline
\end{tabular}

\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\
\hline 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\
\hline 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & 0.0 & \\
\hline \multicolumn{8}{|l|}{C51 ONE TITLE CARD FOLLOWS:} \\
\hline \multicolumn{4}{|l|}{\$\$ C51 File names for spatially/temp} & va & param & 10 & \\
\hline \multicolumn{6}{|l|}{Restart file for end spatial distribution \(=\) wq3drst.out} & & \\
\hline \multicolumn{4}{|l|}{File for initial conditions (ICIFN)} & \(=\mathrm{N}\) & IFN. & & \\
\hline File & gal g & resp & d ( & \(=\mathrm{N}\) & RFN. & & \\
\hline File f & trlin & lgae & (STI & \(=\mathrm{N}\) & FN. & & \\
\hline Input & for I & KT, & SUNFN) & \(=\mathrm{N}\) & NFN. & & \\
\hline Input & for b & flu & BENFN) & \(=\mathrm{b}\) & tse & & \\
\hline \multicolumn{2}{|l|}{Input file for p} & sourc & ut (P & \(=\mathrm{N}\) & FN. & & \\
\hline \multicolumn{5}{|l|}{File for NPS input inc/atm input (NPLFN) \(=\mathrm{N}\)} & FN. & & \\
\hline \multicolumn{5}{|l|}{Diagnostic file-negative conc. (NCOFN) = N} & 3 dnc & & \\
\hline
\end{tabular}

\begin{tabular}{|c|c|c|c|c|c|c|}
\hline \multicolumn{7}{|l|}{912.50} \\
\hline 1 & 0.0 & 0.0 & -0.0278 & 0.0 & 0.0 & -0.696 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -1.044 \\
\hline \multicolumn{7}{|l|}{942.92} \\
\hline 1 & 0.0 & 0.0 & -0.0255 & 0.0 & 0.0 & -0.638 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -0.956 \\
\hline \multicolumn{7}{|l|}{973.33} \\
\hline 1 & 0.0 & 0.0 & -0.0214 & 0.0 & 0.0 & -0.535 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -0.803 \\
\hline \multicolumn{7}{|l|}{1003.75} \\
\hline 1 & 0.0 & 0.0 & -0.0180 & 0.0 & 0.0 & -0.449 \\
\hline 2 & 0.0 & 0.0 & -0.0270 & 0.0 & 0.0 & -0.674 \\
\hline \multicolumn{7}{|l|}{1034.17} \\
\hline 1 & 0.0 & 0.0 & -0.0134 & 0.0 & 0.0 & -0.336 \\
\hline 2 & 0.0 & 0.0 & -0.0202 & 0.0 & 0.0 & -0.504 \\
\hline \multicolumn{7}{|l|}{1064.58} \\
\hline 1 & 0.0 & 0.0 & -0.0089 & 0.0 & 0.0 & -0.223 \\
\hline 2 & 0.0 & 0.0 & -0.0134 & 0.0 & 0.0 & -0.335 \\
\hline \multicolumn{7}{|l|}{1095.00} \\
\hline 1 & 0.0 & 0.0 & -0.0075 & 0.0 & 0.0 & -0.188 \\
\hline 2 & 0.0 & 0.0 & -0.0113 & 0.0 & 0.0 & -0.281 \\
\hline \multicolumn{7}{|l|}{1125.42} \\
\hline 1 & 0.0 & 0.0 & -0.0080 & 0.0 & 0.0 & -0.199 \\
\hline 2 & 0.0 & 0.0 & -0.0119 & 0.0 & 0.0 & -0.298 \\
\hline \multicolumn{7}{|l|}{1155.83} \\
\hline 1 & 0.0 & 0.0 & -0.0095 & 0.0 & 0.0 & -0.237 \\
\hline 2 & 0.0 & 0.0 & -0.0142 & 0.0 & 0.0 & -0.355 \\
\hline \multicolumn{7}{|l|}{1186.25} \\
\hline 1 & 0.0 & 0.0 & -0.0134 & 0.0 & 0.0 & -0.336 \\
\hline 2 & 0.0 & 0.0 & -0.0202 & 0.0 & 0.0 & -0.504 \\
\hline \multicolumn{7}{|l|}{1216.67} \\
\hline 1 & 0.0 & 0.0 & -0.0180 & 0.0 & 0.0 & -0.449 \\
\hline 2 & 0.0 & 0.0 & -0.0270 & 0.0 & 0.0 & -0.674 \\
\hline \multicolumn{7}{|l|}{1247.08} \\
\hline 1 & 0.0 & 0.0 & -0.0241 & 0.0 & 0.0 & -0.601 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -0.902 \\
\hline \multicolumn{7}{|l|}{1277.50} \\
\hline 1 & 0.0 & 0.0 & -0.0278 & 0.0 & 0.0 & -0.696 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -1.044 \\
\hline \multicolumn{7}{|l|}{1307.92} \\
\hline 1 & 0.0 & 0.0 & -0.0255 & 0.0 & 0.0 & -0.638 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -0.956 \\
\hline \multicolumn{7}{|l|}{1338.33} \\
\hline 1 & 0.0 & 0.0 & -0.0214 & 0.0 & 0.0 & -0.535 \\
\hline 2 & 0.0 & 0.0 & -0.0300 & 0.0 & 0.0 & -0.803 \\
\hline \multicolumn{7}{|l|}{1368.75} \\
\hline 1 & 0.0 & 0.0 & -0.0180 & 0.0 & 0.0 & -0.449 \\
\hline 2 & 0.0 & 0.0 & -0.0270 & 0.0 & 0.0 & -0.674 \\
\hline \multicolumn{7}{|l|}{1399.17} \\
\hline 1 & 0.0 & 0.0 & -0.0134 & 0.0 & 0.0 & -0.336 \\
\hline 2 & 0.0 & 0.0 & -0.0202 & 0.0 & 0.0 & -0.504 \\
\hline \multicolumn{7}{|l|}{1429.58} \\
\hline 1 & 0.0 & 0.0 & -0.0089 & 0.0 & 0.0 & -0.223 \\
\hline 2 & 0.0 & 0.0 & -0.0134 & 0.0 & 0.0 & -0.335 \\
\hline \multicolumn{7}{|l|}{1465.00} \\
\hline 1 & 0.0 & 0.0 & -0.0075 & 0.0 & 0.0 & -0.188 \\
\hline 2 & 0.0 & 0.0 & -0.0113 & 0.0 & 0.0 & -0.281 \\
\hline
\end{tabular}```


[^0]:    Assumptions
    $\mathrm{DOC}=\mathrm{BODu} * \mathrm{C}$ to O 2 molecular weight ratio
    $\mathrm{ON}=(\mathrm{TN}-\mathrm{NH} 4) * .25$
    OP = DON

